

**PHASE 2 REPORT- REVIEW COPY
FURTHER SITE CHARACTERIZATION AND ANALYSIS
VOLUME 2E - BASELINE ECOLOGICAL RISK ASSESSMENT
HUDSON RIVER PCBs REASSESSMENT RI/FS**

AUGUST 1999



For

**U.S. Environmental Protection Agency
Region II
and
U.S. Army Corps of Engineers
Kansas City District**

**Book 2 of 3
Tables and Figures**

TAMS Consultants, Inc.

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- 4-9 Selected Mammal Aroclor and Total PCB Toxicity Endpoints
- 4-10 Selected Mammal Dioxin Equivalent Toxicity Endpoints

- 5-1 Complete Linkage Clustering - TI Pool
- 5-2 Relative Percent Grain Size Classes - TI Pool
- 5-3 Mean Sediment TOC - TI Pool
- 5-4 Mean Total PCB Concentrations in Sediment - TI Pool
- 5-5 Biomass of Benthic Invertebrates - TI Pool
- 5-6 Relative Percent Grain Size Classes - Lower Hudson River
- 5-7 Mean Sediment TOC - Lower Hudson River
- 5-8 Mean Total PCB Concentration in Sediment - Lower Hudson River

**TABLE 2-1
HUDSON RIVER FISHES**

| Common Name | Scientific Name | Predominant Habitat |
|--|--------------------------------|----------------------------|
| Anchovies - Family Engraulidae | | |
| Bay anchovy | <i>Anchoa mitchilli</i> | Saltwater |
| Striped anchovy | <i>Anchoa hepsetus</i> | Saltwater |
| Basses (Sea) - Family Serranidae | | |
| Black sea bass | <i>Centropristis striata</i> | Saltwater |
| Basses (Temperate) - Family Percichthyidae or Moronidae | | |
| Striped bass | <i>Morone saxatilis</i> | Anadromous |
| White bass | <i>Morone chrysops</i> | Freshwater |
| White perch | <i>Morone americana</i> | Freshwater/brackish |
| Bluefishes - Family Pomatomidae | | |
| Bluefish | <i>Pomatomus saltatrix</i> | Saltwater |
| Bowfins - Family Amiidae | | |
| Bowfin | <i>Amia calva</i> | Freshwater |
| Butterfishes - Family Stromateidae | | |
| Butterfish | <i>Peprilus triacanthus</i> | Saltwater |
| Catfishes - Family Ictaluridae | | |
| Brown bullhead | <i>Ictalurus nebulosus</i> | Freshwater |
| Channel catfish | <i>Ictalurus punctatus</i> | Freshwater |
| Margined madtom | <i>Noturus insignis</i> | Freshwater |
| Stonecat | <i>Noturus flavus</i> | Freshwater |
| Tadpole madtom | <i>Noturus gyrinus</i> | Freshwater |
| White catfish | <i>Ictalurus catus</i> | Freshwater |
| Yellow bullhead | <i>Ictalurus natalis</i> | Freshwater |
| Codfishes - Family Gadidae | | |
| Atlantic tomcod | <i>Microgadus tomcod</i> | Anadromous |
| Fourbeard rockling | <i>Enchelyopus cimbrius</i> | Saltwater |
| Red hake | <i>Urophycis chuss</i> | Saltwater |
| Silver hake | <i>Merluccius bilinearis</i> | Saltwater |
| Drums - Family Sciaenidae | | |
| Atlantic croaker | <i>Micropogonias undulatus</i> | Saltwater |
| Spot | <i>Leiostomus xanthurus</i> | Saltwater |
| Weakfish | <i>Cynoscion regalis</i> | Saltwater |
| Eels (Freshwater) - Family Anguillidae | | |
| American eel | <i>Anguilla rostrata</i> | Catadromous |
| Flounders (Lefteye) - Family Bothidae | | |
| Smallmouth flounder | <i>Etropus microstomus</i> | Saltwater |
| Summer flounder | <i>Paralichthys dentatus</i> | Saltwater |
| Weakfish | <i>Cynoscion regalis</i> | Saltwater |
| Flounders (Righteye) - Family Pleuronectidae | | |
| Winter flounder | <i>Pleuronectes americanus</i> | Saltwater |
| Gars - Family Lepisosteidae | | |
| Longnose gar | <i>Lepisosteus osseus</i> | Saltwater |
| Gobies - Family Gobiidae | | |
| Naked Goby | <i>Gobiosoma bosc</i> | Saltwater |
| Herrings - Family Clupeidae | | |
| Alewife | <i>Alosa pseudoharengus</i> | Anadromous |
| American shad | <i>Alosa sapidissima</i> | Anadromous |
| Atlantic menhaden | <i>Brevoortia tyrannus</i> | Anadromous |
| Blueback herring | <i>Alosa aestivalis</i> | Anadromous |
| Gizzard shad | <i>Dorosoma cepedianum</i> | Freshwater |
| Hickory shad | <i>Alosa mediocris</i> | Anadromous |
| Jacks - Family Carangidae | | |
| Crevalle jack | <i>Caranx hippos</i> | Saltwater |
| Killifishes- Family Cyprinodontidae | | |

**TABLE 2-1
HUDSON RIVER FISHES**

| Common Name | Scientific Name | Predominant Habitat |
|---|------------------------------------|----------------------------|
| Banded killifish | <i>Fundulus diaphanus</i> | Freshwater/brackish |
| Mummichog | <i>Fundulus heteroclitus</i> | Freshwater/brackish |
| Sheepshead minnow | <i>Cyprinodon variegatus</i> | Saltwater |
| Striped killifish | <i>Fundulus majalis</i> | Freshwater |
| Lampreys - Family Petromyzontidae | | |
| American brook lamprey | <i>Lampetra appendix</i> | Freshwater |
| Sea lamprey | <i>Petromyzon marinus</i> | Anadromous |
| Lizardfishes - Family Synodontidae | | |
| Inshore lizardfish | <i>Synodus foetens</i> | Saltwater |
| Mackerels - Family Scombridae | | |
| Atlantic mackerel | <i>Scomber scombrus</i> | Saltwater |
| Minnows - Family Cyprinidae | | |
| Blacknose dace | <i>Rhinichthys atratulus</i> | Freshwater |
| Bluntnose minnow | <i>Pimephales notatus</i> | Freshwater |
| Bridle shiner | <i>Notropis bifrenatus</i> | Freshwater |
| Central stoneroller | <i>Campostoma anomalum</i> | Freshwater |
| Comely shiner | <i>Notropis amoenus</i> | Freshwater |
| Common carp | <i>Cyprinus carpio</i> | Freshwater |
| Common shiner | <i>Notropis cornutus</i> | Freshwater |
| Creek chub | <i>Semotilus atromaculatus</i> | Freshwater |
| Cutlips minnow | <i>Exoglossum maxillingua</i> | Freshwater |
| Emerald shiner | <i>Notropis atherinoides</i> | Freshwater |
| Fallfish | <i>Semotilus corporalis</i> | Freshwater |
| Fathead minnow | <i>Pimephales promelas</i> | Freshwater |
| Golden shiner | <i>Notemigonus crysoleucas</i> | Freshwater |
| Goldfish | <i>Carassius auratus</i> | Freshwater |
| Hornyhead chub | <i>Nocomis biguttatus</i> | Freshwater |
| Longnose dace | <i>Rhinichthys cataractae</i> | Freshwater |
| Northern redbelly dace | <i>Rhinichthys atratulus</i> | Freshwater |
| Pearl dace | <i>Margariscus margarita</i> | Freshwater |
| Roseface shiner | <i>Notropis rubellus</i> | Freshwater |
| Satinfin shiner | <i>Notropis analostanus</i> | Freshwater |
| Silvery minnow | <i>Hybognathus regius</i> | Freshwater |
| Spotfin shiner | <i>Notropis spilopterus</i> | Freshwater |
| Spottail shiner | <i>Notropis hudsonius</i> | Freshwater |
| Mudminnows - Family Umbridae | | |
| Central mudminnow | <i>Umbra limi</i> | Freshwater |
| Eastern mudminnow | <i>Umbra pygmaea</i> | Freshwater |
| Mullets - Family Mugilidae | | |
| Striped mullet | <i>Mugil cephalus</i> | Saltwater |
| Needlefishes - Family Belontiidae | | |
| Atlantic needlefish | <i>Strongylura marina</i> | Saltwater |
| Perches - Family Percidae | | |
| Fantail darter | <i>Etheostoma flabellare</i> | Freshwater |
| Greenside darter | <i>Etheostoma blennioides</i> | Freshwater |
| Logperch | <i>Percina caprodes</i> | Freshwater |
| Tessellated darter | <i>Etheostoma olmstedi</i> | Freshwater |
| Walleye | <i>Stizostedion v. vitreum</i> | Freshwater |
| Yellow perch | <i>Perca flavescens</i> | Freshwater |
| Pikes - Family Esocidae | | |
| Chain pickerel | <i>Esox niger</i> | Freshwater |
| Northern pike | <i>Esox lucius</i> | Freshwater |
| Redfin pickerel | <i>Esox a. americanus</i> | Freshwater |
| Tiger muskellunge | <i>Northern pike X muskellunge</i> | Freshwater |
| Pipefishes - Family Syngnathidae | | |

**TABLE 2-1
HUDSON RIVER FISHES**

| Common Name | Scientific Name | Predominant Habitat |
|---|--|----------------------------|
| Lined seahorse | <i>Hippocampus erectus</i> | Saltwater |
| Northern pipefish | <i>Syngnathus fuscus</i> | Saltwater |
| Porgies - Family Sparidae | | |
| Scup | <i>Stenotomus chrysops</i> | Saltwater |
| Puffers - Family Tetraodontidae | | |
| Northern puffer | <i>Sphoeroides maculatus</i> | Saltwater |
| Sculpins - Family Cottidae | | |
| Grubby | <i>Myoxocephalus aeneus</i> | Freshwater |
| Longhorn sculpin | <i>Myoxocephalus octodecemspinosus</i> | Freshwater |
| Slimy sculpin | <i>Cottus cognatus</i> | Freshwater |
| Searobins - Family Triglidae | | |
| Northern searobin | <i>Prionotus carolinus</i> | Saltwater |
| Striped searobin | <i>Prionotus evolans</i> | Saltwater |
| Silversides - Family Atherinidae | | |
| Atlantic silverside | <i>Menidia menidia</i> | Saltwater |
| Brook silverside | <i>Labidesthes sicculus</i> | Freshwater |
| Smelts - Family Osmeridae | | |
| Rainbow smelt | <i>Osmerus mordax</i> | Anadromous |
| Soles - Family Soleidae | | |
| Hogchoker | <i>Trinectes maculatus</i> | Saltwater |
| Stickelbacks - Family Gasterosteidae | | |
| Brook stickleback | <i>Culaea inconstans</i> | Freshwater |
| Fourspine stickleback | <i>Apeltes quadracus</i> | Freshwater |
| Threespine stickleback | <i>Gasterosteus aculeatus</i> | Freshwater |
| Sturgeons - Family Acipenseridae | | |
| Atlantic sturgeon | <i>Acipenser oxyrinchus</i> | Freshwater |
| Shortnose sturgeon | <i>Acipenser brevirostrum</i> | Freshwater |
| Suckers - Family Castostomidae | | |
| Creek chubsucker | <i>Erimyzon oblongus</i> | Freshwater |
| Longnose sucker | <i>Catostomus catostomus</i> | Freshwater |
| Northern hogsucker | <i>Hypentelium nigricans</i> | Freshwater |
| Shorthead redhorse | <i>Moxostoma macrolepidotum</i> | Freshwater |
| White sucker | <i>Catostomus commersoni</i> | Freshwater |
| Sunfishes - Family Centrarchidae | | |
| Black crappie | <i>Pomoxis nigromaculatus</i> | Freshwater |
| Bluegill | <i>Lepomis macrochirus</i> | Freshwater |
| Bluespotted sunfish | <i>Enneacanthus gloriosus</i> | Freshwater |
| Green sunfish | <i>Lepomis cyanellus</i> | Freshwater |
| Largemouth bass | <i>Micropterus salmoides</i> | Freshwater |
| Pumpkinseed | <i>Lepomis gibbosus</i> | Freshwater |
| Redbreast sunfish | <i>Lepomis auritus</i> | Freshwater |
| Rock bass | <i>Ambloplites rupestris</i> | Freshwater |
| Smallmouth bass | <i>Micropterus dolomieu</i> | Freshwater |
| Warmouth | <i>Lepomis gulosus</i> | Freshwater |
| White crappie | <i>Pomoxis annularis</i> | Freshwater |
| Trouts - Family Salmonidae | | |
| Atlantic salmon | <i>Salmo salar</i> | Anadromous |
| Brook trout | <i>Salvelinus fontinalis</i> | Freshwater |
| Brown trout | <i>Salmo trutta</i> | Freshwater |
| Lake whitefish | <i>Coregonus clupeaformis</i> | Freshwater |
| Rainbow trout | <i>Oncorhynchus mykiss</i> (formerly <i>Salmo gairdneri</i>) | Freshwater |
| Round whitefish | <i>Prosopium cylindraceum</i> | Freshwater |
| Trout-perches - Family Percopsidae | | |

**TABLE 2-1
HUDSON RIVER FISHES**

| Common Name | Scientific Name | Predominant Habitat |
|---|-------------------------------|---------------------|
| Trout-perch | <i>Percopsis omiscomaycus</i> | Freshwater |
| Wrasses - Family Labridae | | |
| Cunner | <i>Tautoglabrus adspersus</i> | Saltwater |
| Tautog | <i>Tautoga onitis</i> | Saltwater |
| Notes: Fish are not found exclusively in predominant habitats. Source: Haynes and Frisch, 1993 and NYSDEC, 1989. | | |

TABLE 2-2

TYPICAL FISH AGGREGATIONS IN THE UPPER HUDSON RIVER

| | |
|--|---|
| <p>Widespread species American eel Blueback herring Alewife American shad Common carp Spottail shiner White perch Striped bass Pumpkinseed</p> | <p>Shore area Banded killifish Golden shiner Emerald shiner Gizzard shad Bay anchovy Bluegill Smallmouth bass Yellow perch</p> |
| <p>Rock pile White catfish Smallmouth bass Largemouth bass Rock bass Redbreast sunfish</p> | <p>Tailwater White sucker Golden shiner White catfish Largemouth bass Walleye</p> |
| <p>Vegetated backwater Brown bullhead Yellow perch Goldfish Golden shiner Banded killifish Largemouth bass White catfish White sucker Gizzard shad Northern pike Emerald shiner Rock bass Redbreast sunfish Bluegill Smallmouth bass</p> | <p>Major tributaries White sucker Smallmouth bass Redbreast sunfish Yellow perch Largemouth bass Goldfish Golden shiner Rock bass Bluegill Black crappie</p> |
| <p>Offshore shoals and channel Tessellated darter White catfish Brown bullhead</p> | <p>Hogchoker Shortnose sturgeon White sucker</p> |
| <p>Notes: Species are listed in order of abundance, excluding widespread species. Source: NYSDEC, 1989.</p> | |
| | |

TABLE 2-3

AMPHIBIANS POTENTIALLY FOUND ALONG THE HUDSON RIVER

| Common Name | Scientific Name |
|---|--------------------------------------|
| Order Caudata - Salamanders | |
| Allegheny Dusky Salamander | <i>Desmognathus ochrophaeus</i> |
| Blue-spotted Salamander | <i>Ambystoma laterale</i> |
| Common Mudpuppy | <i>Necturus maculosus</i> |
| Four-toed Salamander | <i>Hemidactylum scutatum</i> |
| Jefferson Salamander | <i>Ambystoma jeffersonianum</i> |
| Marbled Salamander | <i>Ambystoma opacum</i> |
| Northern Dusky Salamander | <i>Desmognathus fuscus</i> |
| Northern Spring Salamander | <i>Gyrinophilus p. porphyriticus</i> |
| Northern Redback Salamander | <i>Plethodon c. cinereus</i> |
| Northern Slimy Salamander | <i>Plethodon glutinosus</i> |
| Northern Two-lined Salamander | <i>Eurycea bislineata</i> |
| Northern Red Salamander | <i>Pseudotriton r. ruber</i> |
| Red-spotted or Eastern Newt | <i>Notophthalmus v. viridescens</i> |
| Spotted Salamander | <i>Ambystoma maculatum</i> |
| Order Anura - Toads and Frogs | |
| Toads | |
| Eastern American Toad | <i>Bufo a. americanus</i> |
| Eastern Spadefoot | <i>Scaphiopus holbrookii</i> |
| Fowler's Toad | <i>Bufo fowleri</i> |
| Family Ranida- True Frogs | |
| Bullfrog | <i>Rana catesbeiana</i> |
| Gray Treefrog | <i>Hyla versicolor</i> |
| Green Frog | <i>Rana clamitans melanota</i> |
| Northern Spring Peeper | <i>Pseudacris c. crucifer</i> |
| Northern Cricket Frog | <i>Acris c. crepitans</i> |
| Northern Leopard Frog | <i>Rana pipiens</i> |
| Pickerel Frog | <i>Rana palustris</i> |
| Source: New York State Amphibian and Reptile Atlas 1990- 1998 (NYSDEC, 1999). | |

TABLE 2-4

REPTILES POTENTIALLY FOUND ALONG THE HUDSON RIVER

| Common Name | Scientific Name |
|---|--|
| Turtles - Order Testudines | |
| Blanding's Turtle | <i>Emydoidea blandingii</i> |
| Bog turtle | <i>Clemmys muhlenbergi</i> |
| Common snapping turtle | <i>Chelydra serpentina</i> |
| Diamondback terrapin | <i>Malaclemys terrapin</i> |
| Eastern box turtle | <i>Terrapene carolina</i> |
| Map turtle | <i>Graptemys geographica</i> |
| Northern water snake | <i>Nerodia sipedon</i> |
| Painted turtle | <i>Chrysemys picta</i> |
| Red-eared Slider | <i>Trachemys scripta elegans</i> |
| Spotted turtle | <i>Clemmys guttata</i> |
| Stinkpot/ common musk turtle | <i>Sternotherus odoratus</i> |
| Wood turtle | <i>Clemmys insculpta</i> |
| Order Squamata - Lizards and Snakes | |
| Suborder Lacertilla - Lizards | |
| Five-lined Skink | <i>Eumeces fasciatus</i> |
| Northern Coal Skink | <i>Eumeces a. anthracinus</i> |
| Northern Fence Lizard | <i>Sceloporus undulatus hyacinthinus</i> |
| Suborder Serpente- Snakes | |
| Northern Water Snake | <i>Nerodia s. sipedon</i> |
| Northern Redbelly Snake | <i>Storeria o. occipitomaculata</i> |
| Common Garter Snake | <i>Thamnophis sirtalis</i> |
| Eastern Ribbon Snake | <i>Thamnophis sauritus</i> |
| Eastern Hognose Snake | <i>Heterodon platirhinos</i> |
| Northern Ringneck Snake | <i>Diadophis punctatus edwardsii</i> |
| Eastern Worm Snake | <i>Carphophis a. amoenus</i> |
| Northern Black Racer | <i>Coluber c. constrictor</i> |
| Smooth Green Snake | <i>Liochlorophis vernalis</i> |
| Black Rat Snake | <i>Elaphe o. obsoleta</i> |
| Eastern Milk Snake | <i>Lampropeltis t. triangulum</i> |
| Northern Copperhead | <i>Agkistrodon contortrix mokasen</i> |
| Timber Rattlesnake | <i>Crotalus horridus</i> |
| Northern Brown Snake | <i>Storeria d. dekayi</i> |
| Source: New York State Amphibian and Reptile Atlas 1990- 1998 (NYSDEC, 1999). | |

TABLE 2-5

BREEDING BIRDS OF THE HUDSON RIVER

| Common Name | Scientific Name |
|---|--|
| Acadian Flycatcher | <i>Empidonax vireescens</i> |
| Alder Flycatcher | <i>Empidonax alnorum</i> |
| American Bittern | <i>Botaurus lentiginosus</i> |
| American Robin | <i>Turdus migratorius</i> |
| American Kestrel | <i>Falco sparverius</i> |
| American Goldfinch | <i>Carduelis tristis</i> |
| American Coot | <i>Fulica americana*</i> |
| American Black Duck | <i>Anas rubripes</i> |
| American Crow | <i>Corvus brachyrhynchos</i> |
| American Redstart | <i>Setophaga ruticilla</i> |
| American Woodcock | <i>Scolopax minor</i> |
| Bank Swallow | <i>Riparia riparia</i> |
| Barn Swallow | <i>Hirundo rustica</i> |
| Barred Owl | <i>Strix varia</i> |
| Belted Kingfisher | <i>Ceryle alcyon</i> |
| Black-and-white Warbler | <i>Mniotilta varia</i> |
| Black-billed Cuckoo | <i>Coccyzus erythrophthalmus</i> |
| Black-capped Chickadee | <i>Parus atricapillus</i> |
| Black-crowned Night-Heron | <i>Nycticorax nycticorax</i> |
| Black-throated Blue Warbler | <i>Dendroica caerulescens</i> |
| Black-throated Green Warbler | <i>Dendroica virens</i> |
| Blackburnian Warbler | <i>Dendroica fusca</i> |
| Blue Jay | <i>Cyanocitta cristata</i> |
| Blue-gray Gnatcatcher | <i>Polioptila caerulea</i> |
| Blue-winged Teal | <i>Anas discors</i> |
| Blue-winged Warbler | <i>Vermivora pinus</i> |
| Blue-winged & Golden-winged Warbler Hybrids | <i>Vermivora pinus & Vermivora chrysoptera</i> |
| Bobolink | <i>Dolichonyx oryzivorus</i> |
| Broad-winged Hawk | <i>Buteo platypterus</i> |
| Brown Thrasher | <i>Toxostoma rufum</i> |
| Brown Creeper | <i>Certhia americana</i> |
| Brown-headed Cowbird | <i>Molothrus ater</i> |
| Canada Warbler | <i>Wilsonia canadensis</i> |
| Canada Goose | <i>Branta canadensis</i> |
| Carolina Wren | <i>Thryothorus ludovicianus</i> |
| Cattle Egret | <i>Bubulcus ibis*</i> |
| Cedar Waxwing | <i>Bombycilla cedrorum</i> |
| Cerulean Warbler | <i>Dendroica cerulea</i> |
| Chestnut-sided Warbler | <i>Dendroica pensylvanica</i> |
| Chimney Swift | <i>Chaetura pelagica</i> |
| Chipping Sparrow | <i>Spizella passerina</i> |
| Clapper Rail | <i>Rallus longirostris*</i> |
| Cliff Swallow | <i>Hirundo pyrrhonota</i> |
| Common Grackle | <i>Quiscalus quiscula</i> |
| Common Moorhen | <i>Gallinule chlorous</i> |
| Common Nighthawk | <i>Chordeiles minor</i> |
| Common Barn-Owl | <i>Tyto alba</i> |
| Common Yellowthroat | <i>Geothlypis trichas</i> |

TABLE 2-5

BREEDING BIRDS OF THE HUDSON RIVER

| Common Name | Scientific Name |
|-------------------------------|--------------------------------------|
| Common Merganser | <i>Mergus merganser</i> |
| Common Snipe | <i>Gallinago gallinago</i> |
| Cooper's Hawk | <i>Accipiter cooperii</i> |
| Dark-eyed Junco | <i>Junco hyemalis</i> |
| Double-crested Cormorant | <i>Phalacrocorax auritus</i> |
| Downy Woodpecker | <i>Picoides pubescens</i> |
| Eastern Meadowlark | <i>Sturnella magna</i> |
| Eastern Phoebe | <i>Sayornis phoebe</i> |
| Eastern Wood-Pewee | <i>Contopus virens</i> |
| Eastern Kingbird | <i>Tyrannus tyrannus</i> |
| Eastern Bluebird | <i>Sialia sialis</i> |
| Eastern Screech-Owl | <i>Otus asio</i> |
| European Starling | <i>Sturnus vulgaris</i> |
| Field Sparrow | <i>Spizella pusilla</i> |
| Fish Crow | <i>Corvus ossifragus</i> |
| Gadwall | <i>Anas strepera</i> |
| Glossy Ibis | <i>Plegadis falcinellus*</i> |
| Golden-crowned Kinglet | <i>Regulus satrapa</i> |
| Golden-winged Warbler | <i>Vermivora chrysoptera</i> |
| Grasshopper Sparrow | <i>Ammodramus savannarum</i> |
| Gray Catbird | <i>Dumetella carolinensis</i> |
| Great Horned Owl | <i>Bubo virginianus</i> |
| Great Egret | <i>Casmerodius albus*</i> |
| Great Blue Heron | <i>Ardea herodias</i> |
| Great Crested Flycatcher | <i>Myiarchus crinitus</i> |
| Great Black-backed Gull | <i>Larus marinus</i> |
| Green-backed Heron | <i>Butorides striatus</i> |
| Green-winged Teal | <i>Arias crecca</i> |
| Hairy Woodpecker | <i>Picoides villosus</i> |
| Henslow's Sparrow | <i>Ammodramus henslowii</i> |
| Hermit Thrush | <i>Catharus guttatus</i> |
| Herring Gull | <i>Larus argentatus*</i> |
| Hooded Merganser | <i>Lophodytes cucullatus</i> |
| Hooded Warbler | <i>Wilsonia citrina</i> |
| Horned Lark | <i>Eremophila alpestris</i> |
| House Sparrow | <i>Passer domesticus</i> |
| House Finch | <i>Carpodacus mexicanus</i> |
| House Wren | <i>Troglodytes aedon</i> |
| Indigo Bunting | <i>Passerina cyanea</i> |
| Kentucky Warbler | <i>Oporornis formosus</i> |
| Killdeer | <i>Charadrius vociferus</i> |
| King Rail | <i>Rallus elegans</i> |
| Laughing Gull | <i>Larus atricilla*</i> |
| Least Bittern | <i>Ixobrychus exilis</i> |
| Least Flycatcher | <i>Empidonax minimus</i> |
| Loggerhead Shrike | <i>Lanius ludovicianus</i> |
| Long-earned Owl | <i>Asio otus</i> |
| Louisiana Waterthrush | <i>Seiurus motacilla</i> |
| Magnolia Warbler | <i>Dendroica magnolia</i> |
| Mallard x American Black Duck | <i>Anas platyrhynchos x rubripes</i> |
| Mallard | <i>Anas platyrhynchos</i> |

TABLE 2-5

BREEDING BIRDS OF THE HUDSON RIVER

| Common Name | Scientific Name |
|-------------------------------|-----------------------------------|
| Marsh Wren | <i>Cistothorus palustris</i> |
| Mourning Dove | <i>Zenaida macroura</i> |
| Mute Swan | <i>Cygnus olor</i> |
| Nashville Warbler | <i>Vermivora ruficapilla</i> |
| Northern Bobwhite | <i>Colinus virginianus</i> |
| Northern Rough-winged Swallow | <i>Stelgidopteryx serripennis</i> |
| Northern Saw-whet Owl | <i>Aegolius acadicus</i> |
| Northern Oriole | <i>Icterus galbula</i> |
| Northern Pintail | <i>Anas acuta</i> |
| Northern Waterthrush | <i>Seiurus noveboracensis</i> |
| Northern Goshawk | <i>Accipiter gentilis</i> |
| Northern Flicker | <i>Colaptes auratus</i> |
| Northern Mockingbird | <i>Mimus polyglottos</i> |
| Northern Harrier | <i>Circus cyaneus</i> |
| Northern Cardinal | <i>Cardinalis cardinalis</i> |
| Orchard Oriole | <i>Icterus spurius</i> |
| Osprey | <i>Pandion haliaetus</i> |
| Ovenbird | <i>Seiurus aurocapillus</i> |
| Pied-billed Grebe | <i>Podilymbus podiceps</i> |
| Pileated Woodpecker | <i>Dryocopus pileatus</i> |
| Pine Warbler | <i>Dendroica pinus</i> |
| Prairie Warbler | <i>Dendroica discolor</i> |
| Purple Martin | <i>Progne subis</i> |
| Purple Finch | <i>Carpodacus purpureus</i> |
| Red-bellied Woodpecker | <i>Melanerpes carolinus</i> |
| Red-breasted Merganser | <i>Mergus serrator</i> |
| Red-breasted Nuthatch | <i>Sitta canadensis</i> |
| Red-eyed Vireo | <i>Vireo olivaceus</i> |
| Red-headed Woodpecker | <i>Melanerpes erythrocephalus</i> |
| Red-shouldered Hawk | <i>Buteo lineatus</i> |
| Red-tailed Hawk | <i>Buteo jamaicensis</i> |
| Red-winged Blackbird | <i>Agelaius phoeniceus</i> |
| Ring-necked Pheasant | <i>Phasianus colchicus</i> |
| Rock Dove | <i>Columba livia</i> |
| Rose-breasted Grosbeak | <i>Pheucticus ludovicianus</i> |
| Ruby-throated Hummingbird | <i>Archilochus colubris</i> |
| Ruddy Duck | <i>Oxyura jamaicensis</i> |
| Ruffed Grouse | <i>Bonasa umbellus</i> |
| Rufous-sided Towhee | <i>Pipilo erythrophthalmus</i> |
| Savannah Sparrow | <i>Passerculus sandwichensis</i> |
| Scarlet Tanager | <i>Piranga olivacea</i> |
| Sharp-shinned Hawk | <i>Accipiter striatus</i> |
| Snowy Egret | <i>Egretta thula*</i> |
| Solitary Vireo | <i>Vireo solitarius</i> |
| Song Sparrow | <i>Melospiza melodia</i> |
| Sora | <i>Porzana carolina</i> |
| Spotted Sandpiper | <i>Actitis macularia</i> |
| Swamp Sparrow | <i>Melospiza georgiana</i> |
| Tree Swallow | <i>Tachycineta bicolor</i> |
| Tufted Titmouse | <i>Parus bicolor</i> |
| Turkey Vulture | <i>Cathartes aura</i> |

TABLE 2-5

BREEDING BIRDS OF THE HUDSON RIVER

| Common Name | Scientific Name |
|--|--------------------------------|
| Upland Sandpiper | <i>Bartramia longicauda</i> |
| Veery | <i>Catharus fuscescens</i> |
| Vesper Sparrow | <i>Pooecetes gramineus</i> |
| Virginia Rail | <i>Rallus limicola</i> |
| Warbling Vireo | <i>Vireo gilvus</i> |
| Western Meadowlark | <i>Sturnella neglecta</i> |
| Whip-poor-will | <i>Caprimulgus vociferus</i> |
| White-breasted Nuthatch | <i>Sitta carolinensis</i> |
| White-eyed Vireo | <i>Vireo griseus</i> |
| White-throated Sparrow | <i>Zonotrichia albicollis</i> |
| Wild Turkey | <i>Meleagris gallopavo</i> |
| Willow Flycatcher | <i>Empidonax traillii</i> |
| Winter Wren | <i>Troglodytes troglodytes</i> |
| Wood Duck | <i>Aix sponsa</i> |
| Wood Thrush | <i>Hylocichla mustelina</i> |
| Worm-eating Warbler | <i>Helmitheros vermivorus</i> |
| Yellow Warbler | <i>Dendroica petechia</i> |
| Yellow-bellied Sapsucker | <i>Sphyrapicus varius</i> |
| Yellow-billed Cuckoo | <i>Coccyzus americanus</i> |
| Yellow-breasted Chat | <i>Icteria virens</i> |
| Yellow-crowned Night-Heron | <i>Nycticorax violaceus*</i> |
| Yellow-rumped Warbler | <i>Dendroica coronata</i> |
| Yellow-throated Vireo | <i>Vireo flavifrons</i> |
| Notes: * coastal breeding birds | |
| Source: Andrlle, R.F. and J.R. Carroll (Editors). 1988. The Atlas of Breeding Birds in New York State. Cornell University Press. Ithaca, New York. 551 pp. | |

TABLE 2-6

MAMMALS POTENTIALLY FOUND ALONG THE HUDSON RIVER

| Common Name | Scientific Name |
|--|---|
| Order Artiodactyla - Even-toed Ungulates | |
| Family Cervidae - Cervids | |
| Whitetail deer | <i>Odocoileus virginianus</i> |
| Order Carnivora | |
| Family Canidae - Canids | |
| Coyote | <i>Canis latrans</i> |
| Gray fox | <i>Urocyon cinereoargenteus</i> |
| Red fox | <i>Vulpes vulpes</i> |
| Family Felidae- Cats | |
| Bobcat | <i>Lynx rufus</i> |
| Family Mustelidae - Weasels | |
| Common striped skunk | <i>Mephitis mephitis</i> |
| Ermine | <i>Martes erminea</i> |
| Fisher | <i>Martes pennanti</i> |
| Least weasel | <i>Martes nivalis</i> |
| Longtail weasel | <i>Mustela frenata</i> |
| Marten | <i>Martes americana</i> |
| Mink | <i>Mustela vison</i> |
| River otter | <i>Lutra canadensis</i> |
| Family Procyonidae- Raccoons | |
| Raccoon | <i>Procyon lotor</i> |
| Family Ursidae - Bears | |
| Black bear | <i>Ursus americanus</i> |
| Order Chiroptera - Bats | |
| Family Vespertilionidae - Vespertilionid Bats | |
| Big brown bat | <i>Eptesicus fuscus</i> |
| Eastern pipistrelle | <i>Pipistrellus subflavus</i> |
| Eastern small-footed myotis | <i>Myotis leibii</i> |
| Evening bat | <i>Nycticeius humeralis</i> |
| Hoary bat | <i>Lasiurus cinereus</i> |
| Indiana myotis | <i>Myotis sodalis</i> |
| Little brown bat | <i>Myotis lucifugus</i> |
| Keen's myotis | <i>Myotis keenii</i> or <i>M. septentrionalis</i> |
| Red bat | <i>Lasiurus borealis</i> |
| Silver-haired bat | <i>Lasionycteris noctivagans</i> |
| Order Insectivora- Insectivores | |
| Family Soricidae - Shrews | |
| Least shrew | <i>Cryptotis parva</i> |
| Masked shrew | <i>Sorex cinereus</i> |
| Northern short-tailed shrew | <i>Blarina brevicauda</i> |
| Pygmy shrew | <i>Sorex hoyi</i> |
| Rock shrew | <i>Sorex dispar</i> |
| Smokey shrew | <i>Sorex fumeus</i> |
| Water shrew | <i>Sorex palustris</i> |
| Family Talpidae - Moles | |
| Eastern mole | <i>Scalopus aquaticus</i> |
| Hairy-tailed mole | <i>Parascalops breweri</i> |
| Star-nosed mole | <i>Condylura cristata</i> |

TABLE 2-6

MAMMALS POTENTIALLY FOUND ALONG THE HUDSON RIVER

| Common Name | Scientific Name |
|--|----------------------------------|
| Order Lagomorpha | |
| Family Leporidae - Hares and Rabbits | |
| Black-tailed jackrabbit | <i>Lepus californicus</i> |
| Cottontail | <i>Sylvilagus floridanus</i> |
| European rabbit | <i>Orytolagus cuniculus</i> |
| New England cottontail | <i>Sylvilagus transitionalis</i> |
| Snowshoe hare | <i>Lepus americanus</i> |
| Order Marsupialia - Marsupials | |
| Family Didelphidae - Opossums | |
| Virginia opossum | <i>Didelphis virginiana</i> |
| Order Rodentia - Rodents | |
| Family Castoridae - Beavers | |
| Beaver | <i>Castor canadensis</i> |
| Family Cricetidae - Cricetids | |
| Deer mouse | <i>Peromyscus maniculatus</i> |
| Meadow vole | <i>Microtus pennsylvanicus</i> |
| Muskrat | <i>Ondatra zibethicus</i> |
| Pine vole | <i>Microtus pientorum</i> |
| Rock or yellow nose vole | <i>Microtus chrotorrhinus</i> |
| Southern bog lemming | <i>Synaptomys cooperi</i> |
| Southern red-backed vole | <i>Clethrionomys gapperi</i> |
| White-footed mouse | <i>Peromyscus leucopus</i> |
| Family Erethizontidae - New World Porcupine | |
| Porcupine | <i>Erethizon dorsatum</i> |
| Family Muridae - Murids | |
| Norway rat | <i>Rattus norvegicus</i> |
| Black rat | <i>Rattus rattus</i> |
| House mouse | <i>Mus musculus</i> |
| Eastern woodrat | <i>Neotoma magister</i> |
| Family Myocastoridae - Myocastorids | |
| Nutria | <i>Myocastor coypus</i> |
| Family Sciuridae - Squirrels | |
| Chipmunk | <i>Tamias striatus</i> |
| Eastern gray squirrel | <i>Sciurus carolinensis</i> |
| Fox squirrel | <i>Sciurus niger</i> |
| Northern flying squirrel | <i>Glaucomys sabrinus</i> |
| Red squirrel | <i>Tamiasciurus hudsonicus</i> |
| Southern flying squirrel | <i>Glaucomys volans</i> |
| Woodchuck | <i>Marmota monax</i> |
| Family Zapodidae - Jumping Mice | |
| Meadow jumping mouse | <i>Zapus hudsonius</i> |
| Woodland jumping mouse | <i>Napaeozapus insignis</i> |
| Sources: NYSM, 1999; NYSDOS, 1990. | |

TABLE 2-7

ASSESSMENT AND MEASUREMENT ENDPOINTS

| Assessment Endpoint | Specific Ecological Receptor (“Endpoint Species”) | Measures | |
|--|--|---|---|
| | | Exposure | Effect |
| Benthic community structure as food source for local fish and wildlife. | · Benthic macroinvertebrate community | · Ecological community indices (diversity, evenness, dominance) · PCB levels in sediments and water column | · Differences in benthic community indices · Exceedance of ambient water quality criteria (AWQC) and sediment guidelines |
| Survival, growth, and reproduction of local forage fish populations. | · Spottail shiner · Pumpkinseed | · Measured PCB body burdens · Modeled PCB body burdens · PCB concentrations in sediments and water column | · Estimated exceedance of TRVs ¹ · Exceedance of AWQC and sediment guidelines · Field observations |
| Survival, growth, and reproduction of local piscivorous/semi-piscivorous fish populations. | · Yellow perch · White perch · Largemouth bass · Striped bass | · Measured PCB body burdens · Modeled PCB body burdens · PCB concentrations in sediments and water column | · Estimated exceedance of TRVs · Exceedance of AWQC and sediment guidelines · Field observations |
| Survival, growth, and reproduction of local omnivorous fish populations. | · Shortnose sturgeon · Brown bullhead | · Measured PCB body burdens · Modeled PCB body burdens · PCB concentrations in sediments and water column | · Estimated exceedance of TRVs · Exceedance of AWQC and sediment guidelines · Field observations |
| Protection (i.e., survival and reproduction) of insectivorous birds and mammals. | · Tree swallow · Little brown bat | · Measured PCB concentrations in prey items (aquatic insects/benthic invertebrates) · Modeled PCB concentrations in prey items (aquatic insects) · PCB concentrations in the water column | · Estimated exceedance of TRVs · Exceedance of AWQC for the protection of wildlife · Field observations |
| Protection (i.e., survival and reproduction) of waterfowl. | · Mallard | · Measured PCB concentrations in prey (invertebrates, macrophytes) · Modeled PCB concentrations in prey (invertebrates, macrophytes) · PCB concentrations in the water column | · Estimated exceedance of TRVs · Exceedance of AWQC for the protection of wildlife · Field observations |
| Protection of piscivorous/semi-piscivorous birds and mammals. | · Belted kingfisher · Great blue heron · Mink · River Otter | · Measured PCB concentrations in prey (forage fish, invertebrates) · Modeled PCB concentrations in prey (forage fish, invertebrates) · PCB concentrations in sediments and water column | · Estimated exceedance of TRVs · Exceedance of AWQC for the protection of wildlife · Field observations |

TABLE 2-7

ASSESSMENT AND MEASUREMENT ENDPOINTS

| Assessment Endpoint | Specific Ecological Receptor (“Endpoint Species”) | Measures | |
|---|---|--|---|
| | | Exposure | Effect |
| Protection of omnivorous mammals. | · Raccoon | · Measured PCB concentrations in prey items (fish, invertebrates) · PCB concentrations in the water column | · Estimated exceedance of TRVs · Exceedance of AWQC for the protection of wildlife · Field observations |
| Protection of endangered and threatened species. | · Bald eagle · Shortnose sturgeon | · Modeled PCB body burdens (sturgeon) · Measured PCB concentrations in prey (fish) · Modeled PCB concentrations in prey (fish) · PCB concentrations in sediments and water column | · Estimated exceedance of TRVs · Exceedance of AWQC sediment guidelines for the protection of wildlife · Field observations |
| Protection of significant habitats. | · Hudson River NERR · Selected NYSDOS significant habitats | · PCB concentrations in sediments and water column | · Exceedance of federal and state AWQC and sediment guidelines |
| Notes: 1: Individual-level effects are considered to occur when the TQ is greater to or equal to one. Receptor species are surrogates, representative of a wide range of species likely to use the Hudson River as habitat or foraging source. | | | |

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TABLE 2-8

HUDSON RIVER RECEPTOR SPECIES

| Receptor Species | Habitat/Feeding Characteristics | Similar Feeding Groups (general comparison) | |
|--|--|---|--|
| Benthic Invertebrate Community | Benthic Macroinvertebrates- Planktivorous, Deposit-feeders, Omnivorous | | |
| Spottail Shiner Pumpkinseed | Nektonic Forage Fish - Planktivorous, Insectivorous, Omnivorous | <u>Fish</u> Sunfishes Minnows Killifish River Herring | <u>Amphibians</u> Medium-sized Ranids (True Frogs) |
| Yellow Perch White Perch Largemouth Bass Striped Bass | Nektonic Fishes- Piscivorous | Basses Bluefish Weakfish | |
| Brown Bullhead Shortnose Sturgeon | Aquatic Feeders - Omnivorous, Scavengers, Detritivores | <u>Fish</u> Catfish Sturgeon Flatfishes Eels | <u>Amphibians</u> Salamanders Newts Larger Ranids <u>Reptiles</u> Turtles |
| Tree Swallow | Perching Birds of Wetlands- Insectivorous | Thrushes Wrens Sparrows | Flycatchers Jays Blackbirds |
| Mallard | Swimming Birds - Aquatic Herbivorous/Insectivorous | <u>Birds</u> Ducks Geese Swans Coots | <u>Mammals</u> Muskrat |
| Great Blue Heron | Wading Birds - Piscivorous | Shorebirds Herons, egrets, and bitterns Cormorants Mergansers Rails | |
| Belted Kingfisher | Wide-ranging River Birds - Piscivorous | Gulls Kingfishers | |
| Bald Eagle | Raptors (Birds of Prey) - Piscivorous/Carnivorous/ Scavengers | Eagles Hawks Falcons Osprey | |
| Little Brown Bat | Flying Mammals -Insectivorous | Bats | |
| Mink | Semi-piscivorous/Carnivorous Mammals | Other mustelids | |
| River Otter | Piscivorous mammals | Harbor Seal | |
| Raccoon | Facultative Wetland Mammals - Omnivorous | Foxes Dogs Cats | |
| Notes: Habitat/feeding characteristics are generalized and may not apply to all individuals of a group or species. | | | |

TABLE 2-9

RECEPTOR TROPHIC LEVELS, EXPOSURE PATHWAYS, AND FOOD SOURCES

| Endpoint Species | Level | Exposure Pathways | General Food Sources |
|-------------------------|--------------|---|--|
| FISH | | | |
| Benthic Invertebrates | 1 | A Direct contact with sediments A Direct contact with interstitial water A Direct contact with water (epibenthic and filter feeders) | Species dependent; food sources include detritus, plants, other invertebrates, zooplankton and phytoplankton in interstitial water |
| Pumpkinseed | 2 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | 80% pelagic invertebrates 20% benthic invertebrates |
| Spottail Shiner | 2 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | 50% benthic invertebrates 50% pelagic invertebrates |
| Brown Bullhead | 2 | A Direct contact with water (respiration, Dermal) A Food chain exposure (primarily Sediment-based) A Direct contact with sediment | 90% benthic invertebrates, <10% pelagic invertebrates or forage fish |
| Yellow Perch | 2-3 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | <10% forage fish 20-30% benthic invertebrates 60-80% pelagic invertebrates |
| White Perch | 2-3 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | 10-20% forage fish 30-40% benthic invertebrates 50-60% pelagic invertebrates |
| Largemouth Bass | 3 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | 90% forage fish 10% benthic invertebrates |
| Striped Bass | 3 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | Predominantly forage fish |

TABLE 2-9

RECEPTOR TROPHIC LEVELS, EXPOSURE PATHWAYS, AND FOOD SOURCES

| Endpoint Species | Level | Exposure Pathways | General Food Sources |
|-------------------------|--------------|---|--|
| Shortnose Sturgeon | 3 | A Direct contact with water (respiration, Dermal) A Food chain exposure (both water and Sediment-based) A Direct contact with sediments | Predominantly forage fish |
| BIRDS | | | |
| Tree Swallow | 2 | A Water ingestion A Food chain exposure | Emergent aquatic and terrestrial insects |
| Mallard | 2 | A Water ingestion A Food chain exposure | Vegetation, benthic invertebrates |
| Belted Kingfisher | 3 | A Water ingestion A Food chain exposure | Forage fish, aquatic invertebrates |
| Great Blue Heron | 3 | A Water ingestion A Food chain exposure A Direct contact with sediments | Forage fish, aquatic invertebrates |
| Bald Eagle | 4 | A Ingestion of water A Food chain exposures | Forage fish, small mammals, carrion |
| MAMMALS | | | |
| Little Brown Bat | 2 | A Ingestion of water A Food chain exposure | Emergent aquatic and terrestrial insects |
| Raccoon | 3 | A Ingestion of water A Food chain exposure A Direct contact with sediments | Forage fish, insects, invertebrates |
| Mink | 4-5 | A Ingestion of water A Food chain exposure A Direct contact with sediments | Forage fish, invertebrates, small mammals |
| River Otter | 4-5 | A Ingestion of water A Food chain exposures A Direct contact with sediments | Forage and piscivorous fish, waterfowl, frogs, invertebrates |

TABLE 2-10

NYS RARE AND LISTED SPECIES AND HABITATS OCCURRING IN THE VICINITY OF THE HUDSON RIVER

| Common Name | Scientific Name | NYS Status | State Rank | Precision Value |
|---|--|-------------|------------|-----------------|
| Plants - known occurrences (i.e., precision value S) | | | | |
| American waterwort | <i>Elantine americana</i> | Endangered | S1 | S |
| Bicknell's sedge | <i>Carex bicknelli</i> | Rare | S2/S3 | S |
| Carey's smartweed | <i>Polygonum careyi</i> | Unprotected | S2 | S |
| Clustered sedge | <i>Carex cumulata</i> | Rare | S2S3 | S |
| Corn-salad | <i>Valerianella umbilicata</i> | Unprotected | SH | S |
| Davis' sedge | <i>Carex davisii</i> | Rare | S1 | S |
| Estuary beggar-ticks | <i>Bidens bidentoides</i> | Threatened | S3 | S |
| False hop sedge | <i>Carex lupiformes</i> | Rare | S3 | S |
| Fissidens (non-vascular) | <i>Fissidens Fontanus</i> | Unprotected | S3? | S |
| Frank sedge | <i>Carex frankii</i> | Unprotected | S1 | S |
| Glaucous sedge | <i>Carex Flaccosperma var. glaucodea</i> | Rare | S1 | S |
| Golden club | <i>Orontium aquaticum</i> | Unprotected | S2 | S |
| Golden seal | <i>Hydrastis canadensis</i> | Threatened | S2 | S |
| Gypsy-wort | <i>Lycopus rubellus</i> | Unprotected | S1 | S |
| Heartleaf plantain | <i>Plantago cordata</i> | Threatened | S3 | S |
| Illinois pinweed | <i>Lechea racemulosa</i> | Rare | S3 | S |
| Liliaeopsis | <i>Liliaeopsis chinensis</i> | Unprotected | S2 | S |
| Lined sedge | <i>Carex striatula</i> | Unprotected | S1 | S |
| Long's bittercress | <i>Cardamine longii</i> | Unprotected | S2 | S |
| Marsh straw sedge | <i>Carex hormathodes</i> | Rare | S2/S3 | S |
| Midland sedge | <i>Carex mesocorea</i> | Unprotected | S1 | S |
| Mock-pennyroyal | <i>Hedeoma hispidum</i> | Rare | S2/S3 | S |

TABLE 2-10

NYS RARE AND LISTED SPECIES AND HABITATS OCCURRING IN THE VICINITY OF THE HUDSON RIVER

| Common Name | Scientific Name | NYS Status | State Rank | Precision Value |
|-------------------------------|---|-------------|------------|-----------------|
| Narrow-leaved sedge | <i>Carex amphibola</i> var. <i>amphibola</i> | Unprotected | S1 | S |
| Saltmarsh bulrush | <i>Scirpus novae-angliae</i> | Endangered | S1 | S |
| Schweinitz's flatsedge | <i>Cyperus schweinitzii</i> | Rare | S3 | S |
| Slender crabgrass | <i>Digitaria filiformis</i> | Rare | S2 | S |
| Small-flowered crowfoot | <i>Ranunculus micranthus</i> | Unprotected | S2 | S |
| Smooth bur-marigold | <i>Bidens laevis</i> | Rare | S2 | S |
| Southern yellow flax | <i>Linum medium</i> var. <i>texanum</i> | Threatened | S2 | S |
| Southern dodder | <i>Cuscuta obtusiflora</i> car. <i>glandulosa</i> | Unprotected | S1 | S |
| Spongy arrowhead | <i>Sagittaria calycina</i> var. <i>spongiosa</i> | Rare | S2 | S |
| Starwort | <i>Callitriche terrestris</i> | Unprotected | S2S3 | S |
| Swamp lousewort | <i>Pedicularis lanceolata</i> | Rare | S2 | S |
| Swamp cottonwood | <i>Populus heterophylla</i> | Threatened | S2 | S |
| Taxiphyllum (non-vascular) | <i>Taxiphyllum taxirameum</i> | Unprotected | S1 | S |
| Violet wood-sorrel | <i>Oxalis violacea</i> | Unprotected | S1S2 | S |
| Violet lespedeza | <i>Lespedeza violacea</i> | Rare | S3 | S |
| Water pigmyweed | <i>Crassula aquatica</i> | Endangered | S1 | S |
| Weak stellate sedge | <i>Carex seorsa</i> | Rare | S2 | S |
| Invertebrates | | | | |
| American rubyspot dragonfly | <i>Hetaerina americana</i> | Unprotected | S2/S3 | S |
| Arrowhead spiketail dragonfly | <i>Cordulegaster obliqua</i> | Unprotected | S2S3 | S |
| Gray petaltail dragonfly | <i>Tachopteryx thoreyi</i> | Unprotected | S2 | S |

TABLE 2-10

NYS RARE AND LISTED SPECIES AND HABITATS OCCURRING IN THE VICINITY OF THE HUDSON RIVER

| Common Name | Scientific Name | NYS Status | State Rank | Precision Value |
|---|---------------------------------|----------------------------|-------------|-----------------|
| Tawny emperor butterfly | <i>Asterocampa clyton</i> | Unprotected | S3 | S |
| Riverine clubtail | <i>Stylurus amnicola</i> | Unprotected | SH | M |
| Fish | | | | |
| Shortnose sturgeon | <i>Acipenser brevirostrum</i> | Endangered | S1 | S |
| Bluespotted sunfish | <i>Enneacanthus gloriosus</i> | Unprotected | S2 | M |
| Reptiles | | | | |
| Bog turtle | <i>Clemmys muhlenbergii</i> | Endangered | S2 | M |
| Blanding's turtle | <i>Emydoidea blandingii</i> | Threatened | S2 | M |
| Fence lizard | <i>Sceloporus undulatus</i> | Unprotected | S1 | S |
| Timber rattlesnake | <i>Crotalus horridus</i> | Threatened | S3 | M |
| Birds | | | | |
| Peregrine falcon | <i>Falco peregrinus</i> | Endangered | S2 | S |
| Bald eagle | <i>Haliaeetus Leucocephalus</i> | Endangered | S1B, S1N | S |
| King rail | <i>Rallus elegans</i> | Protected | S1 | M |
| Barn Owl | <i>Tyto alba</i> | Protected- Special Concern | S3 | M |
| Short-eared owl | <i>Asio flammeus</i> | Protected- Special Concern | S2 | S |
| Osprey | <i>Pandion halietus</i> | Threatened | S4 | M |
| Mammals | | | | |
| Eastern woodrat | <i>Neotoma magister</i> | Endangered | SH | M |
| Communities | | | | |
| 19 Freshwater Intertidal Mudflats Communities | | | | S |
| 25 Freshwater Tidal Marsh Communities | | | | S |

TABLE 2-10

NYS RARE AND LISTED SPECIES AND HABITATS OCCURRING IN THE VICINITY OF THE HUDSON RIVER

| Common Name | Scientific Name | NYS Status | State Rank | Precision Value |
|---|------------------------|-------------------|-------------------|------------------------|
| 9 Freshwater Tidal Swamp Communities | | | | S |
| 8 Freshwater Intertidal Shore Communities | | | | S |
| 7 Brackish Intertidal Mudflats Communities | | | | S |
| 7 Brackish Tidal Marsh Communities | | | | S |
| 1 Brackish Subtidal Aquatic Bed Community | | | | S |
| 1 Calcareous Cliff Community | | | | S |
| Areas of Concern | | | | |
| 16 Anadromous Fish Concentration Areas | | | | S |
| 12 Waterfowl Concentration Areas | | | | S |
| 3 Raptor Concentration Areas | | | | S |
| 1 Warm Water Fish Concentration Area | | | | S |
| <p>Notes: State Rank: S1 = Typically 5 or fewer occurrences, very few remaining individuals, acres or miles of stream in NYS S2 = Typically 6 to 20 occurrences, very few remaining individuals, acres or miles of stream in NYS S3 = Typically 21 to 100 occurrences, limited acreage or miles of stream in NYS S4 = Apparently secure in NYS S5 = Demonstrably secure in NYS Precision Rank: A precision value of "S" indicates that a species is known to be found along the Hudson River. A precision value of "M" indicates that a species may occur along the Hudson River in an appropriate habitat. Source: NYSDEC, May 1999.</p> | | | | |

TABLE 2-11

HUDSON RIVER SIGNIFICANT HABITATS

| |
|---|
| <p><u>Freshwater Habitats</u></p> <p>Normans Kill Papascanee Marsh and Creek Shad and Schermerhorn Island Schodack and Houghtaling Islands and Schodack Creek Coeymans Creek Hannacroix Creek Mill Creek Wetlands Stuyvesant Marshes* Coxsackie Creek Coxsackie Island Backwater Stockport Creek and Flats Vosburgh Swamp and Middle Ground Flats Roger's Island Catskill Creek Ramshorn Marsh Inbocht Bay and Duck Cove Roeliff-Jansen Kill Smith's Landing Cementon* Germantown/Clermont Flats Esopus Estuary North and South Tivoli Bays Mudder Kill* The Flats Roundout Creek Kingston Deepwater Habitat Vanderburgh Cove and Shallows Esopus Meadows Poughkeepsie Deepwater Habitat Crum Elbow Marsh*</p> |
| <p><u>Brackish Water Habitats</u></p> <p>Wappinger Creek Fishkill Creek Moodna Creek Hudson River Miles 44-56 Constitution Marsh Iona Island Marsh Camp Smith Marsh and Annsville Creek*</p> |
| <p><u>Salt Water Habitats</u></p> <p>Haverstraw Bay Croton River and Bay Piermont Marsh</p> |
| <p>Notes: * Indicates an area that is recognized by the NYS Natural Heritage Program as containing rare/important species or communities, but is not a designated Significant Habitat. Source: NYSDOS, 1990.</p> |

**TABLE 3-1
AVERAGE PROPORTION OF FISH-BASED TEQ CONGENERS USING EPA 1993 DATASET AND USFWS 1995 DATASET**

| | BZ#77 | BZ#81 | BZ#105 | BZ#114 | BZ#118 | BZ#123 | BZ#126 | BZ#156 | BZ#157 | BZ#167 | BZ#169 | BZ#189 |
|------------------|--------------|--------------|---------------|--------|---------------|--------|---------------|--------|--------|--------|--------|--------|
| Upper River Mean | 0.28 | | 0.06 | 0.01 | 0.11 | 0.00 | 0.52 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 |
| Lower River Mean | 0.05 | | 0.02 | 0.00 | 0.05 | 0.00 | 0.85 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 |
| Whole River Mean | 0.15 | | 0.03 | 0.00 | 0.07 | 0.00 | 0.71 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 |
| Egg Mean | 0.32 | 0.11 | 0.04 | 0.01 | 0.07 | 0.03 | 0.40 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Chick Mean | 0.38 | 0.13 | 0.04 | 0.01 | 0.08 | 0.03 | 0.33 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Odonate Mean | 0.34 | 0.05 | 0.03 | 0.00 | 0.05 | 0.01 | 0.49 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Insect Mean | 0.34 | 0.11 | 0.04 | 0.01 | 0.05 | 0.02 | 0.42 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Source: TAMS/Gradient Database Release 4.1b

Note: Dominant congeners are bold.

TABLE 3-2: FRACTION OF TRI+ CHLORINATED CONGENERS EXPRESSED AS TOXIC EQUIVALENCIES (TEQ)

| <<<<<<< ---- Whole Water Concentrations ---- >>>>>>> | | | | | | |
|---|----------|----------|----------|----------|----------|----------|
| | Value 1 | Value 2 | Value 1 | Value 2 | Value 1 | Value 2 |
| | Fish | Fish | Mammal | Mammal | Avian | Avian |
| Upper River | 2.82E-04 | 4.43E-06 | 6.33E-03 | 4.73E-04 | 8.17E-03 | 8.92E-04 |
| Lower River | 2.92E-05 | 4.72E-07 | 6.53E-04 | 5.49E-05 | 8.50E-04 | 8.83E-05 |
| Whole River | 2.57E-04 | 4.03E-06 | 5.75E-03 | 4.31E-04 | 7.43E-03 | 8.10E-04 |
| <<<<<<< ---- Fish Concentrations ---- >>>>>>> | | | | | | |
| | Value 1 | Value 2 | Value 1 | Value 2 | Value 1 | Value 2 |
| | Fish | Fish | Mammal | Mammal | Avian | Avian |
| Upper River | 2.27E-06 | 1.22E-06 | 4.29E-05 | 1.78E-05 | 2.57E-04 | 2.36E-04 |
| Lower River | 5.46E-06 | 1.20E-06 | 1.17E-04 | 2.19E-05 | 1.90E-04 | 1.02E-04 |
| Whole River | 4.13E-06 | 1.21E-06 | 8.59E-05 | 2.02E-05 | 2.18E-04 | 1.58E-04 |
| <<<<<<< ---- Sediment Concentrations ---- >>>>>>> | | | | | | |
| | Value 1 | Value 2 | Value 1 | Value 2 | Value 1 | Value 2 |
| | Fish | Fish | Mammal | Mammal | Avian | Avian |
| Upper River | 3.94E-05 | 5.75E-06 | 7.78E-04 | 3.86E-05 | 2.78E-03 | 2.10E-03 |
| Lower River | 1.46E-05 | 7.86E-06 | 2.46E-04 | 8.33E-05 | 2.19E-03 | 2.05E-03 |
| Whole River | 2.68E-05 | 6.83E-06 | 5.06E-04 | 6.15E-05 | 2.48E-03 | 2.08E-03 |
| <<<<<<< ---- Dissolved Water Concentrations ---- >>>>>>> | | | | | | |
| | Value 1 | Value 2 | Value 1 | Value 2 | Value 1 | Value 2 |
| | Fish | Fish | Mammal | Mammal | Avian | Avian |
| Upper River | 4.10E-06 | 1.49E-06 | 7.62E-03 | 2.80E-04 | 9.38E-03 | 4.27E-05 |
| Lower River | 5.36E-07 | 5.10E-07 | 5.12E-04 | 8.45E-05 | 6.15E-04 | 4.50E-05 |
| Whole River | 3.74E-06 | 1.39E-06 | 6.90E-03 | 2.60E-04 | 8.50E-03 | 4.30E-05 |
| <<<<<<< ---- Benthic Invertebrate Concentrations ---- >>>>>>> | | | | | | |
| | Value 1 | Value 2 | Value 1 | Value 2 | Value 1 | Value 2 |
| | Fish | Fish | Mammal | Mammal | Avian | Avian |
| Upper River | 1.38E-06 | 1.06E-07 | 2.97E-05 | 1.55E-06 | 5.33E-05 | 1.88E-05 |
| Lower River | 4.82E-06 | 1.05E-07 | 1.08E-04 | 2.07E-06 | 1.39E-04 | 3.96E-06 |
| Whole River | 2.21E-06 | 1.06E-07 | 4.85E-05 | 1.67E-06 | 7.38E-05 | 1.52E-05 |

Factors obtained by multiplying media-specific TEF in Table 4-2 by individual congener concentrations for each sample, averaging across location and summing

Source: TAMS/Gradient Database Release 4.1b

TABLE 3-3: WHOLE WATER CONCENTRATIONS BASED ON 1993 USEPA PHASE 2 DATASET

| Location | Tri+ PCB | | Avian Based TEF | | Mammalian Based TEF | |
|----------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|-----------------------------------|
| | Average Conc. in Water mg/L | 95% UCL Conc. In Water mg/L | Average Conc. in Water mg/L | 95% UCL Conc. In Water mg/L | Average Conc. in Water mg/L | 95% UCL Conc. In Water mg/L |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 7.36E-05 | 2.33E-04 | 6.01E-07 | 1.90E-06 | 4.66E-07 | 1.47E-06 |
| Stillwater (168) | 1.31E-04 | 4.15E-04 | 1.07E-06 | 3.39E-06 | 8.27E-07 | 2.62E-06 |
| Federal Dam (154) | 9.14E-05 | 1.96E-04 | 7.47E-07 | 1.60E-06 | 5.78E-07 | 1.24E-06 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 7.07E-05 | 7.70E-04 | 6.01E-08 | 6.55E-07 | 4.62E-08 | 5.03E-07 |
| 137.2 | 7.07E-05 | 7.70E-04 | 6.01E-08 | 6.55E-07 | 4.62E-08 | 5.03E-07 |
| 122.4 | 3.24E-05 | 4.15E-04 | 2.76E-08 | 3.53E-07 | 2.11E-08 | 2.71E-07 |
| 113.8 | 3.24E-05 | 4.15E-04 | 2.76E-08 | 3.53E-07 | 2.11E-08 | 2.71E-07 |
| 100 | 3.24E-05 | 4.15E-04 | 2.76E-08 | 3.53E-07 | 2.11E-08 | 2.71E-07 |
| 88.9 | 2.13E-05 | 9.48E-05 | 1.82E-08 | 8.06E-08 | 1.39E-08 | 6.19E-08 |
| 58.7 | 2.13E-05 | 9.48E-05 | 1.82E-08 | 8.06E-08 | 1.39E-08 | 6.19E-08 |
| 47.3 | 2.13E-05 | 9.48E-05 | 1.82E-08 | 8.06E-08 | 1.39E-08 | 6.19E-08 |
| 25.8 | 2.13E-05 | 9.48E-05 | 1.82E-08 | 8.06E-08 | 1.39E-08 | 6.19E-08 |

Notes:

Source: TAMS/Gradient Database Release 4.1b

Water concentrations estimated from Phase 2 dataset -- data averaged across appropriate lower river water column sampling locations

TABLE 3-4: DRY WEIGHT SEDIMENT CONCENTRATIONS BASED ON USEPA PHASE 2 DATASET

| Location | Tri+ PCB | | Avian Based TEF | | Mammalian Based TEF | |
|----------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| | Sediment Conc. mg/Kg |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 11.879 | 17.381 | 3.30E-02 | 4.83E-02 | 9.24E-03 | 1.35E-02 |
| Stillwater (168) | 31.030 | 54.170 | 8.62E-02 | 1.50E-01 | 2.41E-02 | 4.21E-02 |
| Federal Dam (154) | 2.793 | 4.684 | 7.76E-03 | 1.30E-02 | 2.17E-03 | 3.64E-03 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 0.860 | 0.942 | 1.88E-03 | 2.06E-03 | 2.12E-04 | 2.32E-04 |
| 137.2 | 1.519 | 3.069 | 3.32E-03 | 6.71E-03 | 3.74E-04 | 7.56E-04 |
| 122.4 | 0.963 | 1.069 | 2.10E-03 | 2.34E-03 | 2.37E-04 | 2.63E-04 |
| 113.8 | 1.009 | 1.667 | 2.21E-03 | 3.64E-03 | 2.48E-04 | 4.11E-04 |
| 100 | 0.399 | 8.613 | 8.72E-04 | 1.88E-02 | 9.82E-05 | 2.12E-03 |
| 88.9 | 0.781 | 2.284 | 1.71E-03 | 4.99E-03 | 1.92E-04 | 5.62E-04 |
| 58.7 | 0.252 | 2.794 | 5.51E-04 | 6.11E-03 | 6.20E-05 | 6.88E-04 |
| 47.3 | 1.537 | 6.000 | 3.36E-03 | 1.31E-02 | 3.79E-04 | 1.48E-03 |
| 25.8 | 0.578 | 1.563 | 1.26E-03 | 3.42E-03 | 1.42E-04 | 3.85E-04 |

Source: TAMS/Gradient Database Release 4.1b

TABLE 3-5: BENTHIC INVERTEBRATE CONCENTRATIONS BASED ON USEPA PHASE 2 DATASET

| Location | Tri+ PCB | | Avian Based TEF | | Mammalian Based TEF | |
|----------------------------|----------------------|-----------------------|----------------------|-----------------------|----------------------|-----------------------|
| | Average Benthic | 95% UCL | Average Benthic | 95% UCL | Average Benthic | 95% UCL |
| | Invert Conc mg/Kg | Benthic Conc mg/Kg | Invert Conc mg/Kg | Benthic Conc mg/Kg | Invert Conc mg/Kg | Benthic Conc mg/Kg |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 14.138 | 22.210 | 7.53E-04 | 1.18E-03 | 4.20E-04 | 6.59E-04 |
| Stillwater (168) | 26.377 | 45.912 | 1.41E-03 | 2.45E-03 | 7.83E-04 | 1.36E-03 |
| Federal Dam (154) | 6.286 | 10.942 | 3.35E-04 | 5.83E-04 | 1.87E-04 | 3.25E-04 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 0.876 | 1.524 | 1.21E-04 | 2.11E-04 | 9.45E-05 | 1.64E-04 |
| 137.2 | 1.725 | 3.002 | 2.39E-04 | 4.16E-04 | 1.86E-04 | 3.24E-04 |
| 122.4 | 0.804 | 2.021 | 1.11E-04 | 2.80E-04 | 8.68E-05 | 2.18E-04 |
| 113.8 | 0.691 | 1.203 | 9.57E-05 | 1.67E-04 | 7.45E-05 | 1.30E-04 |
| 100 | 0.380 | 2.598 | 5.27E-05 | 3.60E-04 | 4.10E-05 | 2.80E-04 |
| 88.9 | 0.191 | 0.339 | 2.64E-05 | 4.69E-05 | 2.06E-05 | 3.65E-05 |
| 58.7 | 0.491 | 0.854 | 6.80E-05 | 1.18E-04 | 5.29E-05 | 9.21E-05 |
| 47.3 | 0.666 | 4.891 | 9.23E-05 | 6.78E-04 | 7.19E-05 | 5.28E-04 |
| 25.8 | 0.197 | 0.335 | 2.73E-05 | 4.64E-05 | 2.13E-05 | 3.61E-05 |

Source: TAMS/Gradient Database Release 4.1b

TABLE 3-6: FORAGE FISH CONCENTRATIONS BASED ON USEPA PHASE 2 DATASET

| Location | Tri+ PCB | | Avian Based TEF | | Mammalian Based TEF | |
|----------------------------|---------------|---------------|-----------------|---------------|---------------------|---------------|
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 20.919 | 42.716 | 5.37E-03 | 1.10E-02 | 8.97E-04 | 1.83E-03 |
| Stillwater (168) | 7.062 | 10.099 | 1.81E-03 | 2.59E-03 | 3.03E-04 | 4.33E-04 |
| Federal Dam (154) | 1.657 | 2.405 | 4.25E-04 | 6.17E-04 | 7.11E-05 | 1.03E-04 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 1.927 | 2.314 | 3.66E-04 | 4.39E-04 | 2.25E-04 | 2.70E-04 |
| 137.2 | 3.898 | 8.453 | 7.40E-04 | 1.60E-03 | 4.55E-04 | 9.86E-04 |
| 122.4 | 1.488 | 2.407 | 2.82E-04 | 4.57E-04 | 1.74E-04 | 2.81E-04 |
| 113.8 | 1.560 | 1.618 | 2.96E-04 | 3.07E-04 | 1.82E-04 | 1.89E-04 |
| 100 | 0.676 | 1.167 | 1.28E-04 | 2.21E-04 | 7.89E-05 | 1.36E-04 |
| 88.9 | 1.345 | 1.845 | 2.55E-04 | 3.50E-04 | 1.57E-04 | 2.15E-04 |
| 58.7 | 1.465 | 1.661 | 2.78E-04 | 3.15E-04 | 1.71E-04 | 1.94E-04 |
| 47.3 | 1.304 | 1.728 | 2.47E-04 | 3.28E-04 | 1.52E-04 | 2.02E-04 |
| 25.8 | 0.981 | 1.179 | 1.86E-04 | 2.24E-04 | 1.14E-04 | 1.38E-04 |

Source: TAMS/Gradient Database Release 4.1b

**TABLE 3-7: OBSERVED CONCENTRATIONS IN PPM FOR FISH SPECIES
FOR RIVER MILES 113, 152, 168 AND 189 FROM NYSDEC DATASET**

| WET WEIGHT CONCENTRATIONS | | | | | | | | | | | | | | | |
|---------------------------|---------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| Largemouth Bass 113 | | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 11.12 | * | 34.36 | 16.79 | 21.63 | 38.12 | 12.75 | 20.03 | 26.00 | 94.49 | 182.86 | 345.84 | 25.52 | 49.75 | 42.28 |
| 1994 | 15.66 | 36.04 | 52.09 | 13.68 | 27.41 | 31.79 | 8.59 | 20.00 | 27.08 | 45.07 | 68.02 | 96.35 | 26.28 | 34.44 | 103.77 |
| 1995 | 8.17 | 10.72 | 29.76 | 13.11 | 18.07 | 28.67 | 9.01 | 11.53 | 19.17 | 56.13 | 93.61 | 128.03 | 19.69 | 26.06 | 27.05 |
| 1996 | 9.32 | 16.09 | 26.70 | | | | | | | 27.90 | 37.34 | 56.96 | 16.13 | * | 18.81 |

| LIPID NORMALIZED CONCENTRATIONS | | | | | | | | | | | | | | | |
|---------------------------------|---------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| Largemouth Bass 113 | | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 129.94 | 269.40 | 225.31 | 504.50 | 589.56 | 790.13 | 244.29 | * | 377.87 | 2215.67 | 3374.69 | 4993.02 | 942.50 | 1583.04 | 1783.53 |
| 1994 | 150.43 | 225.71 | 435.87 | 480.44 | 892.00 | 1115.37 | 165.28 | 344.97 | 763.25 | 1237.36 | 1558.14 | 2569.30 | 718.47 | 1006.55 | 3088.33 |
| 1995 | 143.41 | 174.17 | 226.37 | 557.58 | 658.90 | 978.39 | 161.85 | 194.03 | 489.69 | 1077.61 | 1661.12 | 1822.26 | 341.76 | 421.82 | 612.17 |
| 1996 | 115.88 | 161.28 | 237.33 | | | | | | | 779.19 | 990.00 | 1411.89 | 356.67 | * | 410.64 |

| WET WEIGHT CONCENTRATIONS | | | | | | | | | | | | | | | |
|---------------------------|---------|---------|---------|--------------------|---------|---------|---------------------|---------|---------|---------------------|---------|---------|---------------------|---------|---------|
| White Perch 113 | | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 2.54 | 5.37 | 5.60 | 2.42 | 4.97 | 3.69 | 1.06 | 1.26 | 3.25 | 10.53 | * | 21.21 | 36.25 | 66.22 | 131.33 |
| 1994 | 1.04 | 1.73 | 2.23 | 4.81 | 9.34 | 8.67 | 0.51 | 0.66 | 0.80 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | 4.94 | * | 8.84 | 2.78 | 5.40 | 8.14 | | | | | | | | | |

| LIPID NORMALIZED CONCENTRATIONS | | | | | | | | | | | | | | | |
|---------------------------------|---------|---------|---------|--------------------|---------|---------|---------------------|---------|---------|---------------------|---------|---------|---------------------|---------|----------|
| White Perch 113 | | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 119.96 | 174.09 | 215.41 | 106.16 | 121.40 | 160.87 | 108.84 | 132.50 | 427.33 | 478.90 | * | 708.98 | 2724.70 | 3980.41 | 13539.37 |
| 1994 | 110.50 | 161.09 | 192.59 | 308.69 | 376.11 | 651.61 | 52.32 | 69.06 | 71.97 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | 150.33 | 389.55 | 346.93 | 92.13 | 105.80 | 175.91 | | | | | | | | | |

* Indicates that the calculated UCL exceeded the maximum due to small sample size.

Source: TAMS/Gradient Database Release 4.1b

**TABLE 3-8
OBSERVED STRIPED BASS CONCENTRATIONS FROM NYSDEC
FOR THE HUDSON RIVER**

| River Mile | Year | Average Wet | 95%UCL | Average Lipid | 95% UCL |
|------------|------|-------------------|---------|-------------------------------------|-----------------------|
| | | Weight (mg/Kg) | (mg/Kg) | Normalized (mg PCB /Kg Lipid) | (mg PCB /Kg Lipid) |
| 12 | 1993 | 1.33 | 2.13 | 25.09 | 38.48 |
| 27 | 1993 | 2.55 | 4.09 | 46.12 | 69.02 |
| 33 | 1993 | 4.25 | 8.46 | 54.81 | 96.75 |
| 40 | 1993 | 1.61 | 2.03 | 32.28 | 43.61 |
| 74 | 1993 | 3.15 | 4.90 | 75.92 | 135.09 |
| 112 | 1993 | 3.74 | 5.63 | 121.90 | 204.94 |
| 152 | 1993 | 12.38 | 17.67 | 306.41 | 533.77 |
| 26 | 1994 | 1.65 | 2.56 | 36.28 | 53.31 |
| 37 | 1994 | 1.95 | 3.03 | 46.46 | 84.32 |
| 40 | 1994 | 1.78 | 2.91 | 40.01 | 58.23 |
| 74 | 1994 | 2.54 | 4.22 | 49.58 | 75.02 |
| 112 | 1994 | 3.05 | 8.35 | 132.69 | 476.15 |
| 152 | 1994 | 6.40 | 9.79 | 254.45 | 338.72 |
| 27 | 1995 | 1.99 | 2.86 | 53.47 | 90.59 |
| 36 | 1995 | 1.30 | 1.70 | 26.03 | 35.97 |
| 59 | 1995 | 2.26 | 2.83 | 165.82 | 250.65 |
| 76 | 1995 | 1.72 | 2.05 | 29.85 | 35.79 |
| 113 | 1995 | 1.69 | 2.68 | 35.14 | 58.11 |
| 152 | 1995 | 6.51 | 8.42 | 174.47 | 235.65 |
| 12 | 1996 | 1.26 | 1.87 | 35.97 | 52.95 |
| 29 | 1996 | 1.92 | 2.74 | 44.94 | 58.94 |
| 40 | 1996 | 1.69 | 2.31 | 47.40 | 71.61 |
| 74 | 1996 | 1.81 | 2.49 | 46.47 | 68.76 |
| 112 | 1996 | 1.90 | 3.11 | 86.86 | 186.24 |
| 152 | 1996 | 4.89 | 11.26 | 216.03 | 531.63 |

**TABLE 3-9: OBSERVED MAMMALIAN AND AVIAN
PCB CONCENTRATIONS**

| Species and Statistic | 1983 - 1986 Mink and Otter Concentrations in mg/kg | | | |
|-----------------------|--|---------------------|---------------|----------------|
| | North Hudson Valley | South Hudson Valley | Hudson Valley | Other NY State |
| Mink liver - average | 0.6 | 0.7 | | |
| Mink liver - minimum | 0.1 | 0.1 | | |
| Mink liver - maximum | 1.7 | 3.4 | | |
| | | | | |
| Otter liver - average | | | 2.3 | |
| Otter liver - minimum | | | 0.7 | |
| Otter liver - maximum | | | 7.3 | |
| | | | | |
| | Tree Swallow Concentrations in mg/kg | | | |
| | Lock 9 | Remnant | SA13 | Saratoga |
| Eggs | 0.852 | 6.55 | 29.6 | 18.5 |
| | 2.57 | 22.9 | 77.3 | 2.37 |
| | 5.72 | 12.9 | 17.6 | 15.7 |
| | 16 | 4.6 | 44 | 13 |
| EGG AVERAGE | 6.28 | 11.7 | 42.1 | 12.4 |
| Nestlings | 0.51 | 31.1 | 54.8 | 9.78 |
| | 0.244 | 27.1 | 56.8 | 0.721 |
| NESTLING AVERAGE | 0.377 | 29.1 | 55.8 | 5.25 |

TABLE 3-10: SUMMARY OF TRI+ WHOLE WATER CONCENTRATIONS FROM THE HUDTOX MODEL AND TEQ-BASED PREDICTIONS FOR 1993 - 2018

| Year | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | | Average Avian TEF | | | 95% Avian TEF | | | Average Mammalian TEF | | | 95% UCL Mammalian TEF | | |
|------|--------------------------|------------------|------------------|----------------------|------------------|------------------|-------------------|------------------|------------------|------------------|------------------|------------------|-----------------------|------------------|------------------|-----------------------|------------------|------------------|
| | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 |
| | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc |
| 1993 | 4.9E-05 | 7.3E-05 | 4.9E-05 | 4.9E-05 | 7.5E-05 | 1.1E-05 | 4.0E-07 | 6.0E-07 | 4.0E-07 | 4.0E-07 | 6.1E-07 | 9.1E-08 | 3.1E-07 | 4.6E-07 | 3.1E-07 | 3.1E-07 | 4.7E-07 | 7.1E-08 |
| 1994 | 4.1E-05 | 5.7E-05 | 3.8E-05 | 4.2E-05 | 5.9E-05 | 8.5E-06 | 3.4E-07 | 4.6E-07 | 3.1E-07 | 3.4E-07 | 4.8E-07 | 6.9E-08 | 2.6E-07 | 3.6E-07 | 2.4E-07 | 2.6E-07 | 3.7E-07 | 5.4E-08 |
| 1995 | 6.6E-05 | 9.6E-05 | 6.0E-05 | 6.6E-05 | 9.9E-05 | 1.4E-05 | 5.3E-07 | 7.9E-07 | 4.9E-07 | 5.4E-07 | 8.1E-07 | 1.2E-07 | 4.1E-07 | 6.1E-07 | 3.8E-07 | 4.2E-07 | 6.2E-07 | 8.9E-08 |
| 1996 | 2.6E-05 | 4.0E-05 | 2.6E-05 | 2.6E-05 | 4.4E-05 | 7.0E-06 | 2.1E-07 | 3.3E-07 | 2.1E-07 | 2.1E-07 | 3.6E-07 | 5.7E-08 | 1.6E-07 | 2.5E-07 | 1.6E-07 | 1.7E-07 | 2.8E-07 | 4.4E-08 |
| 1997 | 2.8E-05 | 4.8E-05 | 3.2E-05 | 2.8E-05 | 4.9E-05 | 9.5E-06 | 2.2E-07 | 4.0E-07 | 2.6E-07 | 2.3E-07 | 4.0E-07 | 7.7E-08 | 1.7E-07 | 3.1E-07 | 2.0E-07 | 1.8E-07 | 3.1E-07 | 6.0E-08 |
| 1998 | 2.9E-05 | 5.0E-05 | 2.7E-05 | 2.9E-05 | 5.0E-05 | 2.7E-05 | 2.4E-07 | 4.0E-07 | 2.2E-07 | 2.4E-07 | 4.1E-07 | 2.2E-07 | 1.8E-07 | 3.1E-07 | 1.7E-07 | 1.9E-07 | 3.2E-07 | 1.7E-07 |
| 1999 | 2.4E-05 | 4.3E-05 | 2.5E-05 | 2.4E-05 | 4.4E-05 | 2.7E-05 | 2.0E-07 | 3.5E-07 | 2.0E-07 | 2.0E-07 | 3.6E-07 | 2.2E-07 | 1.5E-07 | 2.7E-07 | 1.6E-07 | 1.5E-07 | 2.8E-07 | 1.7E-07 |
| 2000 | 2.0E-05 | 3.6E-05 | 2.2E-05 | 2.1E-05 | 3.8E-05 | 2.4E-05 | 1.6E-07 | 3.0E-07 | 1.8E-07 | 1.7E-07 | 3.1E-07 | 2.0E-07 | 1.3E-07 | 2.3E-07 | 1.4E-07 | 1.3E-07 | 2.4E-07 | 1.5E-07 |
| 2001 | 2.6E-05 | 4.2E-05 | 2.6E-05 | 2.7E-05 | 4.3E-05 | 2.6E-05 | 2.1E-07 | 3.5E-07 | 2.1E-07 | 2.2E-07 | 3.5E-07 | 2.2E-07 | 1.6E-07 | 2.7E-07 | 1.6E-07 | 1.7E-07 | 2.7E-07 | 1.7E-07 |
| 2002 | 2.2E-05 | 3.9E-05 | 2.2E-05 | 2.2E-05 | 4.1E-05 | 2.4E-05 | 1.8E-07 | 3.2E-07 | 1.8E-07 | 1.8E-07 | 3.3E-07 | 1.9E-07 | 1.4E-07 | 2.5E-07 | 1.4E-07 | 1.4E-07 | 2.6E-07 | 1.5E-07 |
| 2003 | 1.7E-05 | 3.1E-05 | 1.9E-05 | 1.7E-05 | 3.2E-05 | 2.1E-05 | 1.4E-07 | 2.5E-07 | 1.5E-07 | 1.4E-07 | 2.6E-07 | 1.7E-07 | 1.1E-07 | 2.0E-07 | 1.2E-07 | 1.1E-07 | 2.0E-07 | 1.3E-07 |
| 2004 | 1.6E-05 | 2.9E-05 | 1.9E-05 | 1.6E-05 | 3.1E-05 | 2.2E-05 | 1.3E-07 | 2.4E-07 | 1.6E-07 | 1.3E-07 | 2.5E-07 | 1.8E-07 | 9.8E-08 | 1.9E-07 | 1.2E-07 | 1.0E-07 | 2.0E-07 | 1.4E-07 |
| 2005 | 1.9E-05 | 3.4E-05 | 2.2E-05 | 1.9E-05 | 3.5E-05 | 2.3E-05 | 1.5E-07 | 2.8E-07 | 1.8E-07 | 1.6E-07 | 2.9E-07 | 1.8E-07 | 1.2E-07 | 2.2E-07 | 1.4E-07 | 1.2E-07 | 2.2E-07 | 1.4E-07 |
| 2006 | 1.9E-05 | 3.4E-05 | 2.2E-05 | 1.9E-05 | 3.5E-05 | 2.3E-05 | 1.5E-07 | 2.8E-07 | 1.8E-07 | 1.6E-07 | 2.9E-07 | 1.8E-07 | 1.2E-07 | 2.2E-07 | 1.4E-07 | 1.2E-07 | 2.2E-07 | 1.4E-07 |
| 2007 | 1.0E-05 | 1.9E-05 | 1.0E-05 | 1.0E-05 | 1.9E-05 | 1.1E-05 | 8.3E-08 | 1.5E-07 | 8.3E-08 | 8.5E-08 | 1.6E-07 | 9.0E-08 | 6.4E-08 | 1.2E-07 | 6.5E-08 | 6.5E-08 | 1.2E-07 | 7.0E-08 |
| 2008 | 1.4E-05 | 2.4E-05 | 1.5E-05 | 1.4E-05 | 2.6E-05 | 1.7E-05 | 1.1E-07 | 2.0E-07 | 1.2E-07 | 1.1E-07 | 2.1E-07 | 1.4E-07 | 8.7E-08 | 1.5E-07 | 9.7E-08 | 8.9E-08 | 1.6E-07 | 1.1E-07 |
| 2009 | 1.6E-05 | 2.8E-05 | 1.6E-05 | 1.7E-05 | 2.9E-05 | 1.8E-05 | 1.3E-07 | 2.3E-07 | 1.3E-07 | 1.4E-07 | 2.4E-07 | 1.4E-07 | 1.0E-07 | 1.8E-07 | 1.0E-07 | 1.1E-07 | 1.8E-07 | 1.1E-07 |
| 2010 | 9.6E-06 | 1.8E-05 | 1.0E-05 | 9.8E-06 | 1.9E-05 | 1.1E-05 | 7.8E-08 | 1.5E-07 | 8.3E-08 | 8.0E-08 | 1.5E-07 | 9.2E-08 | 6.1E-08 | 1.1E-07 | 6.4E-08 | 6.2E-08 | 1.2E-07 | 7.1E-08 |
| 2011 | 9.4E-06 | 1.8E-05 | 1.1E-05 | 9.6E-06 | 2.0E-05 | 1.3E-05 | 7.7E-08 | 1.5E-07 | 9.0E-08 | 7.9E-08 | 1.6E-07 | 1.0E-07 | 5.9E-08 | 1.2E-07 | 7.0E-08 | 6.1E-08 | 1.3E-07 | 8.1E-08 |
| 2012 | 1.2E-05 | 2.3E-05 | 1.4E-05 | 1.2E-05 | 2.4E-05 | 1.5E-05 | 1.0E-07 | 1.9E-07 | 1.2E-07 | 1.0E-07 | 1.9E-07 | 1.2E-07 | 7.7E-08 | 1.5E-07 | 9.0E-08 | 7.8E-08 | 1.5E-07 | 9.4E-08 |
| 2013 | 8.8E-06 | 1.6E-05 | 8.7E-06 | 8.9E-06 | 1.7E-05 | 9.5E-06 | 7.2E-08 | 1.3E-07 | 7.1E-08 | 7.3E-08 | 1.4E-07 | 7.7E-08 | 5.6E-08 | 1.0E-07 | 5.5E-08 | 5.7E-08 | 1.1E-07 | 6.0E-08 |
| 2014 | 7.4E-06 | 1.9E-05 | 1.1E-05 | 7.5E-06 | 1.9E-05 | 1.1E-05 | 6.0E-08 | 1.5E-07 | 8.8E-08 | 6.1E-08 | 1.5E-07 | 9.1E-08 | 4.7E-08 | 1.2E-07 | 6.8E-08 | 4.7E-08 | 1.2E-07 | 7.1E-08 |
| 2015 | 7.2E-06 | 1.4E-05 | 7.9E-06 | 7.3E-06 | 1.4E-05 | 8.5E-06 | 5.8E-08 | 1.1E-07 | 6.4E-08 | 6.0E-08 | 1.2E-07 | 6.9E-08 | 4.5E-08 | 8.6E-08 | 5.0E-08 | 4.6E-08 | 8.9E-08 | 5.4E-08 |
| 2016 | 1.2E-05 | 2.4E-05 | 1.4E-05 | 1.2E-05 | 2.4E-05 | 1.4E-05 | 9.9E-08 | 1.9E-07 | 1.1E-07 | 1.0E-07 | 2.0E-07 | 1.2E-07 | 7.6E-08 | 1.5E-07 | 8.6E-08 | 7.8E-08 | 1.5E-07 | 8.9E-08 |
| 2017 | 5.5E-06 | 1.1E-05 | 6.2E-06 | 5.6E-06 | 1.2E-05 | 7.0E-06 | 4.5E-08 | 8.8E-08 | 5.1E-08 | 4.6E-08 | 9.6E-08 | 5.7E-08 | 3.5E-08 | 6.8E-08 | 3.9E-08 | 3.6E-08 | 7.4E-08 | 4.4E-08 |
| 2018 | 6.6E-06 | 1.4E-05 | 8.8E-06 | 6.8E-06 | 1.4E-05 | 9.5E-06 | 5.4E-08 | 1.1E-07 | 7.2E-08 | 5.5E-08 | 1.2E-07 | 7.7E-08 | 4.2E-08 | 8.7E-08 | 5.6E-08 | 4.3E-08 | 9.0E-08 | 6.0E-08 |

TABLE 3-11: SUMMARY OF TRI+ SEDIMENT CONCENTRATIONS FROM THE HUDTOX MODEL AND TEQ-BASED PREDICTIONS FOR 1993 - 2018

| Year | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | | Average Avian TEF | | | 95% Avian TEF | | | Average Mammalian TEF | | | 95% UCL Mammalian TEF | | |
|------|--------------------------|-----------|-----------|----------------------|-----------|-----------|-------------------|-----------|-----------|---------------|-----------|-----------|-----------------------|-----------|-----------|-----------------------|-----------|-----------|
| | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total |
| | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc |
| | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg |
| 1993 | 28.81 | 9.28 | 4.13 | 30.40 | 9.33 | 4.16 | 8.0E-02 | 2.6E-02 | 1.1E-02 | 8.4E-02 | 2.6E-02 | 1.2E-02 | 2.2E-02 | 7.2E-03 | 3.2E-03 | 2.4E-02 | 7.3E-03 | 3.2E-03 |
| 1994 | 26.69 | 8.74 | 3.79 | 28.16 | 8.75 | 3.80 | 7.4E-02 | 2.4E-02 | 1.1E-02 | 7.8E-02 | 2.4E-02 | 1.1E-02 | 2.1E-02 | 6.8E-03 | 3.0E-03 | 2.2E-02 | 6.8E-03 | 3.0E-03 |
| 1995 | 25.12 | 8.48 | 3.70 | 26.49 | 8.48 | 3.71 | 7.0E-02 | 2.4E-02 | 1.0E-02 | 7.4E-02 | 2.4E-02 | 1.0E-02 | 2.0E-02 | 6.6E-03 | 2.9E-03 | 2.1E-02 | 6.6E-03 | 2.9E-03 |
| 1996 | 22.81 | 7.61 | 3.11 | 24.04 | 7.64 | 3.13 | 6.3E-02 | 2.1E-02 | 8.6E-03 | 6.7E-02 | 2.1E-02 | 8.7E-03 | 1.8E-02 | 5.9E-03 | 2.4E-03 | 1.9E-02 | 5.9E-03 | 2.4E-03 |
| 1997 | 21.07 | 6.90 | 2.71 | 22.21 | 6.92 | 2.72 | 5.9E-02 | 1.9E-02 | 7.5E-03 | 6.2E-02 | 1.9E-02 | 7.5E-03 | 1.6E-02 | 5.4E-03 | 2.1E-03 | 1.7E-02 | 5.4E-03 | 2.1E-03 |
| 1998 | 19.01 | 6.19 | 2.33 | 20.04 | 6.23 | 2.36 | 5.3E-02 | 1.7E-02 | 6.5E-03 | 5.6E-02 | 1.7E-02 | 6.5E-03 | 1.5E-02 | 4.8E-03 | 1.8E-03 | 1.6E-02 | 4.8E-03 | 1.8E-03 |
| 1999 | 17.34 | 5.66 | 2.06 | 18.28 | 5.67 | 2.06 | 4.8E-02 | 1.6E-02 | 5.7E-03 | 5.1E-02 | 1.6E-02 | 5.7E-03 | 1.3E-02 | 4.4E-03 | 1.6E-03 | 1.4E-02 | 4.4E-03 | 1.6E-03 |
| 2000 | 15.83 | 5.17 | 1.78 | 16.69 | 5.19 | 1.80 | 4.4E-02 | 1.4E-02 | 5.0E-03 | 4.6E-02 | 1.4E-02 | 5.0E-03 | 1.2E-02 | 4.0E-03 | 1.4E-03 | 1.3E-02 | 4.0E-03 | 1.4E-03 |
| 2001 | 14.97 | 5.06 | 1.66 | 15.79 | 5.07 | 1.67 | 4.2E-02 | 1.4E-02 | 4.6E-03 | 4.4E-02 | 1.4E-02 | 4.6E-03 | 1.2E-02 | 3.9E-03 | 1.3E-03 | 1.2E-02 | 3.9E-03 | 1.3E-03 |
| 2002 | 14.16 | 4.88 | 1.59 | 14.94 | 4.88 | 1.60 | 3.9E-02 | 1.4E-02 | 4.4E-03 | 4.1E-02 | 1.4E-02 | 4.4E-03 | 1.1E-02 | 3.8E-03 | 1.2E-03 | 1.2E-02 | 3.8E-03 | 1.2E-03 |
| 2003 | 13.19 | 4.65 | 1.51 | 13.90 | 4.66 | 1.51 | 3.7E-02 | 1.3E-02 | 4.2E-03 | 3.9E-02 | 1.3E-02 | 4.2E-03 | 1.0E-02 | 3.6E-03 | 1.2E-03 | 1.1E-02 | 3.6E-03 | 1.2E-03 |
| 2004 | 12.08 | 4.33 | 1.39 | 12.74 | 4.34 | 1.39 | 3.4E-02 | 1.2E-02 | 3.9E-03 | 3.5E-02 | 1.2E-02 | 3.9E-03 | 9.4E-03 | 3.4E-03 | 1.1E-03 | 9.9E-03 | 3.4E-03 | 1.1E-03 |
| 2005 | 11.13 | 3.88 | 1.22 | 11.74 | 3.89 | 1.23 | 3.1E-02 | 1.1E-02 | 3.4E-03 | 3.3E-02 | 1.1E-02 | 3.4E-03 | 8.7E-03 | 3.0E-03 | 9.5E-04 | 9.1E-03 | 3.0E-03 | 9.5E-04 |
| 2006 | 10.71 | 3.74 | 1.16 | 11.30 | 3.75 | 1.16 | 3.0E-02 | 1.0E-02 | 3.2E-03 | 3.1E-02 | 1.0E-02 | 3.2E-03 | 8.3E-03 | 2.9E-03 | 9.0E-04 | 8.8E-03 | 2.9E-03 | 9.0E-04 |
| 2007 | 10.06 | 3.47 | 1.04 | 10.60 | 3.49 | 1.05 | 2.8E-02 | 9.7E-03 | 2.9E-03 | 2.9E-02 | 9.7E-03 | 2.9E-03 | 7.8E-03 | 2.7E-03 | 8.1E-04 | 8.2E-03 | 2.7E-03 | 8.1E-04 |
| 2008 | 9.37 | 3.25 | 0.95 | 9.87 | 3.26 | 0.96 | 2.6E-02 | 9.0E-03 | 2.6E-03 | 2.7E-02 | 9.1E-03 | 2.7E-03 | 7.3E-03 | 2.5E-03 | 7.4E-04 | 7.7E-03 | 2.5E-03 | 7.4E-04 |
| 2009 | 8.92 | 3.10 | 0.91 | 9.40 | 3.10 | 0.98 | 2.5E-02 | 8.6E-03 | 2.5E-03 | 2.6E-02 | 8.6E-03 | 2.7E-03 | 6.9E-03 | 2.4E-03 | 7.1E-04 | 7.3E-03 | 2.4E-03 | 7.7E-04 |
| 2010 | 8.35 | 2.97 | 0.88 | 8.80 | 2.98 | 0.89 | 2.3E-02 | 8.3E-03 | 2.5E-03 | 2.4E-02 | 8.3E-03 | 2.5E-03 | 6.5E-03 | 2.3E-03 | 6.9E-04 | 6.8E-03 | 2.3E-03 | 6.9E-04 |
| 2011 | 7.38 | 2.68 | 0.80 | 7.78 | 2.70 | 0.80 | 2.1E-02 | 7.5E-03 | 2.2E-03 | 2.2E-02 | 7.5E-03 | 2.2E-03 | 5.7E-03 | 2.1E-03 | 6.2E-04 | 6.0E-03 | 2.1E-03 | 6.2E-04 |
| 2012 | 6.79 | 2.53 | 0.74 | 7.16 | 2.53 | 0.74 | 1.9E-02 | 7.0E-03 | 2.1E-03 | 2.0E-02 | 7.0E-03 | 2.1E-03 | 5.3E-03 | 2.0E-03 | 5.8E-04 | 5.6E-03 | 2.0E-03 | 5.8E-04 |
| 2013 | 6.43 | 2.47 | 0.73 | 6.79 | 2.47 | 0.73 | 1.8E-02 | 6.9E-03 | 2.0E-03 | 1.9E-02 | 6.9E-03 | 2.0E-03 | 5.0E-03 | 1.9E-03 | 5.7E-04 | 5.3E-03 | 1.9E-03 | 5.7E-04 |
| 2014 | 5.93 | 2.27 | 0.64 | 6.27 | 2.27 | 0.65 | 1.6E-02 | 6.3E-03 | 1.8E-03 | 1.7E-02 | 6.3E-03 | 1.8E-03 | 4.6E-03 | 1.8E-03 | 5.0E-04 | 4.9E-03 | 1.8E-03 | 5.0E-04 |
| 2015 | 5.57 | 2.17 | 0.61 | 5.89 | 2.18 | 0.61 | 1.5E-02 | 6.0E-03 | 1.7E-03 | 1.6E-02 | 6.0E-03 | 1.7E-03 | 4.3E-03 | 1.7E-03 | 4.7E-04 | 4.6E-03 | 1.7E-03 | 4.7E-04 |
| 2016 | 5.31 | 2.12 | 0.60 | 5.62 | 2.12 | 0.60 | 1.5E-02 | 5.9E-03 | 1.7E-03 | 1.6E-02 | 5.9E-03 | 1.7E-03 | 4.1E-03 | 1.7E-03 | 4.6E-04 | 4.4E-03 | 1.7E-03 | 4.7E-04 |
| 2017 | 4.87 | 1.94 | 0.51 | 5.15 | 1.95 | 0.51 | 1.4E-02 | 5.4E-03 | 1.4E-03 | 1.4E-02 | 5.4E-03 | 1.4E-03 | 3.8E-03 | 1.5E-03 | 3.9E-04 | 4.0E-03 | 1.5E-03 | 4.0E-04 |
| 2018 | 4.59 | 1.83 | 0.47 | 4.85 | 1.83 | 0.47 | 1.3E-02 | 5.1E-03 | 1.3E-03 | 1.3E-02 | 5.1E-03 | 1.3E-03 | 3.6E-03 | 1.4E-03 | 3.6E-04 | 3.8E-03 | 1.4E-03 | 3.6E-04 |

TABLE 3-12: SUMMARY OF TRI+ BENTHIC INVERTEBRATE CONCENTRATIONS FROM THE FISHRAND MODEL AND TEQ-BASED PREDICTIONS FOR 1993 - 2018

| Year | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | | Average Avian TEF | | | 95% Avian TEF | | | Average Mammalian TEF | | | 95% UCL Mammalian TEF | | |
|------|--------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|
| | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc | 189 Total Benthic Conc | 168 Total Benthic Conc | 154 Total Benthic Conc |
| | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg | mg/kg |
| 1993 | 14.822 | 8.209 | 3.610 | 21.698 | 12.953 | 3.878 | 7.9E-04 | 4.4E-04 | 1.9E-04 | 1.2E-03 | 6.9E-04 | 2.1E-04 | 4.4E-04 | 2.4E-04 | 1.1E-04 | 6.4E-04 | 3.8E-04 | 1.2E-04 |
| 1994 | 13.910 | 7.812 | 3.405 | 20.580 | 12.409 | 3.682 | 7.4E-04 | 4.2E-04 | 1.8E-04 | 1.1E-03 | 6.6E-04 | 2.0E-04 | 4.1E-04 | 2.3E-04 | 1.0E-04 | 6.1E-04 | 3.7E-04 | 1.1E-04 |
| 1995 | 13.004 | 7.352 | 3.124 | 18.806 | 11.637 | 3.363 | 6.9E-04 | 3.9E-04 | 1.7E-04 | 1.0E-03 | 6.2E-04 | 1.8E-04 | 3.9E-04 | 2.2E-04 | 9.3E-05 | 5.6E-04 | 3.5E-04 | 1.0E-04 |
| 1996 | 11.722 | 6.562 | 2.644 | 17.161 | 10.360 | 2.867 | 6.2E-04 | 3.5E-04 | 1.4E-04 | 9.1E-04 | 5.5E-04 | 1.5E-04 | 3.5E-04 | 1.9E-04 | 7.8E-05 | 5.1E-04 | 3.1E-04 | 8.5E-05 |
| 1997 | 10.776 | 5.987 | 2.304 | 15.698 | 9.505 | 2.471 | 5.7E-04 | 3.2E-04 | 1.2E-04 | 8.4E-04 | 5.1E-04 | 1.3E-04 | 3.2E-04 | 1.8E-04 | 6.8E-05 | 4.7E-04 | 2.8E-04 | 7.3E-05 |
| 1998 | 9.882 | 5.399 | 2.006 | 14.126 | 8.557 | 2.157 | 5.3E-04 | 2.9E-04 | 1.1E-04 | 7.5E-04 | 4.6E-04 | 1.1E-04 | 2.9E-04 | 1.6E-04 | 6.0E-05 | 4.2E-04 | 2.5E-04 | 6.4E-05 |
| 1999 | 8.931 | 4.912 | 1.750 | 13.237 | 7.817 | 1.889 | 4.8E-04 | 2.6E-04 | 9.3E-05 | 7.1E-04 | 4.2E-04 | 1.0E-04 | 2.7E-04 | 1.5E-04 | 5.2E-05 | 3.9E-04 | 2.3E-04 | 5.6E-05 |
| 2000 | 8.214 | 4.633 | 1.573 | 11.823 | 7.325 | 1.690 | 4.4E-04 | 2.5E-04 | 8.4E-05 | 6.3E-04 | 3.9E-04 | 9.0E-05 | 2.4E-04 | 1.4E-04 | 4.7E-05 | 3.5E-04 | 2.2E-04 | 5.0E-05 |
| 2001 | 7.757 | 4.521 | 1.482 | 11.289 | 7.181 | 1.599 | 4.1E-04 | 2.4E-04 | 7.9E-05 | 6.0E-04 | 3.8E-04 | 8.5E-05 | 2.3E-04 | 1.3E-04 | 4.4E-05 | 3.3E-04 | 2.1E-04 | 4.7E-05 |
| 2002 | 7.249 | 4.315 | 1.405 | 10.752 | 6.863 | 1.521 | 3.9E-04 | 2.3E-04 | 7.5E-05 | 5.7E-04 | 3.7E-04 | 8.1E-05 | 2.2E-04 | 1.3E-04 | 4.2E-05 | 3.2E-04 | 2.0E-04 | 4.5E-05 |
| 2003 | 6.767 | 4.074 | 1.314 | 9.888 | 6.469 | 1.419 | 3.6E-04 | 2.2E-04 | 7.0E-05 | 5.3E-04 | 3.4E-04 | 7.6E-05 | 2.0E-04 | 1.2E-04 | 3.9E-05 | 2.9E-04 | 1.9E-04 | 4.2E-05 |
| 2004 | 6.287 | 3.741 | 1.189 | 9.075 | 5.911 | 1.277 | 3.3E-04 | 2.0E-04 | 6.3E-05 | 4.8E-04 | 3.1E-04 | 6.8E-05 | 1.9E-04 | 1.1E-04 | 3.5E-05 | 2.7E-04 | 1.8E-04 | 3.8E-05 |
| 2005 | 5.805 | 3.465 | 1.082 | 8.509 | 5.494 | 1.166 | 3.1E-04 | 1.8E-04 | 5.8E-05 | 4.5E-04 | 2.9E-04 | 6.2E-05 | 1.7E-04 | 1.0E-04 | 3.2E-05 | 2.5E-04 | 1.6E-04 | 3.5E-05 |
| 2006 | 5.582 | 3.280 | 1.001 | 8.126 | 5.215 | 1.082 | 3.0E-04 | 1.7E-04 | 5.3E-05 | 4.3E-04 | 2.8E-04 | 5.8E-05 | 1.7E-04 | 9.7E-05 | 3.0E-05 | 2.4E-04 | 1.5E-04 | 3.2E-05 |
| 2007 | 5.200 | 3.078 | 0.913 | 7.514 | 4.881 | 0.983 | 2.8E-04 | 1.6E-04 | 4.9E-05 | 4.0E-04 | 2.6E-04 | 5.2E-05 | 1.5E-04 | 9.1E-05 | 2.7E-05 | 2.2E-04 | 1.4E-04 | 2.9E-05 |
| 2008 | 4.947 | 2.879 | 0.845 | 7.086 | 4.567 | 0.916 | 2.6E-04 | 1.5E-04 | 4.5E-05 | 3.8E-04 | 2.4E-04 | 4.9E-05 | 1.5E-04 | 8.5E-05 | 2.5E-05 | 2.1E-04 | 1.4E-04 | 2.7E-05 |
| 2009 | 4.665 | 2.761 | 0.817 | 6.683 | 4.376 | 0.883 | 2.5E-04 | 1.5E-04 | 4.4E-05 | 3.6E-04 | 2.3E-04 | 4.7E-05 | 1.4E-04 | 8.2E-05 | 2.4E-05 | 2.0E-04 | 1.3E-04 | 2.6E-05 |
| 2010 | 4.224 | 2.575 | 0.766 | 6.202 | 4.101 | 0.832 | 2.3E-04 | 1.4E-04 | 4.1E-05 | 3.3E-04 | 2.2E-04 | 4.4E-05 | 1.3E-04 | 7.6E-05 | 2.3E-05 | 1.8E-04 | 1.2E-04 | 2.5E-05 |
| 2011 | 3.828 | 2.375 | 0.702 | 5.548 | 3.792 | 0.759 | 2.0E-04 | 1.3E-04 | 3.7E-05 | 3.0E-04 | 2.0E-04 | 4.0E-05 | 1.1E-04 | 7.0E-05 | 2.1E-05 | 1.6E-04 | 1.1E-04 | 2.3E-05 |
| 2012 | 3.564 | 2.266 | 0.666 | 5.181 | 3.591 | 0.719 | 1.9E-04 | 1.2E-04 | 3.5E-05 | 2.8E-04 | 1.9E-04 | 3.8E-05 | 1.1E-04 | 6.7E-05 | 2.0E-05 | 1.5E-04 | 1.1E-04 | 2.1E-05 |
| 2013 | 3.299 | 2.157 | 0.625 | 4.927 | 3.416 | 0.673 | 1.8E-04 | 1.1E-04 | 3.3E-05 | 2.6E-04 | 1.8E-04 | 3.6E-05 | 9.8E-05 | 6.4E-05 | 1.9E-05 | 1.5E-04 | 1.0E-04 | 2.0E-05 |
| 2014 | 3.078 | 2.020 | 0.572 | 4.534 | 3.219 | 0.616 | 1.6E-04 | 1.1E-04 | 3.0E-05 | 2.4E-04 | 1.7E-04 | 3.3E-05 | 9.1E-05 | 6.0E-05 | 1.7E-05 | 1.3E-04 | 9.6E-05 | 1.8E-05 |
| 2015 | 2.876 | 1.951 | 0.547 | 4.287 | 3.103 | 0.590 | 1.5E-04 | 1.0E-04 | 2.9E-05 | 2.3E-04 | 1.7E-04 | 3.1E-05 | 8.5E-05 | 5.8E-05 | 1.6E-05 | 1.3E-04 | 9.2E-05 | 1.8E-05 |
| 2016 | 2.730 | 1.853 | 0.503 | 3.942 | 2.944 | 0.544 | 1.5E-04 | 9.9E-05 | 2.7E-05 | 2.1E-04 | 1.6E-04 | 2.9E-05 | 8.1E-05 | 5.5E-05 | 1.5E-05 | 1.2E-04 | 8.7E-05 | 1.6E-05 |
| 2017 | 2.525 | 1.735 | 0.449 | 3.746 | 2.753 | 0.487 | 1.3E-04 | 9.2E-05 | 2.4E-05 | 2.0E-04 | 1.5E-04 | 2.6E-05 | 7.5E-05 | 5.1E-05 | 1.3E-05 | 1.1E-04 | 8.2E-05 | 1.4E-05 |
| 2018 | 2.501 | 1.693 | 0.436 | 3.728 | 2.686 | 0.470 | 1.3E-04 | 9.0E-05 | 2.3E-05 | 2.0E-04 | 1.4E-04 | 2.5E-05 | 7.4E-05 | 5.0E-05 | 1.3E-05 | 1.1E-04 | 8.0E-05 | 1.4E-05 |

TABLE 3-13: LARGEMOUTH BASS PREDICTED TRI+ CONCENTRATIONS FOR 1993 - 2018

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------------------|------------------------------------|--|----------------------------------|------------------------------------|--|----------------------------------|------------------------------------|--|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 21.01 | 43.93 | 81.72 | 16.80 | 33.96 | 58.88 | 7.87 | 16.39 | 29.91 |
| 1994 | 11.76 | 25.15 | 58.63 | 11.99 | 23.60 | 40.09 | 5.70 | 11.74 | 20.66 |
| 1995 | 13.16 | 27.86 | 64.51 | 13.37 | 26.25 | 44.38 | 6.41 | 12.83 | 22.21 |
| 1996 | 10.86 | 23.82 | 66.18 | 10.22 | 20.83 | 35.65 | 5.14 | 10.91 | 20.35 |
| 1997 | 8.05 | 18.55 | 48.77 | 9.32 | 18.43 | 31.17 | 4.48 | 9.45 | 17.26 |
| 1998 | 7.17 | 15.67 | 42.07 | 8.14 | 16.45 | 28.09 | 4.00 | 8.10 | 14.41 |
| 1999 | 5.70 | 13.14 | 37.31 | 6.40 | 13.23 | 23.55 | 2.86 | 6.02 | 11.34 |
| 2000 | 5.35 | 12.26 | 35.22 | 6.22 | 12.62 | 21.92 | 2.86 | 5.85 | 10.51 |
| 2001 | 4.97 | 11.51 | 32.22 | 5.95 | 12.11 | 21.39 | 2.68 | 5.59 | 10.05 |
| 2002 | 4.99 | 11.39 | 32.09 | 6.25 | 12.92 | 22.01 | 2.76 | 5.90 | 10.58 |
| 2003 | 4.30 | 9.84 | 28.50 | 5.17 | 10.78 | 18.82 | 2.33 | 5.01 | 8.98 |
| 2004 | 4.07 | 9.32 | 26.13 | 5.14 | 10.57 | 18.24 | 2.43 | 5.19 | 9.33 |
| 2005 | 3.49 | 8.10 | 23.80 | 4.02 | 8.10 | 14.19 | 1.85 | 3.92 | 7.07 |
| 2006 | 3.64 | 8.10 | 23.84 | 4.40 | 8.81 | 15.77 | 2.08 | 4.20 | 7.52 |
| 2007 | 3.21 | 7.34 | 22.00 | 3.77 | 7.41 | 13.23 | 1.62 | 3.25 | 5.90 |
| 2008 | 2.90 | 6.87 | 20.32 | 3.28 | 6.61 | 12.04 | 1.45 | 3.07 | 5.81 |
| 2009 | 3.09 | 7.20 | 19.44 | 3.59 | 7.19 | 12.75 | 1.58 | 3.32 | 6.25 |
| 2010 | 2.84 | 6.33 | 17.95 | 3.19 | 6.53 | 11.42 | 1.38 | 2.81 | 5.17 |
| 2011 | 2.27 | 5.36 | 15.79 | 2.73 | 5.57 | 9.79 | 1.24 | 2.67 | 4.98 |
| 2012 | 2.29 | 5.45 | 14.83 | 2.73 | 5.63 | 9.88 | 1.25 | 2.61 | 5.00 |
| 2013 | 2.24 | 5.23 | 14.56 | 2.82 | 5.85 | 10.13 | 1.19 | 2.51 | 4.80 |
| 2014 | 2.11 | 4.84 | 13.04 | 2.76 | 5.59 | 9.86 | 1.14 | 2.34 | 4.25 |
| 2015 | 1.90 | 4.39 | 12.12 | 2.45 | 4.93 | 8.58 | 0.99 | 2.13 | 3.88 |
| 2016 | 2.00 | 4.54 | 12.27 | 2.73 | 5.52 | 9.68 | 1.13 | 2.33 | 4.19 |
| 2017 | 1.61 | 3.75 | 11.17 | 2.05 | 4.17 | 7.42 | 0.82 | 1.81 | 3.76 |
| 2018 | 1.63 | 3.68 | 10.94 | 1.97 | 3.99 | 7.21 | 0.79 | 1.68 | 3.34 |

TABLE 3-14: BROWN BULLHEAD PREDICTED TRI+ CONCENTRATIONS FOR 1993 - 2018

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 9.80 | 25.30 | 82.32 | 6.83 | 14.97 | 33.37 | 3.20 | 6.87 | 15.14 |
| 1994 | 7.81 | 22.41 | 76.09 | 5.76 | 13.16 | 30.81 | 2.64 | 5.95 | 13.55 |
| 1995 | 8.13 | 22.69 | 73.63 | 6.22 | 13.99 | 31.16 | 2.91 | 6.42 | 13.45 |
| 1996 | 7.03 | 20.65 | 72.72 | 4.97 | 11.61 | 27.65 | 2.24 | 5.03 | 11.74 |
| 1997 | 6.06 | 18.30 | 61.04 | 4.69 | 10.65 | 24.92 | 2.08 | 4.56 | 9.89 |
| 1998 | 5.38 | 16.19 | 56.16 | 3.96 | 9.30 | 22.66 | 1.70 | 3.77 | 8.65 |
| 1999 | 4.66 | 14.62 | 51.42 | 3.48 | 8.31 | 20.35 | 1.38 | 3.20 | 7.49 |
| 2000 | 4.47 | 13.51 | 46.82 | 3.39 | 7.83 | 19.06 | 1.32 | 2.92 | 6.68 |
| 2001 | 4.29 | 12.74 | 44.27 | 3.21 | 7.70 | 18.90 | 1.23 | 2.79 | 6.41 |
| 2002 | 4.06 | 11.87 | 42.79 | 3.26 | 7.47 | 17.84 | 1.23 | 2.69 | 5.97 |
| 2003 | 3.68 | 11.18 | 38.83 | 2.85 | 6.78 | 16.88 | 1.07 | 2.46 | 5.65 |
| 2004 | 3.44 | 10.46 | 35.04 | 2.75 | 6.45 | 15.63 | 1.07 | 2.39 | 5.31 |
| 2005 | 3.08 | 9.56 | 31.95 | 2.30 | 5.65 | 13.97 | 0.86 | 1.94 | 4.51 |
| 2006 | 3.03 | 9.19 | 32.51 | 2.36 | 5.68 | 13.82 | 0.86 | 1.95 | 4.39 |
| 2007 | 2.78 | 8.57 | 29.40 | 2.08 | 5.06 | 12.51 | 0.72 | 1.66 | 3.81 |
| 2008 | 2.70 | 7.95 | 27.31 | 2.02 | 4.81 | 11.90 | 0.68 | 1.58 | 3.68 |
| 2009 | 2.62 | 7.69 | 25.27 | 1.91 | 4.68 | 11.43 | 0.69 | 1.57 | 3.54 |
| 2010 | 2.19 | 6.95 | 23.97 | 1.77 | 4.30 | 10.56 | 0.62 | 1.39 | 3.22 |
| 2011 | 2.03 | 6.37 | 21.30 | 1.59 | 3.88 | 9.59 | 0.57 | 1.30 | 2.91 |
| 2012 | 1.95 | 5.96 | 19.94 | 1.63 | 3.91 | 9.38 | 0.57 | 1.30 | 2.87 |
| 2013 | 1.82 | 5.55 | 19.62 | 1.51 | 3.64 | 8.86 | 0.52 | 1.18 | 2.68 |
| 2014 | 1.66 | 5.14 | 17.33 | 1.47 | 3.46 | 8.45 | 0.50 | 1.11 | 2.50 |
| 2015 | 1.58 | 4.83 | 16.28 | 1.37 | 3.26 | 8.05 | 0.45 | 1.03 | 2.32 |
| 2016 | 1.60 | 4.64 | 15.69 | 1.37 | 3.29 | 7.82 | 0.47 | 1.03 | 2.24 |
| 2017 | 1.39 | 4.17 | 14.76 | 1.16 | 2.85 | 7.12 | 0.38 | 0.85 | 1.94 |
| 2018 | 1.33 | 4.15 | 14.36 | 1.18 | 2.83 | 7.10 | 0.37 | 0.83 | 1.90 |

TABLE 3-15: WHITE PERCH PREDICTED TRI+ CONCENTRATIONS FOR 1993 - 2018

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 6.78 | 20.61 | 60.01 | 3.53 | 20.43 | 58.28 | 1.05 | 3.07 | 8.85 |
| 1994 | 6.08 | 18.92 | 55.70 | 3.11 | 18.84 | 54.09 | 0.93 | 2.82 | 8.32 |
| 1995 | 6.04 | 18.14 | 53.62 | 3.17 | 18.05 | 52.03 | 0.96 | 2.76 | 7.91 |
| 1996 | 5.03 | 15.63 | 45.59 | 2.38 | 15.50 | 44.32 | 0.72 | 2.17 | 6.35 |
| 1997 | 4.57 | 14.62 | 43.14 | 2.18 | 14.49 | 41.84 | 0.69 | 2.03 | 5.76 |
| 1998 | 4.10 | 13.25 | 39.03 | 2.05 | 13.47 | 37.80 | 0.57 | 1.68 | 4.88 |
| 1999 | 3.69 | 12.02 | 35.66 | 1.65 | 11.88 | 34.40 | 0.51 | 1.52 | 4.36 |
| 2000 | 3.42 | 11.08 | 32.99 | 1.63 | 10.98 | 31.91 | 0.48 | 1.37 | 3.94 |
| 2001 | 3.23 | 10.48 | 31.22 | 1.57 | 10.32 | 30.20 | 0.45 | 1.27 | 3.64 |
| 2002 | 3.03 | 9.82 | 29.31 | 1.57 | 9.72 | 28.37 | 0.42 | 1.21 | 3.47 |
| 2003 | 2.82 | 9.13 | 27.09 | 1.39 | 9.02 | 26.22 | 0.40 | 1.14 | 3.28 |
| 2004 | 2.59 | 8.38 | 24.87 | 1.34 | 8.27 | 24.08 | 0.37 | 1.03 | 2.94 |
| 2005 | 2.40 | 7.83 | 23.33 | 1.17 | 7.69 | 22.63 | 0.33 | 0.94 | 2.70 |
| 2006 | 2.30 | 7.47 | 22.23 | 1.13 | 7.37 | 21.47 | 0.32 | 0.87 | 2.48 |
| 2007 | 2.14 | 7.00 | 20.75 | 1.03 | 6.86 | 20.10 | 0.30 | 0.82 | 2.31 |
| 2008 | 2.02 | 6.55 | 19.51 | 0.96 | 6.45 | 18.90 | 0.27 | 0.74 | 2.10 |
| 2009 | 1.93 | 6.21 | 18.45 | 0.94 | 6.14 | 17.87 | 0.26 | 0.71 | 2.01 |
| 2010 | 1.75 | 5.73 | 16.85 | 0.90 | 5.58 | 16.24 | 0.28 | 0.71 | 1.93 |
| 2011 | 1.57 | 5.13 | 15.18 | 0.81 | 5.02 | 14.63 | 0.25 | 0.66 | 1.78 |
| 2012 | 1.48 | 4.76 | 14.16 | 0.77 | 4.69 | 13.68 | 0.26 | 0.62 | 1.70 |
| 2013 | 1.38 | 4.47 | 13.24 | 0.78 | 4.39 | 12.73 | 0.25 | 0.59 | 1.58 |
| 2014 | 1.29 | 4.17 | 12.32 | 0.72 | 4.08 | 11.87 | 0.22 | 0.53 | 1.46 |
| 2015 | 1.21 | 3.91 | 11.61 | 0.69 | 3.82 | 11.22 | 0.21 | 0.51 | 1.38 |
| 2016 | 1.17 | 3.69 | 10.98 | 0.67 | 3.64 | 10.55 | 0.21 | 0.49 | 1.33 |
| 2017 | 1.07 | 3.44 | 10.23 | 0.59 | 3.37 | 9.83 | 0.18 | 0.43 | 1.14 |
| 2018 | 1.03 | 3.33 | 10.07 | 0.56 | 3.29 | 9.69 | 0.15 | 0.40 | 1.11 |

TABLE 3-16: YELLOW PERCH PREDICTED TRI+ CONCENTRATIONS FOR 1993 - 2018

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------------------|------------------------------------|--|----------------------------------|------------------------------------|--|----------------------------------|------------------------------------|--|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 4.08 | 12.79 | 38.05 | 2.21 | 5.34 | 13.00 | 0.62 | 1.86 | 5.29 |
| 1994 | 3.47 | 11.25 | 34.55 | 1.88 | 4.57 | 11.83 | 0.58 | 1.73 | 4.90 |
| 1995 | 3.58 | 11.75 | 35.10 | 2.00 | 4.67 | 12.42 | 0.59 | 1.79 | 5.06 |
| 1996 | 2.94 | 9.50 | 28.63 | 1.44 | 3.62 | 9.74 | 0.44 | 1.33 | 3.85 |
| 1997 | 2.60 | 8.69 | 26.89 | 1.46 | 3.38 | 9.04 | 0.43 | 1.23 | 3.45 |
| 1998 | 2.28 | 7.78 | 24.10 | 1.13 | 3.03 | 8.14 | 0.35 | 1.04 | 2.99 |
| 1999 | 2.08 | 7.08 | 22.16 | 1.04 | 2.63 | 7.26 | 0.32 | 0.91 | 2.61 |
| 2000 | 1.90 | 6.56 | 20.29 | 0.95 | 2.47 | 6.78 | 0.32 | 0.85 | 2.33 |
| 2001 | 1.82 | 6.21 | 19.17 | 1.02 | 2.51 | 6.66 | 0.28 | 0.79 | 2.17 |
| 2002 | 1.68 | 5.82 | 18.05 | 0.90 | 2.35 | 6.41 | 0.25 | 0.74 | 2.05 |
| 2003 | 1.56 | 5.39 | 16.80 | 0.85 | 2.17 | 6.00 | 0.27 | 0.71 | 1.95 |
| 2004 | 1.43 | 4.97 | 15.45 | 0.77 | 2.04 | 5.60 | 0.23 | 0.64 | 1.78 |
| 2005 | 1.35 | 4.61 | 14.37 | 0.68 | 1.81 | 5.05 | 0.22 | 0.58 | 1.60 |
| 2006 | 1.28 | 4.42 | 13.67 | 0.69 | 1.79 | 4.88 | 0.20 | 0.54 | 1.49 |
| 2007 | 1.19 | 4.12 | 12.75 | 0.60 | 1.62 | 4.46 | 0.21 | 0.52 | 1.39 |
| 2008 | 1.12 | 3.88 | 12.01 | 0.55 | 1.50 | 4.23 | 0.17 | 0.46 | 1.26 |
| 2009 | 1.12 | 3.69 | 11.38 | 0.60 | 1.52 | 4.07 | 0.17 | 0.45 | 1.22 |
| 2010 | 0.99 | 3.37 | 10.61 | 0.52 | 1.39 | 3.86 | 0.17 | 0.45 | 1.18 |
| 2011 | 0.88 | 3.02 | 9.49 | 0.46 | 1.25 | 3.49 | 0.16 | 0.41 | 1.09 |
| 2012 | 0.84 | 2.83 | 8.70 | 0.52 | 1.25 | 3.39 | 0.18 | 0.41 | 1.04 |
| 2013 | 0.78 | 2.64 | 8.21 | 0.46 | 1.18 | 3.20 | 0.16 | 0.38 | 0.96 |
| 2014 | 0.73 | 2.45 | 7.57 | 0.44 | 1.10 | 2.98 | 0.14 | 0.34 | 0.89 |
| 2015 | 0.68 | 2.30 | 7.16 | 0.41 | 1.04 | 2.86 | 0.13 | 0.33 | 0.84 |
| 2016 | 0.65 | 2.19 | 6.74 | 0.43 | 1.05 | 2.78 | 0.15 | 0.34 | 0.82 |
| 2017 | 0.59 | 2.02 | 6.28 | 0.34 | 0.92 | 2.55 | 0.11 | 0.28 | 0.71 |
| 2018 | 0.58 | 1.98 | 6.08 | 0.33 | 0.88 | 2.45 | 0.10 | 0.25 | 0.66 |

TABLE 3-17
EXPOSURE PARAMETERS FOR THE TREE SWALLOW (*Tachycineta bicolor*)

| | Exposure Parameters | Range Reported for Species |
|---|------------------------|--------------------------------|
| Common Name | Tree Swallow | - |
| Genus | <i>Tachycineta</i> | - |
| Species | <i>bicolor</i> | - |
| Sex (M/F) | Female Male | - |
| Age (Adult/Juv.) | Adult, Breeding | - |
| Male/Female Body Weight (kg) ¹ | 0.0210 0.0206 | 0.017-0.0255 (M and F) |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.018 0.018 | 0.016-0.020 |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.005 - | No Contact with Sediments |
| General Dietary Characterization | Insectivore | - |
| Percent Diet Composition (% wet wt.) ⁴ | | |
| Fish (Total Component) | 0% | 0% |
| Aquatic Invertebrates (Total Component) ⁵ | 100% | 95.0% - 100.0% |
| Non-river Related Diet Sources | 0% | 0% |
| Water Consumption Rate (L/day) ⁶ | 0.0044 | 0.0038-0.0050 |
| Percent Incidental Sediment Ingestion in Diet ⁷ | 0.00% | No Contact with Sediments |
| Foraging Territory (km) ⁸ | 0.1 | 0.1-0.2 |
| Behavioral Modification Factors in the Exposure Assessment ⁹ | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | - |
| Temporal Hibernation/Asivation Correction Factor (1-% Temporal Hibernation/Asivation) | 1 | - |
| Habitat Use Factor (Temporal use factor %) | 1 | Feeds over open water habitats |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{10, 11} | April - June | April - June |

Notes: ¹ Secord and McCarty (1997), Robertson et al. (1992); ² Estimated from Nagy (1987) and USEPA (December, 1993); ³ No contact with sediments; ⁴ Secord and McCarty (1997), McCarty and Winkler (In Press); ⁵ Emergent forms of insects with partial aquatic life histories; ⁶ Calder and Braun (1983 In USEPA Report, December 1993), Davis (1982); ⁷ Robertson et al. (1992); ⁸ McCarty and Winkler (In Press); ⁹ Robertson et al. (1992), see text for rationale; ¹⁰ Bull (1998), And (1988).

TABLE 3-18
EXPOSURE PARAMETERS FOR THE MALLARD (*Anas platyrhynchos*)

| | Exposure Parameters | | Range Reported for Species |
|---|----------------------------|-------|-----------------------------------|
| Common Name | Mallard | | - |
| Genus | <i>Anas</i> | | - |
| Species | <i>platyrhynchos</i> | | - |
| Sex (M/F) | Female | Male | - |
| Age (Adult/Juv.) | Adult, Breeding | | - |
| Male/Female Body Weight (kg) ¹ | 1.06 | 1.24 | 1.01 - 1.11 F/M 1.21 - 1.27 |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.292 | 0.322 | 0.270-0.279 F/0.317-0.326 M |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.061 | 0.067 | 0.058-0.063 F/ 0.066-0.068 M |
| General Dietary Characterization | Opportunistic Omnivore | | - |
| Percent Diet Composition (% wet wt.) ⁴ | | | |
| Fish (Total Component) | 0% | | 0% |
| Aquatic Invertebrates (Total Component) | 50% | | 10 - 100% |
| Aquatic Vegetation/Seeds | 50% | | 8 - 90 % |
| Water Consumption Rate (L/day) ⁵ | 0.061 | 0.068 | 0.059-0.063 F/ 0.067 - 0.069 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 2.00% | | 2.00% |
| Foraging Territory (km) ⁷ | 540.0 | 620.0 | 40.0-1440.0 Ha |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | | |
| Temporal Migration CorrectionFactor (1-% Annual Temporal Displaceme | 1 | | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/ | 1 | | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{9,10} | February -May | | February -May |

Notes: ¹ Dunning (1993), USEPA (December 1993); ² Estimated from Nagy (1987) and USEPA (December 1993); ³ Estimated from USEPA (December 1993); ⁴ Average of diet study summaries presented in USEPA (December 1993); ⁵ Calder and Braun (1983 In USEPA, December 1993); ⁶ Beyer et al. (1994); ⁷ Kirby et al. (1985 In USEPA, December 1993); ⁸ Bull (1998), USEPA (December 1993); ^{9,10} Bull (1998), Andrie and Carroll (1988).

TABLE 3-19
EXPOSURE PARAMETERS FOR BELTED KINGFISHER (*Ceryle alcyon*)

| | Exposure Parameters | Range Reported for Species |
|---|----------------------------|-----------------------------------|
| Common Name | Belted Kingfisher | - |
| Genus | <i>Ceryle</i> | - |
| Species | <i>alcyon</i> | - |
| Sex (M/F) | Female Male | - |
| Age (Adult/Juv.) | Adult, Breeding | - |
| Male/Female Body Weight (kg) ¹ | 0.147 0.147 | 0.136-0.158 M and F |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.058 0.058 | 0.055-0.060 M and F |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.017 0.017 | - |
| General Dietary Characterization | Opportunistic Piscivore | - |
| Percent Diet Composition (% wet wt.) ⁴ | | |
| Fish (Total Component) | 78% | 46% - 100% |
| Aquatic Invertebrates (Total Component) | 22% | 5% - 41% |
| Non-river Related Diet Sources | 0% | 0-4.3% |
| Water Consumption Rate (L/day) ⁵ | 0.016 | 0.015-0.017 |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 1.00% | nests in banks, grooming |
| Foraging Territory (km) ⁷ | 0.70 | 0.389-1.03 |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Hatching) ^{9,10} | April - June | April - June |
| Notes: ¹ Brooks and Davis (1987), Poole (1932); ² Estimated from Nagy (1987) and USEPA (December 1993); ³ Estimated from USEPA (1993b); ⁴ Gould unpublished data (In USEPA, December 1993), Davis (1982); ⁵ Calder and Braun (1983 In USEPA December 1993); ⁶ Best Professional Judgment based on Davis (1982); ⁷ Davis (1982); ⁸ Bull (1998), USEPA (December 1993); ^{9,10} Bull (1998), Andrle and Carroll (1988). | | |

TABLE 3-20
EXPOSURE PARAMETERS FOR GREAT BLUE HERON (*Ardea herodias*)

| | Exposure Parameters | | Range Reported for Species |
|--|----------------------------|-------|-----------------------------------|
| Common Name | Great Blue Heron | | - |
| Genus | <i>Ardea</i> | | - |
| Species | <i>herodias</i> | | - |
| Sex (M/F) | Female | Male | - |
| Age (Adult/Juvenile) | Adult, Breeding | | - |
| Male/Female Body Weight (kg) ¹ | 2.20 | 2.58 | 1.87-2.54 F/ 2.28-2.88 M |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.352 | 0.390 | 0.284-0.431 F/ 0.331-0.455 M |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.097 | 0.108 | |
| General Dietary Characterization | Opportunistic Piscivore | | - |
| Percent Diet Composition (% wet wt.) ⁴ | | | |
| Fish (Total Component) | 98% | | 72-98% |
| Aquatic Invertebrates (Total Component) | 1% | | 1-18% |
| Non-river Related Diet Sources | 1% | | 0-4.3% |
| Water Consumption Rate (L/day) ⁵ | 0.100 | 0.111 | 0.089-0.110 F/ 0.102-0.119 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 2.00% | | - |
| Foraging Territory (km) ⁷ | 0.98 | | 0.6-1.37 |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | | |
| Temporal Migration CorrectionFactor (1-% Annual Temporal Displacement) | 1 | | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{9,10} | March - June | | March -June |
| Notes: ¹ Dunning (1993) ; ² Estimated from Nagy (1987) and USEPA (December 1993); ³ Estimated from USEPA (1993b); ⁴ Alexander (1977 In USEPA, December 1993), Cotaam and Uhler (1945); ⁵ Calder and Braun (1983 In USEPA, December 1993); ⁶ Best Professional Judgement based on Eckert and Karalus (1988); ⁷ Peifer (1979 In USEPA (December, 1993); ⁸ USEPA (December, 1993); ^{9,10} Bull (1998) and Andrle and Carroll (1988). | | | |

TABLE 3-21
EXPOSURE PARAMETERS FOR BALD EAGLE (*Haliaeetus leucocephalus*)

| | Exposure Parameters | Range Reported for Species |
|---|-------------------------|-----------------------------|
| Common Name | Bald Eagle | - |
| Genus | <i>Haliaeetus</i> | - |
| Species | <i>leucocephalus</i> | - |
| Sex (M/F) | Female Male | - |
| Age (Adult/Juvenile) | Adult, Breeding | - |
| Male/Female Body Weight (kg) ¹ | 5.10 3.20 | 4.5-5.6 F/M 3.0-3.4 |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.65 0.46 | 0.60-0.69 F/0.46-0.49 M |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | - - | - |
| General Dietary Characterization ⁴ | Opportunistic Piscivore | - |
| Percent Diet Composition (% wet wt.) ⁴ | | |
| Fish (Total Component) | 100% | 70-100% |
| Aquatic Invertebrates (Total Component) | 0% | 0-18% |
| Non-river Related Diet Sources | 0% | 0-4.3% |
| Water Consumption Rate (L/day) ⁵ | 0.175 0.129 | 0.162-0.187 F/0.123-0.134 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 0.00% | 0.00% |
| Foraging Territory (km) ⁷ | 5.0 | 3.0-7.0 Km |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | Resident |
| Temporal Hibernation/Asivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{9,10} | February - May | February - May |

¹ Bopp (1999), USEPA (December 1993), Dunning (1993); ^{2, 3} Estimated from Nagy (1987) and USEPA (December 1993);
⁴ Nye (1999), Bull (1998), USEPA (December 1993), Nye and Suring (1978); ⁵ Caluder and Braun (1983 In USEPA December 1993);
⁶ Best Professional Judgement - USEPA (December 1993);
⁷ Craig et al. (1988 In USEPA, December 1993); ⁸ Nye (1999), USEPA (December 1993); ^{9, 10} Nye (1999), Andrlle and Carroll (1988).

TABLE 3-22
EXPOSURE PARAMETERS FOR LITTLE BROWN BAT (*Myotis lucifugus*)

| | Exposure Parameters | Proximal Range Reported for Species |
|---|---|-------------------------------------|
| Common Name | Little Brown Bat | - |
| Genus | <i>Myotis</i> | - |
| Species | <i>lucifugus</i> | - |
| Sex (M/F) | Female Male | - |
| Age (Adult/Juv.) | Adult, Breeding | - |
| Male/Female Body Weight (kg) ¹ | 0.0071 0.0069 | 0.0042-0.0094 /0.0055-0.0077 |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.0025 0.0025 | 0.0025-0.0037 F/ No Male Data |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | - - | - |
| General Dietary Characterization ⁴ | Insectivore | - |
| Percent Diet Composition (% wet wt.) ⁴ | | |
| Fish (Total Component) | 0.0% | 0% |
| Aquatic Invertebrates (Total Component) | 100.0% | 87.0 % - 100.0% |
| Non-river Related Diet Sources | 0.0% | 0 % - 13.0 % |
| Water Consumption Rate (L/day) ⁵ | 0.0011 0.0011 | Based upon 0.007 Kg |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 0.00% 0.00% | 0.00% |
| Home Range (km) ⁷ | 0.1 >0.1 | 0.1 - >0.1 |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | |
| Temporal Migration CorrectionFactor (1-% Annual Temporal Displacement) | 1 | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | See text |
| Habitat Use Factor (Temporal use factor %) | 1 | Feeds over waterbody |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{9, 10} | April to July - | April to July |
| ¹ Bopp (1999); ² Fenton and Barclay (1980); ³ Dry weight basis of ingestion not required; ⁴ Anthony and Kunz (1977), Belwood and Fenton (1976), Buchler (1976); ⁵ Farrell and Wood (1968c In USEPA, December 1993); ⁶ No contact with sediments; ⁷ Bulcher (1976); ⁸ Davis and Hitchcock (1965); ^{9, 10} Belwood and Fenton (1976), Wimbatt (1945). | | |

TABLE 3-23
EXPOSURE PARAMETERS FOR RACCOON (*Procyon lotor*)

| | Exposure Parameters | | Proximal Range Reported for Species |
|---|------------------------|-------|-------------------------------------|
| Common Name | Raccoon | | - |
| Genus | <i>Procyon</i> | | - |
| Species | <i>lotor</i> | | - |
| Sex (M/F) | Female | Male | - |
| Age (Adult/Juv.) | Adult, Breeding | | - |
| Male/Female Body Weight (kg) ¹ | 6.400 | 7.600 | 5.6-7.1 F/7.0-8.3 M |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.99 | 1.20 | 0.866-1.1 F/1.1-1.30 M |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.316 | 0.364 | 0.283-0.344 F/0.340-0.391 M |
| General Dietary Characterization ⁴ | Opportunistic Omnivore | | - |
| Percent Diet Composition (% wet wt.) ⁴ | | | |
| Fish (Total Component) | 3.0% | | 0-3% |
| Aquatic Invertebrates (Total Component) | 37.0% | | 1.4-37.0% |
| Non-river Related Diet Sources | 60.0% | | 0-1.5% |
| Water Consumption Rate (L/day) ⁵ | 0.526 | 0.614 | 0.467-0.578 F/0.571-0.665 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 9.4% | 9.4% | 9.40% |
| Home Range (km) ⁷ | 48.0 | 48.0 | 5.3-376 F/18.2-814 M |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | | Resident |
| Temporal Hibernation/Asivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ^{9, 10} | January to May | - | January to May |

¹ Bopp (1999), Sanderson (1984), USEPA (December 1993); ^{2, 3} Estimated from NFMR and ME in USEPA (December 1993) and Nagy (1987);
⁴ Tabatabai and Kennedy (1988), Newell et al. (1987), Llewellyn and Uhler (1952), Hamilton (1951); ⁵ Farrell and Wood (1968c In USEPA, 1993a);
⁶ Beyer et al. (1994); ⁷ Urban (1970), Stuewer (1943); ⁸ USEPA (December, 1993), Hamilton (1951); ^{9, 10} USEPA (December, 1993), Stuewer (1943).

TABLE 3-24
EXPOSURE PARAMETERS FOR MINK (*Mustela vison*)

| | Exposure Parameters | | Proximal Range Reported for Species |
|--|-----------------------------------|-------|--|
| Common Name | Mink | | - |
| Genus | <i>Mustela</i> | | - |
| Species | <i>vison</i> | | - |
| Sex (M/F) | Female | Male | - |
| Age (Adult/Juv.) | Adult, Breeding | | - |
| Male/Female Body Weight (kg) ¹ | 0.83 | 1.02 | 0.550-1.101 F/0.681-1.362 M |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.132 | 0.132 | 0.145 F/ 0.119 M |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.059 | 0.069 | 0.042-1.013 F/0.050-0.089 M |
| General Dietary Characterization ⁴ | Opportunistic Piscivore/Carnivore | | - |
| Percent Diet Composition (% wet wt.) ⁴ | | | |
| Fish (Total Component) | 34.0% | | 18.8-34.0% |
| Aquatic Invertebrates (Total Component) | 16.5% | | 13.9-16.5% |
| Non-river Related Diet Sources | 49.5% | | 49.5 % - 67.0 % |
| Water Consumption Rate (L/day) ⁵ | 0.084 | 0.101 | 0.052-0.107 F/0.070-0.131 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 1.0% | | 1.0% |
| Home Range (km) ⁷ | 1.9 | 3.4 | 1.0-2.8 km F/1.8-5.0 km M |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ⁸ | March to June | | March to June |
| ¹ Mitchell (1961); J. Bopp (1999), ² Bleavins and Aulerich (1981); ³ Estimated from Nagy (1987) and USEPA (December, 1993); ⁴ Hamilton (1951), Hamilton (1940), Hamilton (1936); ⁵ Farrell and Wood (1968c In USEPA, December 1993); ⁶ Best Professional Judgement - based upon observations in Hamilton (1940); ⁷ Gerell (1970), Mitchell (1961); ⁸ Allen (1986). | | | |

TABLE 3-25
EXPOSURE PARAMETERS FOR RIVER OTTER (*Lutra canadensis*)

| | Exposure Parameters | Proximal Range Reported for Species |
|---|------------------------------|-------------------------------------|
| Common Name | River Otter | - |
| Genus | <i>Lutra</i> | - |
| Species | <i>canadensis</i> | - |
| Sex (M/F) | Female Male | - |
| Age (Adult/Juv.) | Adult, Breeding | - |
| Male/Female Body Weight (kg) ¹ | 7.32 10.9 | 6.73-7.90 F/9.20-12.7 M |
| Total Daily Dietary Ingestion (kg/day wet wt.) ² | 0.900 0.900 | 0.7-1.1 |
| Total Daily Dietary Ingestion (kg/day dry wt.) ³ | 0.353 0.491 | 0.329-0.376 F/0.425-0.555 M |
| General Dietary Characterization ⁴ | Opportunistic Piscivore | - |
| Percent Diet Composition (% wet wt.) ⁴ | | |
| Fish (Total Component) | 100% | 70-100% |
| Aquatic Invertebrates (Total Component) | 0.0% | 5-15% |
| Non-river Related Diet Sources | 0.0% | 0-25% |
| Water Consumption Rate (L/day) ⁵ | 0.594 0.853 | 0.551-0.636 F/0.730-0.975 M |
| Percent Incidental Sediment Ingestion in Diet ⁶ | 1.0% | 1.0% |
| Home Range (km) ⁷ | 10.0 | 1.5-22.3 Km |
| Behavioral Modification Factors in the Exposure Assessment ⁸ | | |
| Temporal Migration Correction Factor (1-% Annual Temporal Displacement) | 1 | Resident |
| Temporal Hibernation/Asetivation Correction Factor (1-% Temporal Hib/Aset.) | 1 | Active Year Round |
| Habitat Use Factor (Temporal use factor %) | 1 | Riparian habitats preferred |
| Temporal Reproductive Period (Mating/Gestation/Birth) ⁹ | March to March ¹⁰ | March to March |
| ¹ Spinola et al., (undated), Bopp (1999), USEPA (December 1993); ^{2, 3} Harris (1968 In USEPA, December 1993), Penrod (1999); ⁴ Spinola (1999), Newell et al. (1987), Hamilton (1961); ⁵ Farrell and Wood (1968c In USEPA, December 1993); ⁶ Best Professional Judgement - based upon Liers (1951) In USEPA, 1993; ⁷ Spinola et al. (undated); ⁸ USEPA (December 1993a); ⁹ Hamilton and Eadie (1964); ¹⁰ Period between mating and birth extends for one full year due to delayed implantation of zygote. | | |

**TABLE 3-26: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE TREE SWALLOW BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water Expected | Benthic Invertebrate Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Average Predicted Egg Conc (mg/Kg) |
|----------------------------|-------------------------------|-------------------------------------|--|---|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 1.54E-05 | 1.09E+01 | 1.09E+01 | 2.54E+01 |
| Stillwater (168) | 2.74E-05 | 1.91E+01 | 1.91E+01 | 4.45E+01 |
| Federal Dam (154) | 1.92E-05 | 4.54E+00 | 4.54E+00 | 1.06E+01 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.48E-05 | 6.33E-01 | 6.33E-01 | 1.48E+00 |
| 137.2 | 1.48E-05 | 1.25E+00 | 1.25E+00 | 2.91E+00 |
| 122.4 | 6.79E-06 | 6.89E-01 | 6.89E-01 | 1.61E+00 |
| 113.8 | 6.79E-06 | 7.09E-01 | 7.09E-01 | 1.65E+00 |
| 100 | 6.79E-06 | 3.26E-01 | 3.26E-01 | 7.60E-01 |
| 88.9 | 4.47E-06 | 1.63E-01 | 1.63E-01 | 3.81E-01 |
| 58.7 | 4.47E-06 | 5.03E-01 | 5.03E-01 | 1.17E+00 |
| 47.3 | 4.47E-06 | 5.71E-01 | 5.71E-01 | 1.33E+00 |
| 25.8 | 4.47E-06 | 1.69E-01 | 1.69E-01 | 3.95E-01 |

**TABLE 3-27: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE TREE SWALLOW BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Benthic Invertebrate 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) | 95% UCL Predicted Egg Conc (mg/Kg) |
|----------------------------|------------------------------|------------------------------------|--|---|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 4.88E-05 | 1.90E+01 | 1.90E+01 | 4.44E+01 |
| Stillwater (168) | 8.69E-05 | 8.93E+01 | 8.93E+01 | 2.08E+02 |
| Federal Dam (154) | 4.11E-05 | 7.72E+00 | 7.72E+00 | 1.80E+01 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.61E-04 | 1.55E+00 | 1.55E+00 | 3.62E+00 |
| 137.2 | 1.61E-04 | 5.06E+00 | 5.06E+00 | 1.18E+01 |
| 122.4 | 8.70E-05 | 1.73E+00 | 1.73E+00 | 4.04E+00 |
| 113.8 | 8.70E-05 | 2.75E+00 | 2.75E+00 | 6.41E+00 |
| 100 | 8.70E-05 | 2.23E+00 | 2.23E+00 | 5.20E+00 |
| 88.9 | 1.99E-05 | 2.90E-01 | 2.90E-01 | 6.77E-01 |
| 58.7 | 1.99E-05 | 4.60E+00 | 4.60E+00 | 1.07E+01 |
| 47.3 | 1.99E-05 | 4.19E+00 | 4.19E+00 | 9.78E+00 |
| 25.8 | 1.99E-05 | 2.87E-01 | 2.87E-01 | 6.69E-01 |

**TABLE 3-28: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE TREE SWALLOW BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|---|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 1.27E+01 | 7.04E+00 | 3.09E+00 | 2.96E+01 | 1.64E+01 |
| 1994 | 1.19E+01 | 6.70E+00 | 2.92E+00 | 2.78E+01 | 1.56E+01 | 6.81E+00 |
| 1995 | 1.11E+01 | 6.30E+00 | 2.68E+00 | 2.60E+01 | 1.47E+01 | 6.25E+00 |
| 1996 | 1.00E+01 | 5.62E+00 | 2.27E+00 | 2.34E+01 | 1.31E+01 | 5.29E+00 |
| 1997 | 9.24E+00 | 5.13E+00 | 1.97E+00 | 2.16E+01 | 1.20E+01 | 4.61E+00 |
| 1998 | 8.47E+00 | 4.63E+00 | 1.72E+00 | 1.98E+01 | 1.08E+01 | 4.01E+00 |
| 1999 | 7.66E+00 | 4.21E+00 | 1.50E+00 | 1.79E+01 | 9.82E+00 | 3.50E+00 |
| 2000 | 7.04E+00 | 3.97E+00 | 1.35E+00 | 1.64E+01 | 9.27E+00 | 3.15E+00 |
| 2001 | 6.65E+00 | 3.88E+00 | 1.27E+00 | 1.55E+01 | 9.04E+00 | 2.96E+00 |
| 2002 | 6.21E+00 | 3.70E+00 | 1.20E+00 | 1.45E+01 | 8.63E+00 | 2.81E+00 |
| 2003 | 5.80E+00 | 3.49E+00 | 1.13E+00 | 1.35E+01 | 8.15E+00 | 2.63E+00 |
| 2004 | 5.39E+00 | 3.21E+00 | 1.02E+00 | 1.26E+01 | 7.48E+00 | 2.38E+00 |
| 2005 | 4.98E+00 | 2.97E+00 | 9.28E-01 | 1.16E+01 | 6.93E+00 | 2.16E+00 |
| 2006 | 4.78E+00 | 2.81E+00 | 8.58E-01 | 1.12E+01 | 6.56E+00 | 2.00E+00 |
| 2007 | 4.46E+00 | 2.64E+00 | 7.82E-01 | 1.04E+01 | 6.16E+00 | 1.83E+00 |
| 2008 | 4.24E+00 | 2.47E+00 | 7.24E-01 | 9.89E+00 | 5.76E+00 | 1.69E+00 |
| 2009 | 4.00E+00 | 2.37E+00 | 7.00E-01 | 9.33E+00 | 5.52E+00 | 1.63E+00 |
| 2010 | 3.62E+00 | 2.21E+00 | 6.57E-01 | 8.45E+00 | 5.15E+00 | 1.53E+00 |
| 2011 | 3.28E+00 | 2.04E+00 | 6.02E-01 | 7.66E+00 | 4.75E+00 | 1.40E+00 |
| 2012 | 3.05E+00 | 1.94E+00 | 5.71E-01 | 7.13E+00 | 4.53E+00 | 1.33E+00 |
| 2013 | 2.83E+00 | 1.85E+00 | 5.36E-01 | 6.60E+00 | 4.31E+00 | 1.25E+00 |
| 2014 | 2.64E+00 | 1.73E+00 | 4.90E-01 | 6.16E+00 | 4.04E+00 | 1.14E+00 |
| 2015 | 2.47E+00 | 1.67E+00 | 4.69E-01 | 5.75E+00 | 3.90E+00 | 1.09E+00 |
| 2016 | 2.34E+00 | 1.59E+00 | 4.31E-01 | 5.46E+00 | 3.71E+00 | 1.01E+00 |
| 2017 | 2.16E+00 | 1.49E+00 | 3.84E-01 | 5.05E+00 | 3.47E+00 | 8.97E-01 |
| 2018 | 2.14E+00 | 1.45E+00 | 3.74E-01 | 5.00E+00 | 3.39E+00 | 8.72E-01 |

TABLE 3-29: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE TREE SWALLOW BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|---|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 1.86E+01 | 1.11E+01 | 3.32E+00 | 4.34E+01 | 2.59E+01 |
| 1994 | 1.76E+01 | 1.06E+01 | 3.16E+00 | 4.12E+01 | 2.48E+01 | 7.36E+00 |
| 1995 | 1.61E+01 | 9.97E+00 | 2.88E+00 | 3.76E+01 | 2.33E+01 | 6.73E+00 |
| 1996 | 1.47E+01 | 8.88E+00 | 2.46E+00 | 3.43E+01 | 2.07E+01 | 5.73E+00 |
| 1997 | 1.35E+01 | 8.15E+00 | 2.12E+00 | 3.14E+01 | 1.90E+01 | 4.94E+00 |
| 1998 | 1.21E+01 | 7.33E+00 | 1.85E+00 | 2.83E+01 | 1.71E+01 | 4.31E+00 |
| 1999 | 1.13E+01 | 6.70E+00 | 1.62E+00 | 2.65E+01 | 1.56E+01 | 3.78E+00 |
| 2000 | 1.01E+01 | 6.28E+00 | 1.45E+00 | 2.36E+01 | 1.47E+01 | 3.38E+00 |
| 2001 | 9.68E+00 | 6.15E+00 | 1.37E+00 | 2.26E+01 | 1.44E+01 | 3.20E+00 |
| 2002 | 9.22E+00 | 5.88E+00 | 1.30E+00 | 2.15E+01 | 1.37E+01 | 3.04E+00 |
| 2003 | 8.48E+00 | 5.55E+00 | 1.22E+00 | 1.98E+01 | 1.29E+01 | 2.84E+00 |
| 2004 | 7.78E+00 | 5.07E+00 | 1.09E+00 | 1.81E+01 | 1.18E+01 | 2.55E+00 |
| 2005 | 7.29E+00 | 4.71E+00 | 1.00E+00 | 1.70E+01 | 1.10E+01 | 2.33E+00 |
| 2006 | 6.96E+00 | 4.47E+00 | 9.28E-01 | 1.63E+01 | 1.04E+01 | 2.16E+00 |
| 2007 | 6.44E+00 | 4.18E+00 | 8.42E-01 | 1.50E+01 | 9.76E+00 | 1.97E+00 |
| 2008 | 6.07E+00 | 3.91E+00 | 7.85E-01 | 1.42E+01 | 9.13E+00 | 1.83E+00 |
| 2009 | 5.73E+00 | 3.75E+00 | 7.57E-01 | 1.34E+01 | 8.75E+00 | 1.77E+00 |
| 2010 | 5.32E+00 | 3.52E+00 | 7.14E-01 | 1.24E+01 | 8.20E+00 | 1.66E+00 |
| 2011 | 4.76E+00 | 3.25E+00 | 6.51E-01 | 1.11E+01 | 7.58E+00 | 1.52E+00 |
| 2012 | 4.44E+00 | 3.08E+00 | 6.17E-01 | 1.04E+01 | 7.18E+00 | 1.44E+00 |
| 2013 | 4.22E+00 | 2.93E+00 | 5.77E-01 | 9.85E+00 | 6.83E+00 | 1.35E+00 |
| 2014 | 3.89E+00 | 2.76E+00 | 5.28E-01 | 9.07E+00 | 6.44E+00 | 1.23E+00 |
| 2015 | 3.67E+00 | 2.66E+00 | 5.06E-01 | 8.57E+00 | 6.21E+00 | 1.18E+00 |
| 2016 | 3.38E+00 | 2.52E+00 | 4.66E-01 | 7.88E+00 | 5.89E+00 | 1.09E+00 |
| 2017 | 3.21E+00 | 2.36E+00 | 4.17E-01 | 7.49E+00 | 5.51E+00 | 9.74E-01 |
| 2018 | 3.20E+00 | 2.30E+00 | 4.03E-01 | 7.46E+00 | 5.37E+00 | 9.41E-01 |

**TABLE 3-30: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE MALLARD BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water Expected | Macrophyte Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Average Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 4.24E-06 | 1.43E-01 | 1.95E+00 | 1.37E-02 | 2.10E+00 | 4.24E+01 |
| Stillwater (168) | 7.53E-06 | 2.31E-01 | 3.63E+00 | 3.57E-02 | 3.90E+00 | 7.91E+01 |
| Federal Dam (154) | 5.26E-06 | 2.25E-01 | 8.66E-01 | 3.21E-03 | 1.09E+00 | 1.89E+01 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 4.07E-06 | 1.28E-01 | 1.21E-01 | 9.90E-04 | 2.50E-01 | 2.63E+00 |
| 137.2 | 4.07E-06 | 1.28E-01 | 2.38E-01 | 1.75E-03 | 3.67E-01 | 5.17E+00 |
| 122.4 | 1.86E-06 | 7.60E-02 | 1.11E-01 | 1.11E-03 | 1.88E-01 | 2.41E+00 |
| 113.8 | 1.86E-06 | 7.60E-02 | 9.52E-02 | 1.16E-03 | 1.72E-01 | 2.07E+00 |
| 100 | 1.86E-06 | 7.60E-02 | 5.23E-02 | 4.59E-04 | 1.29E-01 | 1.14E+00 |
| 88.9 | 1.23E-06 | 5.56E-02 | 2.63E-02 | 8.98E-04 | 8.28E-02 | 5.72E-01 |
| 58.7 | 1.23E-06 | 5.56E-02 | 6.76E-02 | 2.90E-04 | 1.23E-01 | 1.47E+00 |
| 47.3 | 1.23E-06 | 5.56E-02 | 9.18E-02 | 1.77E-03 | 1.49E-01 | 2.00E+00 |
| 25.8 | 1.23E-06 | 5.56E-02 | 2.72E-02 | 6.66E-04 | 8.35E-02 | 5.92E-01 |

**TABLE 3-31: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE MALLARD BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Macrophyte 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (95% UCL) (mg/Kg) |
|----------------------------|------------------------------|-----------------------|------------------------------------|---------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 1.34E-05 | 5.65E-01 | 3.06E+00 | 2.00E-02 | 3.64E+00 | 6.66E+01 |
| Stillwater (168) | 2.39E-05 | 6.19E-01 | 6.32E+00 | 6.23E-02 | 7.01E+00 | 1.38E+02 |
| Federal Dam (154) | 1.13E-05 | 4.68E-01 | 1.51E+00 | 5.39E-03 | 1.98E+00 | 3.28E+01 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 4.43E-05 | 3.83E-01 | 2.10E-01 | 1.08E-03 | 5.94E-01 | 4.57E+00 |
| 137.2 | 4.43E-05 | 3.83E-01 | 4.14E-01 | 3.53E-03 | 8.00E-01 | 9.01E+00 |
| 122.4 | 2.39E-05 | 1.17E+00 | 2.78E-01 | 1.23E-03 | 1.45E+00 | 6.06E+00 |
| 113.8 | 2.39E-05 | 1.17E+00 | 1.66E-01 | 1.92E-03 | 1.34E+00 | 3.61E+00 |
| 100 | 2.39E-05 | 1.17E+00 | 3.58E-01 | 9.91E-03 | 1.54E+00 | 7.79E+00 |
| 88.9 | 5.46E-06 | 1.08E+00 | 4.67E-02 | 2.63E-03 | 1.13E+00 | 1.02E+00 |
| 58.7 | 5.46E-06 | 1.08E+00 | 1.18E-01 | 3.22E-03 | 1.20E+00 | 2.56E+00 |
| 47.3 | 5.46E-06 | 1.08E+00 | 6.74E-01 | 6.91E-03 | 1.76E+00 | 1.47E+01 |
| 25.8 | 5.46E-06 | 1.08E+00 | 4.61E-02 | 1.80E-03 | 1.13E+00 | 1.00E+00 |

TABLE 3-32: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE MALLARD BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 2.36E+00 | 1.47E+00 | 5.59E-01 | 4.45E+01 | 2.46E+01 | 1.08E+01 |
| 1994 | 2.27E+00 | 1.46E+00 | 5.33E-01 | 4.17E+01 | 2.34E+01 | 1.02E+01 |
| 1995 | 2.17E+00 | 1.38E+00 | 5.30E-01 | 3.90E+01 | 2.21E+01 | 9.37E+00 |
| 1996 | 1.78E+00 | 1.08E+00 | 4.00E-01 | 3.52E+01 | 1.97E+01 | 7.93E+00 |
| 1997 | 1.66E+00 | 1.10E+00 | 3.74E-01 | 3.23E+01 | 1.80E+01 | 6.91E+00 |
| 1998 | 1.42E+00 | 8.43E-01 | 3.16E-01 | 2.96E+01 | 1.62E+01 | 6.02E+00 |
| 1999 | 1.31E+00 | 8.03E-01 | 2.81E-01 | 2.68E+01 | 1.47E+01 | 5.25E+00 |
| 2000 | 1.19E+00 | 7.30E-01 | 2.55E-01 | 2.46E+01 | 1.39E+01 | 4.72E+00 |
| 2001 | 1.15E+00 | 7.80E-01 | 2.33E-01 | 2.33E+01 | 1.36E+01 | 4.45E+00 |
| 2002 | 1.07E+00 | 7.10E-01 | 2.15E-01 | 2.17E+01 | 1.29E+01 | 4.22E+00 |
| 2003 | 9.90E-01 | 6.69E-01 | 2.17E-01 | 2.03E+01 | 1.22E+01 | 3.94E+00 |
| 2004 | 9.18E-01 | 6.39E-01 | 1.93E-01 | 1.89E+01 | 1.12E+01 | 3.57E+00 |
| 2005 | 8.49E-01 | 5.47E-01 | 1.81E-01 | 1.74E+01 | 1.04E+01 | 3.25E+00 |
| 2006 | 8.27E-01 | 5.51E-01 | 1.61E-01 | 1.67E+01 | 9.84E+00 | 3.00E+00 |
| 2007 | 7.61E-01 | 4.94E-01 | 1.58E-01 | 1.56E+01 | 9.23E+00 | 2.74E+00 |
| 2008 | 7.19E-01 | 4.47E-01 | 1.33E-01 | 1.48E+01 | 8.64E+00 | 2.53E+00 |
| 2009 | 7.00E-01 | 4.65E-01 | 1.33E-01 | 1.40E+01 | 8.28E+00 | 2.45E+00 |
| 2010 | 6.25E-01 | 4.21E-01 | 1.34E-01 | 1.27E+01 | 7.73E+00 | 2.30E+00 |
| 2011 | 5.56E-01 | 3.67E-01 | 1.17E-01 | 1.15E+01 | 7.13E+00 | 2.11E+00 |
| 2012 | 5.35E-01 | 3.81E-01 | 1.27E-01 | 1.07E+01 | 6.80E+00 | 2.00E+00 |
| 2013 | 4.85E-01 | 3.51E-01 | 1.15E-01 | 9.90E+00 | 6.47E+00 | 1.87E+00 |
| 2014 | 4.56E-01 | 3.28E-01 | 1.10E-01 | 9.23E+00 | 6.06E+00 | 1.71E+00 |
| 2015 | 4.23E-01 | 3.12E-01 | 9.79E-02 | 8.63E+00 | 5.85E+00 | 1.64E+00 |
| 2016 | 4.11E-01 | 3.10E-01 | 1.01E-01 | 8.19E+00 | 5.56E+00 | 1.51E+00 |
| 2017 | 3.68E-01 | 2.65E-01 | 7.78E-02 | 7.58E+00 | 5.21E+00 | 1.35E+00 |
| 2018 | 3.78E-01 | 2.97E-01 | 9.80E-02 | 7.50E+00 | 5.08E+00 | 1.31E+00 |

TABLE 3-33: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE MALLARD BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 3.32E+00 | 2.13E+00 | 5.97E-01 | 6.51E+01 | 3.89E+01 |
| 1994 | 3.20E+00 | 2.10E+00 | 5.72E-01 | 6.17E+01 | 3.72E+01 | 1.10E+01 |
| 1995 | 2.98E+00 | 1.97E+00 | 5.65E-01 | 5.64E+01 | 3.49E+01 | 1.01E+01 |
| 1996 | 2.53E+00 | 1.60E+00 | 4.31E-01 | 5.15E+01 | 3.11E+01 | 8.60E+00 |
| 1997 | 2.34E+00 | 1.59E+00 | 3.98E-01 | 4.71E+01 | 2.85E+01 | 7.41E+00 |
| 1998 | 2.01E+00 | 1.28E+00 | 3.37E-01 | 4.24E+01 | 2.57E+01 | 6.47E+00 |
| 1999 | 1.90E+00 | 1.21E+00 | 3.01E-01 | 3.97E+01 | 2.35E+01 | 5.67E+00 |
| 2000 | 1.69E+00 | 1.10E+00 | 2.72E-01 | 3.55E+01 | 2.20E+01 | 5.07E+00 |
| 2001 | 1.64E+00 | 1.15E+00 | 2.50E-01 | 3.39E+01 | 2.15E+01 | 4.80E+00 |
| 2002 | 1.55E+00 | 1.06E+00 | 2.31E-01 | 3.23E+01 | 2.06E+01 | 4.56E+00 |
| 2003 | 1.42E+00 | 1.00E+00 | 2.32E-01 | 2.97E+01 | 1.94E+01 | 4.26E+00 |
| 2004 | 1.30E+00 | 9.40E-01 | 2.06E-01 | 2.72E+01 | 1.77E+01 | 3.83E+00 |
| 2005 | 1.22E+00 | 8.28E-01 | 1.93E-01 | 2.55E+01 | 1.65E+01 | 3.50E+00 |
| 2006 | 1.18E+00 | 8.19E-01 | 1.73E-01 | 2.44E+01 | 1.56E+01 | 3.25E+00 |
| 2007 | 1.08E+00 | 7.44E-01 | 1.68E-01 | 2.25E+01 | 1.46E+01 | 2.95E+00 |
| 2008 | 1.01E+00 | 6.80E-01 | 1.43E-01 | 2.13E+01 | 1.37E+01 | 2.75E+00 |
| 2009 | 9.79E-01 | 6.89E-01 | 1.43E-01 | 2.00E+01 | 1.31E+01 | 2.65E+00 |
| 2010 | 8.98E-01 | 6.32E-01 | 1.43E-01 | 1.86E+01 | 1.23E+01 | 2.50E+00 |
| 2011 | 7.94E-01 | 5.62E-01 | 1.26E-01 | 1.66E+01 | 1.14E+01 | 2.28E+00 |
| 2012 | 7.59E-01 | 5.65E-01 | 1.35E-01 | 1.55E+01 | 1.08E+01 | 2.16E+00 |
| 2013 | 7.10E-01 | 5.25E-01 | 1.22E-01 | 1.48E+01 | 1.02E+01 | 2.02E+00 |
| 2014 | 6.57E-01 | 4.94E-01 | 1.17E-01 | 1.36E+01 | 9.66E+00 | 1.85E+00 |
| 2015 | 6.18E-01 | 4.72E-01 | 1.04E-01 | 1.29E+01 | 9.31E+00 | 1.77E+00 |
| 2016 | 5.78E-01 | 4.61E-01 | 1.07E-01 | 1.18E+01 | 8.83E+00 | 1.63E+00 |
| 2017 | 5.36E-01 | 4.06E-01 | 8.34E-02 | 1.12E+01 | 8.26E+00 | 1.46E+00 |
| 2018 | 5.48E-01 | 4.35E-01 | 1.04E-01 | 1.12E+01 | 8.06E+00 | 1.41E+00 |

**TABLE 3-34: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE BELTED KINGFISHER BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Average Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 8.01E-06 | 6.37E+00 | 1.09E+00 | 1.35E-02 | 7.48E+00 | 5.68E+02 |
| Stillwater (168) | 1.42E-05 | 2.15E+00 | 1.91E+00 | 3.53E-02 | 4.10E+00 | 3.09E+02 |
| Federal Dam (154) | 9.95E-06 | 5.05E-01 | 4.55E-01 | 3.17E-03 | 9.64E-01 | 7.30E+01 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 7.70E-06 | 5.87E-01 | 6.34E-02 | 9.78E-04 | 6.52E-01 | 4.95E+01 |
| 137.2 | 7.70E-06 | 1.19E+00 | 1.25E-01 | 1.73E-03 | 1.31E+00 | 9.98E+01 |
| 122.4 | 3.53E-06 | 4.53E-01 | 6.91E-02 | 1.09E-03 | 5.24E-01 | 3.97E+01 |
| 113.8 | 3.53E-06 | 4.75E-01 | 7.11E-02 | 1.15E-03 | 5.47E-01 | 4.15E+01 |
| 100 | 3.53E-06 | 2.06E-01 | 3.27E-02 | 4.53E-04 | 2.39E-01 | 1.81E+01 |
| 88.9 | 2.32E-06 | 4.10E-01 | 1.64E-02 | 8.87E-04 | 4.27E-01 | 3.24E+01 |
| 58.7 | 2.32E-06 | 4.47E-01 | 5.05E-02 | 2.86E-04 | 4.97E-01 | 3.78E+01 |
| 47.3 | 2.32E-06 | 3.97E-01 | 5.73E-02 | 1.75E-03 | 4.56E-01 | 3.46E+01 |
| 25.8 | 2.32E-06 | 2.99E-01 | 1.70E-02 | 6.57E-04 | 3.16E-01 | 2.40E+01 |

**TABLE 3-35: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE BELTED KINGFISHER BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (95 % UCL) (mg/Kg) |
|----------------------------|------------------------------|-----------------|------------------------------------|---------------------|--|---|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 2.54E-05 | 1.30E+01 | 1.91E+00 | 1.98E-02 | 1.49E+01 | 1.13E+03 |
| Stillwater (168) | 4.52E-05 | 3.08E+00 | 8.95E+00 | 6.16E-02 | 1.21E+01 | 9.14E+02 |
| Federal Dam (154) | 2.13E-05 | 7.33E-01 | 7.74E-01 | 5.32E-03 | 1.51E+00 | 1.15E+02 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 8.39E-05 | 7.05E-01 | 1.56E-01 | 1.07E-03 | 8.62E-01 | 6.54E+01 |
| 137.2 | 8.39E-05 | 2.58E+00 | 5.07E-01 | 3.49E-03 | 3.09E+00 | 2.34E+02 |
| 122.4 | 4.52E-05 | 7.33E-01 | 1.74E-01 | 1.22E-03 | 9.08E-01 | 6.90E+01 |
| 113.8 | 4.52E-05 | 4.93E-01 | 2.75E-01 | 1.89E-03 | 7.70E-01 | 5.84E+01 |
| 100 | 4.52E-05 | 3.56E-01 | 2.23E-01 | 9.79E-03 | 5.89E-01 | 4.40E+01 |
| 88.9 | 1.03E-05 | 5.62E-01 | 2.91E-02 | 2.60E-03 | 5.94E-01 | 4.49E+01 |
| 58.7 | 1.03E-05 | 5.06E-01 | 4.62E-01 | 3.17E-03 | 9.71E-01 | 7.36E+01 |
| 47.3 | 1.03E-05 | 5.27E-01 | 4.20E-01 | 6.82E-03 | 9.54E-01 | 7.20E+01 |
| 25.8 | 1.03E-05 | 3.59E-01 | 2.88E-02 | 1.78E-03 | 3.90E-01 | 2.95E+01 |

TABLE 3-36: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE BELTED KINGFISHER BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 7.92E+00 | 3.92E+00 | 1.93E+00 | 5.94E+02 | 2.95E+02 | 1.46E+02 |
| 1994 | 4.33E+00 | 3.26E+00 | 1.51E+00 | 3.21E+02 | 2.45E+02 | 1.13E+02 |
| 1995 | 4.01E+00 | 3.52E+00 | 1.73E+00 | 2.97E+02 | 2.65E+02 | 1.30E+02 |
| 1996 | 4.02E+00 | 2.45E+00 | 1.16E+00 | 2.99E+02 | 1.84E+02 | 8.73E+01 |
| 1997 | 3.12E+00 | 2.17E+00 | 1.07E+00 | 2.31E+02 | 1.63E+02 | 8.02E+01 |
| 1998 | 2.72E+00 | 1.94E+00 | 8.71E-01 | 2.01E+02 | 1.45E+02 | 6.55E+01 |
| 1999 | 2.36E+00 | 1.58E+00 | 6.60E-01 | 1.74E+02 | 1.18E+02 | 4.95E+01 |
| 2000 | 1.98E+00 | 1.47E+00 | 5.82E-01 | 1.46E+02 | 1.10E+02 | 4.37E+01 |
| 2001 | 1.92E+00 | 1.54E+00 | 6.55E-01 | 1.42E+02 | 1.16E+02 | 4.93E+01 |
| 2002 | 1.79E+00 | 1.48E+00 | 6.02E-01 | 1.32E+02 | 1.11E+02 | 4.53E+01 |
| 2003 | 1.71E+00 | 1.27E+00 | 5.38E-01 | 1.26E+02 | 9.52E+01 | 4.04E+01 |
| 2004 | 1.52E+00 | 1.21E+00 | 5.25E-01 | 1.12E+02 | 9.07E+01 | 3.95E+01 |
| 2005 | 1.45E+00 | 1.02E+00 | 4.32E-01 | 1.07E+02 | 7.61E+01 | 3.25E+01 |
| 2006 | 1.29E+00 | 1.05E+00 | 4.24E-01 | 9.48E+01 | 7.84E+01 | 3.19E+01 |
| 2007 | 1.27E+00 | 9.57E-01 | 3.58E-01 | 9.33E+01 | 7.17E+01 | 2.69E+01 |
| 2008 | 1.16E+00 | 8.52E-01 | 3.19E-01 | 8.55E+01 | 6.38E+01 | 2.40E+01 |
| 2009 | 1.11E+00 | 9.03E-01 | 3.71E-01 | 8.19E+01 | 6.78E+01 | 2.79E+01 |
| 2010 | 1.07E+00 | 8.93E-01 | 3.39E-01 | 7.86E+01 | 6.70E+01 | 2.55E+01 |
| 2011 | 1.00E+00 | 7.02E-01 | 2.80E-01 | 7.39E+01 | 5.26E+01 | 2.10E+01 |
| 2012 | 8.84E-01 | 8.05E-01 | 3.25E-01 | 6.51E+01 | 6.04E+01 | 2.45E+01 |
| 2013 | 8.43E-01 | 7.51E-01 | 2.76E-01 | 6.22E+01 | 5.64E+01 | 2.08E+01 |
| 2014 | 7.98E-01 | 6.79E-01 | 2.55E-01 | 5.89E+01 | 5.10E+01 | 1.92E+01 |
| 2015 | 7.28E-01 | 6.58E-01 | 2.39E-01 | 5.37E+01 | 4.94E+01 | 1.80E+01 |
| 2016 | 6.93E-01 | 6.47E-01 | 2.73E-01 | 5.11E+01 | 4.86E+01 | 2.05E+01 |
| 2017 | 6.52E-01 | 5.29E-01 | 1.97E-01 | 4.81E+01 | 3.97E+01 | 1.48E+01 |
| 2018 | 6.14E-01 | 5.24E-01 | 1.84E-01 | 4.53E+01 | 3.93E+01 | 1.38E+01 |

TABLE 3-37: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE BELTED KINGFISHER BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 1.35E+01 | 5.24E+00 | 3.39E+00 | 1.01E+03 | 3.96E+02 |
| 1994 | 6.99E+00 | 4.16E+00 | 2.31E+00 | 5.23E+02 | 3.13E+02 | 1.75E+02 |
| 1995 | 6.31E+00 | 4.71E+00 | 2.97E+00 | 4.72E+02 | 3.56E+02 | 2.25E+02 |
| 1996 | 6.55E+00 | 3.15E+00 | 2.03E+00 | 4.91E+02 | 2.37E+02 | 1.54E+02 |
| 1997 | 5.55E+00 | 2.77E+00 | 1.67E+00 | 4.15E+02 | 2.09E+02 | 1.26E+02 |
| 1998 | 4.50E+00 | 2.54E+00 | 1.41E+00 | 3.36E+02 | 1.91E+02 | 1.07E+02 |
| 1999 | 4.14E+00 | 2.04E+00 | 9.89E-01 | 3.10E+02 | 1.53E+02 | 7.46E+01 |
| 2000 | 3.67E+00 | 1.90E+00 | 9.33E-01 | 2.74E+02 | 1.43E+02 | 7.04E+01 |
| 2001 | 3.42E+00 | 1.99E+00 | 1.07E+00 | 2.55E+02 | 1.50E+02 | 8.11E+01 |
| 2002 | 3.26E+00 | 1.90E+00 | 9.52E-01 | 2.43E+02 | 1.43E+02 | 7.19E+01 |
| 2003 | 3.07E+00 | 1.64E+00 | 8.94E-01 | 2.29E+02 | 1.23E+02 | 6.75E+01 |
| 2004 | 2.83E+00 | 1.54E+00 | 8.90E-01 | 2.12E+02 | 1.16E+02 | 6.73E+01 |
| 2005 | 2.65E+00 | 1.32E+00 | 6.95E-01 | 1.98E+02 | 9.89E+01 | 5.25E+01 |
| 2006 | 2.37E+00 | 1.35E+00 | 7.12E-01 | 1.77E+02 | 1.01E+02 | 5.38E+01 |
| 2007 | 2.32E+00 | 1.24E+00 | 6.00E-01 | 1.74E+02 | 9.29E+01 | 4.53E+01 |
| 2008 | 2.08E+00 | 1.10E+00 | 5.16E-01 | 1.55E+02 | 8.27E+01 | 3.90E+01 |
| 2009 | 2.05E+00 | 1.17E+00 | 6.53E-01 | 1.53E+02 | 8.82E+01 | 4.93E+01 |
| 2010 | 1.97E+00 | 1.14E+00 | 5.65E-01 | 1.47E+02 | 8.61E+01 | 4.27E+01 |
| 2011 | 1.74E+00 | 9.17E-01 | 4.82E-01 | 1.30E+02 | 6.89E+01 | 3.64E+01 |
| 2012 | 1.59E+00 | 1.05E+00 | 5.34E-01 | 1.19E+02 | 7.89E+01 | 4.04E+01 |
| 2013 | 1.51E+00 | 9.57E-01 | 4.56E-01 | 1.13E+02 | 7.20E+01 | 3.44E+01 |
| 2014 | 1.43E+00 | 8.81E-01 | 4.27E-01 | 1.06E+02 | 6.63E+01 | 3.23E+01 |
| 2015 | 1.29E+00 | 8.38E-01 | 3.69E-01 | 9.64E+01 | 6.31E+01 | 2.79E+01 |
| 2016 | 1.21E+00 | 8.62E-01 | 5.63E-01 | 9.02E+01 | 6.49E+01 | 4.26E+01 |
| 2017 | 1.22E+00 | 6.91E-01 | 3.66E-01 | 9.13E+01 | 5.20E+01 | 2.77E+01 |
| 2018 | 1.13E+00 | 6.78E-01 | 2.98E-01 | 8.44E+01 | 5.10E+01 | 2.25E+01 |

**TABLE 3-38: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE GREAT BLUE HERON BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 3.34E-06 | 3.24E+00 | 2.01E-02 | 1.05E-02 | 3.27E+00 | 6.28E+02 |
| Stillwater (168) | 5.93E-06 | 1.09E+00 | 3.52E-02 | 2.74E-02 | 1.16E+00 | 2.12E+02 |
| Federal Dam (154) | 4.15E-06 | 2.57E-01 | 8.38E-03 | 2.47E-03 | 2.68E-01 | 4.97E+01 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 3.21E-06 | 2.99E-01 | 1.17E-03 | 7.60E-04 | 3.01E-01 | 5.78E+01 |
| 137.2 | 3.21E-06 | 6.04E-01 | 2.30E-03 | 1.34E-03 | 6.08E-01 | 1.17E+02 |
| 122.4 | 1.47E-06 | 2.31E-01 | 1.27E-03 | 8.50E-04 | 2.33E-01 | 4.46E+01 |
| 113.8 | 1.47E-06 | 2.42E-01 | 1.31E-03 | 8.91E-04 | 2.44E-01 | 4.68E+01 |
| 100 | 1.47E-06 | 1.05E-01 | 6.01E-04 | 3.52E-04 | 1.06E-01 | 2.03E+01 |
| 88.9 | 9.68E-07 | 2.08E-01 | 3.01E-04 | 6.90E-04 | 2.09E-01 | 4.04E+01 |
| 58.7 | 9.68E-07 | 2.27E-01 | 9.28E-04 | 2.23E-04 | 2.28E-01 | 4.40E+01 |
| 47.3 | 9.68E-07 | 2.02E-01 | 1.05E-03 | 1.36E-03 | 2.05E-01 | 3.91E+01 |
| 25.8 | 9.68E-07 | 1.52E-01 | 3.12E-04 | 5.11E-04 | 1.53E-01 | 2.94E+01 |

**TABLE 3-39: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE GREAT BLUE HERON BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Fish 95% UCL | Benthic Invertebrate 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|------------------------------|-----------------|------------------------------------|--|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 1.06E-05 | 6.62E+00 | 3.51E-02 | 6.67E+00 | 1.28E+03 |
| Stillwater (168) | 1.88E-05 | 1.56E+00 | 1.65E-01 | 1.78E+00 | 3.03E+02 |
| Federal Dam (154) | 8.90E-06 | 3.73E-01 | 1.42E-02 | 3.91E-01 | 7.22E+01 |
| <i>Lower River</i> | | | | | |
| 143.5 | 3.50E-05 | 3.59E-01 | 2.86E-03 | 3.62E-01 | 6.94E+01 |
| 137.2 | 3.50E-05 | 1.31E+00 | 9.33E-03 | 1.32E+00 | 2.54E+02 |
| 122.4 | 1.88E-05 | 3.73E-01 | 3.20E-03 | 3.77E-01 | 7.22E+01 |
| 113.8 | 1.88E-05 | 2.51E-01 | 5.07E-03 | 2.57E-01 | 4.85E+01 |
| 100 | 1.88E-05 | 1.81E-01 | 4.11E-03 | 1.93E-01 | 3.50E+01 |
| 88.9 | 4.30E-06 | 2.86E-01 | 5.36E-04 | 2.88E-01 | 5.53E+01 |
| 58.7 | 4.30E-06 | 2.57E-01 | 8.49E-03 | 2.68E-01 | 4.98E+01 |
| 47.3 | 4.30E-06 | 2.68E-01 | 7.73E-03 | 2.81E-01 | 5.18E+01 |
| 25.8 | 4.30E-06 | 1.83E-01 | 5.29E-04 | 1.85E-01 | 3.54E+01 |

**TABLE 3-40: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE GREAT BLUE HERON BASED ON TRI+ CONGENERS
FOR PERIOD 1993 - 2018**

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 3.37E+00 | 1.64E+00 | 8.25E-01 | 6.43E+02 | 3.13E+02 | 1.58E+02 |
| 1994 | 1.58E+00 | 1.32E+00 | 6.18E-01 | 2.98E+02 | 2.51E+02 | 1.18E+02 |
| 1995 | 1.46E+00 | 1.47E+00 | 7.42E-01 | 2.75E+02 | 2.81E+02 | 1.42E+02 |
| 1996 | 1.52E+00 | 9.59E-01 | 4.75E-01 | 2.88E+02 | 1.82E+02 | 9.07E+01 |
| 1997 | 1.11E+00 | 8.44E-01 | 4.42E-01 | 2.08E+02 | 1.60E+02 | 8.44E+01 |
| 1998 | 9.43E-01 | 7.50E-01 | 3.56E-01 | 1.76E+02 | 1.43E+02 | 6.79E+01 |
| 1999 | 8.03E-01 | 5.90E-01 | 2.59E-01 | 1.50E+02 | 1.12E+02 | 4.93E+01 |
| 2000 | 6.44E-01 | 5.44E-01 | 2.28E-01 | 1.19E+02 | 1.03E+02 | 4.33E+01 |
| 2001 | 6.34E-01 | 5.88E-01 | 2.69E-01 | 1.18E+02 | 1.12E+02 | 5.13E+01 |
| 2002 | 5.88E-01 | 5.67E-01 | 2.45E-01 | 1.09E+02 | 1.08E+02 | 4.68E+01 |
| 2003 | 5.69E-01 | 4.69E-01 | 2.17E-01 | 1.06E+02 | 8.88E+01 | 4.13E+01 |
| 2004 | 4.93E-01 | 4.53E-01 | 2.16E-01 | 9.14E+01 | 8.58E+01 | 4.11E+01 |
| 2005 | 4.83E-01 | 3.67E-01 | 1.73E-01 | 8.97E+01 | 6.93E+01 | 3.29E+01 |
| 2006 | 4.09E-01 | 3.90E-01 | 1.72E-01 | 7.56E+01 | 7.38E+01 | 3.29E+01 |
| 2007 | 4.14E-01 | 3.53E-01 | 1.42E-01 | 7.68E+01 | 6.68E+01 | 2.71E+01 |
| 2008 | 3.72E-01 | 3.08E-01 | 1.26E-01 | 6.89E+01 | 5.83E+01 | 2.39E+01 |
| 2009 | 3.59E-01 | 3.40E-01 | 1.53E-01 | 6.66E+01 | 6.44E+01 | 2.92E+01 |
| 2010 | 3.55E-01 | 3.42E-01 | 1.39E-01 | 6.60E+01 | 6.49E+01 | 2.65E+01 |
| 2011 | 3.40E-01 | 2.54E-01 | 1.12E-01 | 6.33E+01 | 4.80E+01 | 2.13E+01 |
| 2012 | 2.92E-01 | 3.11E-01 | 1.36E-01 | 5.42E+01 | 5.91E+01 | 2.60E+01 |
| 2013 | 2.83E-01 | 2.88E-01 | 1.13E-01 | 5.26E+01 | 5.48E+01 | 2.16E+01 |
| 2014 | 2.70E-01 | 2.58E-01 | 1.05E-01 | 5.03E+01 | 4.89E+01 | 2.01E+01 |
| 2015 | 2.43E-01 | 2.50E-01 | 9.76E-02 | 4.52E+01 | 4.75E+01 | 1.86E+01 |
| 2016 | 2.31E-01 | 2.49E-01 | 1.17E-01 | 4.30E+01 | 4.72E+01 | 2.23E+01 |
| 2017 | 2.20E-01 | 1.94E-01 | 8.08E-02 | 4.09E+01 | 3.67E+01 | 1.54E+01 |
| 2018 | 2.02E-01 | 1.93E-01 | 7.46E-02 | 3.75E+01 | 3.66E+01 | 1.42E+01 |

**TABLE 3-41: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR
FEMALE GREAT BLUE HERON BASED ON TRI+ CONGENERS
FOR PERIOD 1993 - 2018**

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 5.89E+00 | 2.11E+00 | 1.10E+00 | 1.13E+03 | 4.03E+02 | 2.10E+02 |
| 1994 | 2.66E+00 | 1.58E+00 | 7.70E-01 | 5.03E+02 | 3.01E+02 | 1.47E+02 |
| 1995 | 2.39E+00 | 1.90E+00 | 9.74E-01 | 4.52E+02 | 3.62E+02 | 1.87E+02 |
| 1996 | 2.58E+00 | 1.16E+00 | 6.35E-01 | 4.91E+02 | 2.20E+02 | 1.22E+02 |
| 1997 | 2.14E+00 | 1.00E+00 | 5.56E-01 | 4.05E+02 | 1.90E+02 | 1.06E+02 |
| 1998 | 1.67E+00 | 9.22E-01 | 4.57E-01 | 3.16E+02 | 1.75E+02 | 8.75E+01 |
| 1999 | 1.53E+00 | 7.01E-01 | 3.21E-01 | 2.89E+02 | 1.32E+02 | 6.13E+01 |
| 2000 | 1.35E+00 | 6.53E-01 | 2.93E-01 | 2.54E+02 | 1.23E+02 | 5.59E+01 |
| 2001 | 1.25E+00 | 7.02E-01 | 3.47E-01 | 2.35E+02 | 1.33E+02 | 6.64E+01 |
| 2002 | 1.19E+00 | 6.71E-01 | 3.11E-01 | 2.24E+02 | 1.27E+02 | 5.94E+01 |
| 2003 | 1.13E+00 | 5.55E-01 | 2.82E-01 | 2.14E+02 | 1.05E+02 | 5.40E+01 |
| 2004 | 1.04E+00 | 5.30E-01 | 2.83E-01 | 1.97E+02 | 1.00E+02 | 5.42E+01 |
| 2005 | 9.78E-01 | 4.34E-01 | 2.22E-01 | 1.85E+02 | 8.16E+01 | 4.24E+01 |
| 2006 | 8.51E-01 | 4.62E-01 | 2.26E-01 | 1.60E+02 | 8.73E+01 | 4.31E+01 |
| 2007 | 8.54E-01 | 4.19E-01 | 1.87E-01 | 1.61E+02 | 7.90E+01 | 3.58E+01 |
| 2008 | 7.49E-01 | 3.63E-01 | 1.62E-01 | 1.41E+02 | 6.84E+01 | 3.10E+01 |
| 2009 | 7.52E-01 | 4.08E-01 | 2.05E-01 | 1.42E+02 | 7.72E+01 | 3.92E+01 |
| 2010 | 7.32E-01 | 4.06E-01 | 1.81E-01 | 1.38E+02 | 7.68E+01 | 3.46E+01 |
| 2011 | 6.41E-01 | 3.03E-01 | 1.49E-01 | 1.21E+02 | 5.71E+01 | 2.84E+01 |
| 2012 | 5.82E-01 | 3.79E-01 | 1.75E-01 | 1.10E+02 | 7.18E+01 | 3.36E+01 |
| 2013 | 5.55E-01 | 3.40E-01 | 1.47E-01 | 1.05E+02 | 6.43E+01 | 2.81E+01 |
| 2014 | 5.27E-01 | 3.10E-01 | 1.37E-01 | 9.95E+01 | 5.86E+01 | 2.62E+01 |
| 2015 | 4.69E-01 | 2.93E-01 | 1.22E-01 | 8.85E+01 | 5.54E+01 | 2.34E+01 |
| 2016 | 4.42E-01 | 3.12E-01 | 1.68E-01 | 8.35E+01 | 5.92E+01 | 3.23E+01 |
| 2017 | 4.57E-01 | 2.33E-01 | 1.11E-01 | 8.65E+01 | 4.40E+01 | 2.13E+01 |
| 2018 | 4.12E-01 | 2.30E-01 | 9.59E-02 | 7.78E+01 | 4.33E+01 | 1.83E+01 |

**TABLE 3-42: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE EAGLE BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water Expected | Piscivorous Fish Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|---------------------------------|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 2.53E-06 | 1.20E+01 | 1.20E+01 | 2.65E+03 |
| Stillwater (168) | 4.49E-06 | 2.14E+00 | 2.14E+00 | 4.70E+02 |
| Federal Dam (154) | 3.14E-06 | 1.55E+00 | 1.55E+00 | 3.41E+02 |
| <i>Lower River</i> | | | | |
| 143.5 | 2.43E-06 | 1.55E+00 | 1.55E+00 | 3.41E+02 |
| 137.2 | 2.43E-06 | 5.86E+00 | 5.86E+00 | 1.29E+03 |
| 122.4 | 1.11E-06 | 1.36E+00 | 1.36E+00 | 3.00E+02 |
| 113.8 | 1.11E-06 | 1.23E+00 | 1.23E+00 | 2.71E+02 |
| 100 | 1.11E-06 | 1.42E+00 | 1.42E+00 | 3.11E+02 |
| 88.9 | 7.32E-07 | 9.13E-01 | 9.13E-01 | 2.01E+02 |
| 58.7 | 7.32E-07 | 1.06E+00 | 1.06E+00 | 2.33E+02 |
| 47.3 | 7.32E-07 | 1.21E+00 | 1.21E+00 | 2.67E+02 |
| 25.8 | 7.32E-07 | 8.58E-01 | 8.58E-01 | 1.88E+02 |

**TABLE 3-43: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE EAGLE BASED ON 1993 DATA
USING SUM OF TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Piscivorous Fish 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (95% UCL) (mg/Kg) |
|----------------------------|------------------------------|--------------------------------|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 8.00E-06 | 2.33E+01 | 2.33E+01 | 5.12E+03 |
| Stillwater (168) | 1.42E-05 | 2.76E+00 | 2.76E+00 | 6.06E+02 |
| Federal Dam (154) | 6.73E-06 | 2.79E+00 | 2.79E+00 | 6.13E+02 |
| <i>Lower River</i> | | | | |
| 143.5 | 2.64E-05 | 2.79E+00 | 2.79E+00 | 6.13E+02 |
| 137.2 | 2.64E-05 | 1.39E+01 | 1.39E+01 | 3.06E+03 |
| 122.4 | 1.42E-05 | 1.88E+00 | 1.88E+00 | 4.12E+02 |
| 113.8 | 1.42E-05 | 1.73E+00 | 1.73E+00 | 3.79E+02 |
| 100 | 1.42E-05 | 4.38E+00 | 4.38E+00 | 9.62E+02 |
| 88.9 | 3.25E-06 | 1.74E+00 | 1.74E+00 | 3.81E+02 |
| 58.7 | 3.25E-06 | 1.57E+00 | 1.57E+00 | 3.45E+02 |
| 47.3 | 3.25E-06 | 3.27E+00 | 3.27E+00 | 7.18E+02 |
| 25.8 | 3.25E-06 | 1.70E+00 | 1.70E+00 | 3.74E+02 |

TABLE 3-44: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE EAGLE BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 6.68E+00 | 5.04E+00 | 2.48E+00 | 1.47E+03 | 1.11E+03 | 5.45E+02 |
| 1994 | 4.06E+00 | 3.47E+00 | 1.76E+00 | 8.92E+02 | 7.61E+02 | 3.86E+02 |
| 1995 | 4.48E+00 | 3.85E+00 | 1.90E+00 | 9.83E+02 | 8.45E+02 | 4.17E+02 |
| 1996 | 4.08E+00 | 3.09E+00 | 1.67E+00 | 8.97E+02 | 6.79E+02 | 3.66E+02 |
| 1997 | 3.18E+00 | 2.71E+00 | 1.44E+00 | 6.99E+02 | 5.95E+02 | 3.16E+02 |
| 1998 | 2.65E+00 | 2.44E+00 | 1.21E+00 | 5.83E+02 | 5.35E+02 | 2.65E+02 |
| 1999 | 2.31E+00 | 1.99E+00 | 9.20E-01 | 5.07E+02 | 4.37E+02 | 2.02E+02 |
| 2000 | 2.16E+00 | 1.88E+00 | 8.77E-01 | 4.74E+02 | 4.12E+02 | 1.93E+02 |
| 2001 | 2.01E+00 | 1.81E+00 | 8.42E-01 | 4.43E+02 | 3.97E+02 | 1.85E+02 |
| 2002 | 1.99E+00 | 1.93E+00 | 8.96E-01 | 4.37E+02 | 4.23E+02 | 1.97E+02 |
| 2003 | 1.74E+00 | 1.62E+00 | 7.64E-01 | 3.81E+02 | 3.55E+02 | 1.68E+02 |
| 2004 | 1.63E+00 | 1.58E+00 | 7.88E-01 | 3.57E+02 | 3.46E+02 | 1.73E+02 |
| 2005 | 1.44E+00 | 1.20E+00 | 5.94E-01 | 3.17E+02 | 2.65E+02 | 1.31E+02 |
| 2006 | 1.42E+00 | 1.31E+00 | 6.27E-01 | 3.12E+02 | 2.88E+02 | 1.38E+02 |
| 2007 | 1.31E+00 | 1.10E+00 | 4.86E-01 | 2.87E+02 | 2.41E+02 | 1.07E+02 |
| 2008 | 1.24E+00 | 9.90E-01 | 4.71E-01 | 2.71E+02 | 2.17E+02 | 1.03E+02 |
| 2009 | 1.25E+00 | 1.07E+00 | 5.06E-01 | 2.75E+02 | 2.35E+02 | 1.11E+02 |
| 2010 | 1.10E+00 | 9.75E-01 | 4.23E-01 | 2.41E+02 | 2.14E+02 | 9.30E+01 |
| 2011 | 9.62E-01 | 8.31E-01 | 4.11E-01 | 2.11E+02 | 1.83E+02 | 9.02E+01 |
| 2012 | 9.55E-01 | 8.42E-01 | 3.99E-01 | 2.10E+02 | 1.85E+02 | 8.77E+01 |
| 2013 | 9.18E-01 | 8.75E-01 | 3.85E-01 | 2.02E+02 | 1.92E+02 | 8.46E+01 |
| 2014 | 8.35E-01 | 8.34E-01 | 3.52E-01 | 1.83E+02 | 1.83E+02 | 7.74E+01 |
| 2015 | 7.65E-01 | 7.32E-01 | 3.25E-01 | 1.68E+02 | 1.61E+02 | 7.14E+01 |
| 2016 | 7.80E-01 | 8.21E-01 | 3.50E-01 | 1.71E+02 | 1.80E+02 | 7.69E+01 |
| 2017 | 6.72E-01 | 6.22E-01 | 2.87E-01 | 1.48E+02 | 1.37E+02 | 6.31E+01 |
| 2018 | 6.51E-01 | 5.97E-01 | 2.61E-01 | 1.43E+02 | 1.31E+02 | 5.73E+01 |

TABLE 3-45: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE EAGLE BASED ON TRI+ CONGENERS FOR PERIOD 1993 - 2018

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 8.38E+00 | 6.15E+00 | 3.09E+00 | 1.84E+03 | 1.35E+03 | 6.79E+02 |
| 1994 | 5.47E+00 | 4.17E+00 | 2.16E+00 | 1.20E+03 | 9.17E+02 | 4.76E+02 |
| 1995 | 5.99E+00 | 4.62E+00 | 2.31E+00 | 1.32E+03 | 1.02E+03 | 5.07E+02 |
| 1996 | 5.89E+00 | 3.77E+00 | 2.10E+00 | 1.29E+03 | 8.29E+02 | 4.62E+02 |
| 1997 | 4.58E+00 | 3.26E+00 | 1.80E+00 | 1.01E+03 | 7.17E+02 | 3.95E+02 |
| 1998 | 3.78E+00 | 2.96E+00 | 1.48E+00 | 8.29E+02 | 6.51E+02 | 3.26E+02 |
| 1999 | 3.42E+00 | 2.46E+00 | 1.16E+00 | 7.52E+02 | 5.40E+02 | 2.55E+02 |
| 2000 | 3.21E+00 | 2.29E+00 | 1.08E+00 | 7.04E+02 | 5.04E+02 | 2.38E+02 |
| 2001 | 2.98E+00 | 2.22E+00 | 1.05E+00 | 6.54E+02 | 4.87E+02 | 2.30E+02 |
| 2002 | 2.93E+00 | 2.36E+00 | 1.12E+00 | 6.43E+02 | 5.18E+02 | 2.47E+02 |
| 2003 | 2.59E+00 | 2.00E+00 | 9.61E-01 | 5.68E+02 | 4.39E+02 | 2.11E+02 |
| 2004 | 2.39E+00 | 1.93E+00 | 9.88E-01 | 5.25E+02 | 4.24E+02 | 2.17E+02 |
| 2005 | 2.17E+00 | 1.47E+00 | 7.44E-01 | 4.78E+02 | 3.23E+02 | 1.63E+02 |
| 2006 | 2.10E+00 | 1.61E+00 | 7.70E-01 | 4.62E+02 | 3.53E+02 | 1.69E+02 |
| 2007 | 1.97E+00 | 1.33E+00 | 5.97E-01 | 4.34E+02 | 2.93E+02 | 1.31E+02 |
| 2008 | 1.88E+00 | 1.22E+00 | 5.96E-01 | 4.14E+02 | 2.68E+02 | 1.31E+02 |
| 2009 | 1.83E+00 | 1.31E+00 | 6.38E-01 | 4.02E+02 | 2.87E+02 | 1.40E+02 |
| 2010 | 1.60E+00 | 1.20E+00 | 5.25E-01 | 3.53E+02 | 2.63E+02 | 1.15E+02 |
| 2011 | 1.46E+00 | 1.02E+00 | 5.22E-01 | 3.21E+02 | 2.24E+02 | 1.15E+02 |
| 2012 | 1.41E+00 | 1.04E+00 | 5.04E-01 | 3.10E+02 | 2.28E+02 | 1.11E+02 |
| 2013 | 1.36E+00 | 1.08E+00 | 4.88E-01 | 2.98E+02 | 2.37E+02 | 1.07E+02 |
| 2014 | 1.21E+00 | 1.02E+00 | 4.37E-01 | 2.66E+02 | 2.25E+02 | 9.60E+01 |
| 2015 | 1.12E+00 | 8.92E-01 | 4.10E-01 | 2.47E+02 | 1.96E+02 | 9.00E+01 |
| 2016 | 1.13E+00 | 1.00E+00 | 4.34E-01 | 2.48E+02 | 2.21E+02 | 9.53E+01 |
| 2017 | 1.02E+00 | 7.65E-01 | 3.78E-01 | 2.24E+02 | 1.68E+02 | 8.31E+01 |
| 2018 | 9.73E-01 | 7.36E-01 | 3.35E-01 | 2.14E+02 | 1.62E+02 | 7.36E+01 |

TABLE 3-46: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE TREE SWALLOW BASED ON 1993 DATA ON TEQ BASIS

| Location | Drinking Water Expected | Benthic Invertebrate Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Average Predicted Egg Conc (mg/Kg) |
|----------------------------|-------------------------|-------------------------------|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 1.26E-07 | 5.80E-04 | 5.80E-04 | 4.74E-03 |
| Stillwater (168) | 2.24E-07 | 1.02E-03 | 1.02E-03 | 8.29E-03 |
| Federal Dam (154) | 1.56E-07 | 2.42E-04 | 2.42E-04 | 1.98E-03 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.21E-07 | 3.37E-05 | 3.38E-05 | 2.75E-04 |
| 137.2 | 1.21E-07 | 6.64E-05 | 6.65E-05 | 5.42E-04 |
| 122.4 | 5.54E-08 | 3.67E-05 | 3.68E-05 | 3.00E-04 |
| 113.8 | 5.54E-08 | 3.78E-05 | 3.78E-05 | 3.08E-04 |
| 100 | 5.54E-08 | 1.74E-05 | 1.74E-05 | 1.42E-04 |
| 88.9 | 3.65E-08 | 8.70E-06 | 8.74E-06 | 7.11E-05 |
| 58.7 | 3.65E-08 | 2.68E-05 | 2.68E-05 | 2.19E-04 |
| 47.3 | 3.65E-08 | 3.04E-05 | 3.05E-05 | 2.49E-04 |
| 25.8 | 3.65E-08 | 9.01E-06 | 9.05E-06 | 7.36E-05 |

TABLE 3-47: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE TREE SWALLOW BASED ON 1993 DATA ON TEQ BASIS

| Location | Drinking Water 95% UCL | Benthic Invertebrate 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) | 95% UCL Predicted Egg Conc (mg/Kg) |
|----------------------------|---------------------------|---------------------------------|--|---|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 3.99E-07 | 1.01E-03 | 1.01E-03 | 8.28E-03 |
| Stillwater (168) | 7.10E-07 | 4.76E-03 | 4.76E-03 | 3.88E-02 |
| Federal Dam (154) | 3.36E-07 | 4.11E-04 | 4.12E-04 | 3.36E-03 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.32E-06 | 8.27E-05 | 8.40E-05 | 6.75E-04 |
| 137.2 | 1.32E-06 | 2.69E-04 | 2.71E-04 | 2.20E-03 |
| 122.4 | 7.10E-07 | 9.23E-05 | 9.30E-05 | 7.54E-04 |
| 113.8 | 7.10E-07 | 1.46E-04 | 1.47E-04 | 1.20E-03 |
| 100 | 7.10E-07 | 1.19E-04 | 1.19E-04 | 9.69E-04 |
| 88.9 | 1.62E-07 | 1.55E-05 | 1.56E-05 | 1.26E-04 |
| 58.7 | 1.62E-07 | 2.45E-04 | 2.45E-04 | 2.00E-03 |
| 47.3 | 1.62E-07 | 2.23E-04 | 2.24E-04 | 1.82E-03 |
| 25.8 | 1.62E-07 | 1.53E-05 | 1.54E-05 | 1.25E-04 |

TABLE 3-48: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE TREE SWALLOW FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | Total Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|---|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 6.77E-04 | 3.75E-04 | 1.65E-04 | 5.53E-03 | 3.06E-03 | 1.35E-03 |
| 1994 | 6.35E-04 | 3.57E-04 | 1.56E-04 | 5.19E-03 | 2.91E-03 | 1.27E-03 |
| 1995 | 5.94E-04 | 3.36E-04 | 1.43E-04 | 4.85E-03 | 2.74E-03 | 1.16E-03 |
| 1996 | 5.35E-04 | 3.00E-04 | 1.21E-04 | 4.37E-03 | 2.45E-03 | 9.86E-04 |
| 1997 | 4.92E-04 | 2.73E-04 | 1.05E-04 | 4.02E-03 | 2.23E-03 | 8.59E-04 |
| 1998 | 4.51E-04 | 2.47E-04 | 9.17E-05 | 3.69E-03 | 2.01E-03 | 7.48E-04 |
| 1999 | 4.08E-04 | 2.24E-04 | 8.00E-05 | 3.33E-03 | 1.83E-03 | 6.53E-04 |
| 2000 | 3.75E-04 | 2.12E-04 | 7.19E-05 | 3.06E-03 | 1.73E-03 | 5.87E-04 |
| 2001 | 3.54E-04 | 2.07E-04 | 6.77E-05 | 2.89E-03 | 1.69E-03 | 5.53E-04 |
| 2002 | 3.31E-04 | 1.97E-04 | 6.42E-05 | 2.70E-03 | 1.61E-03 | 5.24E-04 |
| 2003 | 3.09E-04 | 1.86E-04 | 6.00E-05 | 2.52E-03 | 1.52E-03 | 4.90E-04 |
| 2004 | 2.87E-04 | 1.71E-04 | 5.43E-05 | 2.34E-03 | 1.40E-03 | 4.44E-04 |
| 2005 | 2.65E-04 | 1.58E-04 | 4.95E-05 | 2.16E-03 | 1.29E-03 | 4.04E-04 |
| 2006 | 2.55E-04 | 1.50E-04 | 4.58E-05 | 2.08E-03 | 1.22E-03 | 3.73E-04 |
| 2007 | 2.37E-04 | 1.41E-04 | 4.17E-05 | 1.94E-03 | 1.15E-03 | 3.40E-04 |
| 2008 | 2.26E-04 | 1.32E-04 | 3.86E-05 | 1.84E-03 | 1.07E-03 | 3.15E-04 |
| 2009 | 2.13E-04 | 1.26E-04 | 3.73E-05 | 1.74E-03 | 1.03E-03 | 3.05E-04 |
| 2010 | 1.93E-04 | 1.18E-04 | 3.50E-05 | 1.58E-03 | 9.60E-04 | 2.86E-04 |
| 2011 | 1.75E-04 | 1.08E-04 | 3.21E-05 | 1.43E-03 | 8.86E-04 | 2.62E-04 |
| 2012 | 1.63E-04 | 1.04E-04 | 3.04E-05 | 1.33E-03 | 8.45E-04 | 2.48E-04 |
| 2013 | 1.51E-04 | 9.85E-05 | 2.86E-05 | 1.23E-03 | 8.04E-04 | 2.33E-04 |
| 2014 | 1.41E-04 | 9.23E-05 | 2.61E-05 | 1.15E-03 | 7.53E-04 | 2.13E-04 |
| 2015 | 1.31E-04 | 8.91E-05 | 2.50E-05 | 1.07E-03 | 7.28E-04 | 2.04E-04 |
| 2016 | 1.25E-04 | 8.46E-05 | 2.30E-05 | 1.02E-03 | 6.91E-04 | 1.88E-04 |
| 2017 | 1.15E-04 | 7.93E-05 | 2.05E-05 | 9.42E-04 | 6.47E-04 | 1.67E-04 |
| 2018 | 1.14E-04 | 7.73E-05 | 1.99E-05 | 9.33E-04 | 6.31E-04 | 1.63E-04 |

TABLE 3-49: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE TREE SWALLOW FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|---|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 9.91E-04 | 5.92E-04 | 1.77E-04 | 8.09E-03 | 4.83E-03 | 1.45E-03 |
| 1994 | 9.40E-04 | 5.67E-04 | 1.68E-04 | 7.68E-03 | 4.63E-03 | 1.37E-03 |
| 1995 | 8.59E-04 | 5.32E-04 | 1.54E-04 | 7.01E-03 | 4.34E-03 | 1.25E-03 |
| 1996 | 7.84E-04 | 4.73E-04 | 1.31E-04 | 6.40E-03 | 3.86E-03 | 1.07E-03 |
| 1997 | 7.17E-04 | 4.34E-04 | 1.13E-04 | 5.85E-03 | 3.54E-03 | 9.21E-04 |
| 1998 | 6.45E-04 | 3.91E-04 | 9.85E-05 | 5.27E-03 | 3.19E-03 | 8.04E-04 |
| 1999 | 6.05E-04 | 3.57E-04 | 8.63E-05 | 4.94E-03 | 2.92E-03 | 7.04E-04 |
| 2000 | 5.40E-04 | 3.35E-04 | 7.72E-05 | 4.41E-03 | 2.73E-03 | 6.30E-04 |
| 2001 | 5.16E-04 | 3.28E-04 | 7.31E-05 | 4.21E-03 | 2.68E-03 | 5.96E-04 |
| 2002 | 4.91E-04 | 3.13E-04 | 6.95E-05 | 4.01E-03 | 2.56E-03 | 5.67E-04 |
| 2003 | 4.52E-04 | 2.95E-04 | 6.48E-05 | 3.69E-03 | 2.41E-03 | 5.29E-04 |
| 2004 | 4.14E-04 | 2.70E-04 | 5.83E-05 | 3.38E-03 | 2.20E-03 | 4.76E-04 |
| 2005 | 3.89E-04 | 2.51E-04 | 5.33E-05 | 3.17E-03 | 2.05E-03 | 4.35E-04 |
| 2006 | 3.71E-04 | 2.38E-04 | 4.95E-05 | 3.03E-03 | 1.94E-03 | 4.04E-04 |
| 2007 | 3.43E-04 | 2.23E-04 | 4.49E-05 | 2.80E-03 | 1.82E-03 | 3.66E-04 |
| 2008 | 3.24E-04 | 2.09E-04 | 4.18E-05 | 2.64E-03 | 1.70E-03 | 3.41E-04 |
| 2009 | 3.05E-04 | 2.00E-04 | 4.03E-05 | 2.49E-03 | 1.63E-03 | 3.29E-04 |
| 2010 | 2.83E-04 | 1.87E-04 | 3.80E-05 | 2.31E-03 | 1.53E-03 | 3.10E-04 |
| 2011 | 2.53E-04 | 1.73E-04 | 3.47E-05 | 2.07E-03 | 1.41E-03 | 2.83E-04 |
| 2012 | 2.37E-04 | 1.64E-04 | 3.29E-05 | 1.93E-03 | 1.34E-03 | 2.68E-04 |
| 2013 | 2.25E-04 | 1.56E-04 | 3.08E-05 | 1.84E-03 | 1.27E-03 | 2.51E-04 |
| 2014 | 2.07E-04 | 1.47E-04 | 2.82E-05 | 1.69E-03 | 1.20E-03 | 2.30E-04 |
| 2015 | 1.96E-04 | 1.42E-04 | 2.70E-05 | 1.60E-03 | 1.16E-03 | 2.20E-04 |
| 2016 | 1.80E-04 | 1.34E-04 | 2.49E-05 | 1.47E-03 | 1.10E-03 | 2.03E-04 |
| 2017 | 1.71E-04 | 1.26E-04 | 2.22E-05 | 1.40E-03 | 1.03E-03 | 1.82E-04 |
| 2018 | 1.70E-04 | 1.23E-04 | 2.15E-05 | 1.39E-03 | 1.00E-03 | 1.75E-04 |

**TABLE 3-50: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS
FOR FEMALE MALLARD BASED ON 1993 DATA ON A TEQ BASIS**

| Location | Drinking Water Expected | Macrophyte Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Average Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 3.46E-08 | 1.16E-03 | 1.04E-04 | 3.80E-05 | 1.31E-03 | 2.11E-02 |
| Stillwater (168) | 6.15E-08 | 1.89E-03 | 1.94E-04 | 9.92E-05 | 2.18E-03 | 3.93E-02 |
| Federal Dam (154) | 4.30E-08 | 1.84E-03 | 4.61E-05 | 8.93E-06 | 1.89E-03 | 9.38E-03 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 3.32E-08 | 1.05E-03 | 6.43E-06 | 2.75E-06 | 1.05E-03 | 1.31E-03 |
| 137.2 | 3.32E-08 | 1.05E-03 | 1.27E-05 | 4.86E-06 | 1.06E-03 | 2.57E-03 |
| 122.4 | 1.52E-08 | 6.21E-04 | 5.90E-06 | 3.08E-06 | 6.30E-04 | 1.20E-03 |
| 113.8 | 1.52E-08 | 6.21E-04 | 5.07E-06 | 3.23E-06 | 6.29E-04 | 1.03E-03 |
| 100 | 1.52E-08 | 6.21E-04 | 2.79E-06 | 1.27E-06 | 6.25E-04 | 5.67E-04 |
| 88.9 | 1.00E-08 | 4.54E-04 | 1.40E-06 | 2.50E-06 | 4.58E-04 | 2.84E-04 |
| 58.7 | 1.00E-08 | 4.54E-04 | 3.60E-06 | 8.05E-07 | 4.59E-04 | 7.32E-04 |
| 47.3 | 1.00E-08 | 4.54E-04 | 4.89E-06 | 4.91E-06 | 4.64E-04 | 9.94E-04 |
| 25.8 | 1.00E-08 | 4.54E-04 | 1.45E-06 | 1.85E-06 | 4.58E-04 | 2.94E-04 |

**TABLE 3-51: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS
FOR FEMALE MALLARD BASED ON 1993 DATA ON A TEQ BASIS**

| Location | Drinking Water 95% UCL | Macrophyte 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (95% UCL) (mg/Kg) |
|----------------------------|------------------------------|-----------------------|------------------------------------|---------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 1.10E-07 | 4.61E-03 | 1.63E-04 | 5.56E-05 | 4.83E-03 | 3.31E-02 |
| Stillwater (168) | 1.95E-07 | 5.05E-03 | 3.37E-04 | 1.73E-04 | 5.56E-03 | 6.85E-02 |
| Federal Dam (154) | 9.22E-08 | 3.82E-03 | 8.03E-05 | 1.50E-05 | 3.91E-03 | 1.63E-02 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 3.62E-07 | 3.13E-03 | 1.12E-05 | 3.01E-06 | 3.14E-03 | 2.27E-03 |
| 137.2 | 3.62E-07 | 3.13E-03 | 2.20E-05 | 9.81E-06 | 3.16E-03 | 4.48E-03 |
| 122.4 | 1.95E-07 | 9.56E-03 | 1.48E-05 | 3.42E-06 | 9.58E-03 | 3.02E-03 |
| 113.8 | 1.95E-07 | 9.56E-03 | 8.82E-06 | 5.33E-06 | 9.57E-03 | 1.79E-03 |
| 100 | 1.95E-07 | 9.56E-03 | 1.91E-05 | 2.75E-05 | 9.61E-03 | 3.88E-03 |
| 88.9 | 4.45E-08 | 8.82E-03 | 2.49E-06 | 7.30E-06 | 8.83E-03 | 5.05E-04 |
| 58.7 | 4.45E-08 | 8.82E-03 | 6.27E-06 | 8.93E-06 | 8.84E-03 | 1.27E-03 |
| 47.3 | 4.45E-08 | 8.82E-03 | 3.59E-05 | 1.92E-05 | 8.88E-03 | 7.30E-03 |
| 25.8 | 4.45E-08 | 8.82E-03 | 2.46E-06 | 5.00E-06 | 8.83E-03 | 4.99E-04 |

TABLE 3-52: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE MALLARD ON A TEQ BASIS FOR PERIOD 1993 - 2018

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 2.55E-03 | 2.75E-03 | 5.04E-04 | 2.21E-02 | 1.22E-02 | 5.38E-03 |
| 1994 | 2.85E-03 | 3.12E-03 | 5.22E-04 | 2.08E-02 | 1.17E-02 | 5.08E-03 |
| 1995 | 3.04E-03 | 2.94E-03 | 8.17E-04 | 1.94E-02 | 1.10E-02 | 4.66E-03 |
| 1996 | 1.29E-03 | 1.39E-03 | 2.92E-04 | 1.75E-02 | 9.79E-03 | 3.94E-03 |
| 1997 | 1.37E-03 | 2.22E-03 | 4.61E-04 | 1.61E-02 | 8.93E-03 | 3.44E-03 |
| 1998 | 4.50E-04 | 7.94E-04 | 3.21E-04 | 1.47E-02 | 8.05E-03 | 2.99E-03 |
| 1999 | 5.87E-04 | 1.02E-03 | 3.28E-04 | 1.33E-02 | 7.33E-03 | 2.61E-03 |
| 2000 | 4.51E-04 | 7.36E-04 | 3.15E-04 | 1.23E-02 | 6.91E-03 | 2.35E-03 |
| 2001 | 6.11E-04 | 1.27E-03 | 2.38E-04 | 1.16E-02 | 6.74E-03 | 2.21E-03 |
| 2002 | 5.19E-04 | 9.30E-04 | 1.76E-04 | 1.08E-02 | 6.44E-03 | 2.10E-03 |
| 2003 | 4.38E-04 | 8.64E-04 | 2.93E-04 | 1.01E-02 | 6.08E-03 | 1.96E-03 |
| 2004 | 3.95E-04 | 9.94E-04 | 2.42E-04 | 9.38E-03 | 5.58E-03 | 1.77E-03 |
| 2005 | 3.74E-04 | 5.57E-04 | 2.62E-04 | 8.66E-03 | 5.17E-03 | 1.61E-03 |
| 2006 | 4.53E-04 | 7.98E-04 | 1.89E-04 | 8.33E-03 | 4.89E-03 | 1.49E-03 |
| 2007 | 3.39E-04 | 5.65E-04 | 2.61E-04 | 7.76E-03 | 4.59E-03 | 1.36E-03 |
| 2008 | 2.86E-04 | 4.01E-04 | 1.35E-04 | 7.38E-03 | 4.30E-03 | 1.26E-03 |
| 2009 | 4.46E-04 | 6.82E-04 | 1.71E-04 | 6.96E-03 | 4.12E-03 | 1.22E-03 |
| 2010 | 3.28E-04 | 5.33E-04 | 2.31E-04 | 6.30E-03 | 3.84E-03 | 1.14E-03 |
| 2011 | 2.20E-04 | 3.15E-04 | 1.68E-04 | 5.71E-03 | 3.54E-03 | 1.05E-03 |
| 2012 | 3.45E-04 | 5.56E-04 | 2.86E-04 | 5.32E-03 | 3.38E-03 | 9.93E-04 |
| 2013 | 2.31E-04 | 4.33E-04 | 2.36E-04 | 4.92E-03 | 3.22E-03 | 9.32E-04 |
| 2014 | 2.47E-04 | 4.01E-04 | 2.56E-04 | 4.59E-03 | 3.01E-03 | 8.53E-04 |
| 2015 | 2.04E-04 | 3.48E-04 | 1.84E-04 | 4.29E-03 | 2.91E-03 | 8.16E-04 |
| 2016 | 2.70E-04 | 4.39E-04 | 2.56E-04 | 4.07E-03 | 2.76E-03 | 7.50E-04 |
| 2017 | 1.50E-04 | 2.06E-04 | 1.31E-04 | 3.77E-03 | 2.59E-03 | 6.69E-04 |
| 2018 | 2.62E-04 | 5.12E-04 | 3.10E-04 | 3.73E-03 | 2.53E-03 | 6.51E-04 |

TABLE 3-53: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE MALLARD ON A TEQ BASIS FOR PERIOD 1993 - 2018

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 2.65E-03 | 2.84E-03 | 5.17E-04 | 3.24E-02 | 1.93E-02 |
| 1994 | 2.96E-03 | 3.22E-03 | 5.34E-04 | 3.07E-02 | 1.85E-02 | 5.49E-03 |
| 1995 | 3.15E-03 | 3.03E-03 | 8.36E-04 | 2.81E-02 | 1.74E-02 | 5.02E-03 |
| 1996 | 1.36E-03 | 1.45E-03 | 3.00E-04 | 2.56E-02 | 1.55E-02 | 4.28E-03 |
| 1997 | 1.43E-03 | 2.29E-03 | 4.72E-04 | 2.34E-02 | 1.42E-02 | 3.69E-03 |
| 1998 | 4.91E-04 | 8.32E-04 | 3.29E-04 | 2.11E-02 | 1.28E-02 | 3.22E-03 |
| 1999 | 6.31E-04 | 1.06E-03 | 3.36E-04 | 1.97E-02 | 1.17E-02 | 2.82E-03 |
| 2000 | 4.88E-04 | 7.69E-04 | 3.23E-04 | 1.76E-02 | 1.09E-02 | 2.52E-03 |
| 2001 | 6.50E-04 | 1.32E-03 | 2.44E-04 | 1.68E-02 | 1.07E-02 | 2.39E-03 |
| 2002 | 5.56E-04 | 9.66E-04 | 1.80E-04 | 1.60E-02 | 1.02E-02 | 2.27E-03 |
| 2003 | 4.70E-04 | 8.98E-04 | 3.00E-04 | 1.48E-02 | 9.65E-03 | 2.12E-03 |
| 2004 | 4.24E-04 | 1.03E-03 | 2.47E-04 | 1.35E-02 | 8.82E-03 | 1.90E-03 |
| 2005 | 4.01E-04 | 5.83E-04 | 2.68E-04 | 1.27E-02 | 8.20E-03 | 1.74E-03 |
| 2006 | 4.82E-04 | 8.28E-04 | 1.94E-04 | 1.21E-02 | 7.78E-03 | 1.61E-03 |
| 2007 | 3.63E-04 | 5.89E-04 | 2.67E-04 | 1.12E-02 | 7.28E-03 | 1.47E-03 |
| 2008 | 3.08E-04 | 4.21E-04 | 1.38E-04 | 1.06E-02 | 6.81E-03 | 1.37E-03 |
| 2009 | 4.70E-04 | 7.07E-04 | 1.76E-04 | 9.97E-03 | 6.53E-03 | 1.32E-03 |
| 2010 | 3.50E-04 | 5.54E-04 | 2.36E-04 | 9.25E-03 | 6.12E-03 | 1.24E-03 |
| 2011 | 2.38E-04 | 3.31E-04 | 1.72E-04 | 8.28E-03 | 5.66E-03 | 1.13E-03 |
| 2012 | 3.64E-04 | 5.77E-04 | 2.92E-04 | 7.73E-03 | 5.36E-03 | 1.07E-03 |
| 2013 | 2.48E-04 | 4.51E-04 | 2.41E-04 | 7.35E-03 | 5.10E-03 | 1.00E-03 |
| 2014 | 2.62E-04 | 4.18E-04 | 2.62E-04 | 6.76E-03 | 4.80E-03 | 9.19E-04 |
| 2015 | 2.18E-04 | 3.63E-04 | 1.88E-04 | 6.40E-03 | 4.63E-03 | 8.81E-04 |
| 2016 | 2.85E-04 | 4.56E-04 | 2.62E-04 | 5.88E-03 | 4.39E-03 | 8.12E-04 |
| 2017 | 1.62E-04 | 2.17E-04 | 1.34E-04 | 5.59E-03 | 4.11E-03 | 7.26E-04 |
| 2018 | 2.76E-04 | 5.30E-04 | 3.17E-04 | 5.56E-03 | 4.01E-03 | 7.02E-04 |

**TABLE 3-54: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE BELTED KINGFISHER BASED ON 1993 DATA ON TEQ BASIS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 6.54E-08 | 1.64E-03 | 5.82E-05 | 3.75E-05 | 1.73E-03 | 8.16E-02 |
| Stillwater (168) | 1.16E-07 | 5.52E-04 | 1.02E-04 | 9.79E-05 | 7.52E-04 | 3.15E-02 |
| Federal Dam (154) | 8.13E-08 | 1.30E-04 | 2.43E-05 | 8.82E-06 | 1.63E-04 | 7.41E-03 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 6.29E-08 | 1.51E-04 | 3.38E-06 | 2.72E-06 | 1.57E-04 | 7.42E-03 |
| 137.2 | 6.29E-08 | 3.05E-04 | 6.66E-06 | 4.79E-06 | 3.16E-04 | 1.50E-02 |
| 122.4 | 2.88E-08 | 1.16E-04 | 3.68E-06 | 3.04E-06 | 1.23E-04 | 5.78E-03 |
| 113.8 | 2.88E-08 | 1.22E-04 | 3.79E-06 | 3.18E-06 | 1.29E-04 | 6.06E-03 |
| 100 | 2.88E-08 | 5.29E-05 | 1.74E-06 | 1.26E-06 | 5.59E-05 | 2.63E-03 |
| 88.9 | 1.90E-08 | 1.05E-04 | 8.73E-07 | 2.46E-06 | 1.09E-04 | 5.11E-03 |
| 58.7 | 1.90E-08 | 1.15E-04 | 2.69E-06 | 7.95E-07 | 1.18E-04 | 5.65E-03 |
| 47.3 | 1.90E-08 | 1.02E-04 | 3.05E-06 | 4.85E-06 | 1.10E-04 | 5.06E-03 |
| 25.8 | 1.90E-08 | 7.67E-05 | 9.03E-07 | 1.83E-06 | 7.94E-05 | 3.74E-03 |

TABLE 3-55: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE BELTED KINGFISHER BASED ON 1993 DATA ON TEQ BASIS

| Location | Drinking | Fish | Benthic | Sediment | Total Upper Bound | Total |
|----------------------------|----------|----------|--------------|----------|------------------------------|----------------------------|
| | Water | | Invertebrate | | Daily Dose _{95%UCL} | Concentration in |
| | 95% UCL | 95% UCL | 95% UCL | 95% UCL | (mg/Kg/day) | Eggs (95 % UCL) (mg/Kg) |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 2.07E-07 | 3.34E-03 | 1.02E-04 | 5.49E-05 | 3.50E-03 | 1.66E-01 |
| Stillwater (168) | 3.69E-07 | 7.90E-04 | 4.77E-04 | 1.71E-04 | 1.44E-03 | 6.10E-02 |
| Federal Dam (154) | 1.74E-07 | 1.88E-04 | 4.12E-05 | 1.48E-05 | 2.44E-04 | 1.10E-02 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 6.85E-07 | 1.81E-04 | 8.29E-06 | 2.97E-06 | 1.93E-04 | 9.11E-03 |
| 137.2 | 6.85E-07 | 6.61E-04 | 2.70E-05 | 9.69E-06 | 6.98E-04 | 3.31E-02 |
| 122.4 | 3.69E-07 | 1.88E-04 | 9.26E-06 | 3.38E-06 | 2.01E-04 | 9.51E-03 |
| 113.8 | 3.69E-07 | 1.27E-04 | 1.47E-05 | 5.26E-06 | 1.47E-04 | 6.80E-03 |
| 100 | 3.69E-07 | 9.12E-05 | 1.19E-05 | 2.72E-05 | 1.31E-04 | 4.97E-03 |
| 88.9 | 8.43E-08 | 1.44E-04 | 1.55E-06 | 7.21E-06 | 1.53E-04 | 7.02E-03 |
| 58.7 | 8.43E-08 | 1.30E-04 | 2.46E-05 | 8.82E-06 | 1.63E-04 | 7.44E-03 |
| 47.3 | 8.43E-08 | 1.35E-04 | 2.24E-05 | 1.89E-05 | 1.77E-04 | 7.59E-03 |
| 25.8 | 8.43E-08 | 9.22E-05 | 1.53E-06 | 4.93E-06 | 9.87E-05 | 4.51E-03 |

TABLE 3-56: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE BELTED KINGFISHER FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 2.06E-03 | 9.55E-04 | 4.74E-04 | 8.40E-02 | 4.11E-02 | 2.06E-02 |
| 1994 | 1.13E-03 | 7.86E-04 | 3.65E-04 | 4.05E-02 | 3.32E-02 | 1.56E-02 |
| 1995 | 1.05E-03 | 8.58E-04 | 4.25E-04 | 3.74E-02 | 3.68E-02 | 1.85E-02 |
| 1996 | 1.05E-03 | 5.89E-04 | 2.83E-04 | 3.87E-02 | 2.43E-02 | 1.20E-02 |
| 1997 | 8.22E-04 | 5.21E-04 | 2.60E-04 | 2.85E-02 | 2.14E-02 | 1.11E-02 |
| 1998 | 7.13E-04 | 4.64E-04 | 2.12E-04 | 2.43E-02 | 1.91E-02 | 8.96E-03 |
| 1999 | 6.21E-04 | 3.76E-04 | 1.59E-04 | 2.08E-02 | 1.51E-02 | 6.58E-03 |
| 2000 | 5.22E-04 | 3.47E-04 | 1.40E-04 | 1.68E-02 | 1.40E-02 | 5.78E-03 |
| 2001 | 5.06E-04 | 3.67E-04 | 1.59E-04 | 1.65E-02 | 1.50E-02 | 6.77E-03 |
| 2002 | 4.73E-04 | 3.54E-04 | 1.46E-04 | 1.53E-02 | 1.45E-02 | 6.18E-03 |
| 2003 | 4.51E-04 | 3.01E-04 | 1.30E-04 | 1.48E-02 | 1.20E-02 | 5.47E-03 |
| 2004 | 3.99E-04 | 2.88E-04 | 1.28E-04 | 1.29E-02 | 1.16E-02 | 5.42E-03 |
| 2005 | 3.82E-04 | 2.39E-04 | 1.04E-04 | 1.25E-02 | 9.46E-03 | 4.37E-03 |
| 2006 | 3.40E-04 | 2.48E-04 | 1.03E-04 | 1.07E-02 | 9.99E-03 | 4.35E-03 |
| 2007 | 3.34E-04 | 2.26E-04 | 8.63E-05 | 1.08E-02 | 9.06E-03 | 3.61E-03 |
| 2008 | 3.05E-04 | 2.01E-04 | 7.66E-05 | 9.74E-03 | 7.95E-03 | 3.19E-03 |
| 2009 | 2.93E-04 | 2.14E-04 | 8.99E-05 | 9.39E-03 | 8.69E-03 | 3.85E-03 |
| 2010 | 2.83E-04 | 2.14E-04 | 8.23E-05 | 9.22E-03 | 8.72E-03 | 3.50E-03 |
| 2011 | 2.63E-04 | 1.65E-04 | 6.74E-05 | 8.79E-03 | 6.55E-03 | 2.83E-03 |
| 2012 | 2.32E-04 | 1.92E-04 | 7.91E-05 | 7.59E-03 | 7.91E-03 | 3.42E-03 |
| 2013 | 2.23E-04 | 1.80E-04 | 6.71E-05 | 7.33E-03 | 7.35E-03 | 2.85E-03 |
| 2014 | 2.10E-04 | 1.62E-04 | 6.20E-05 | 6.99E-03 | 6.58E-03 | 2.64E-03 |
| 2015 | 1.92E-04 | 1.56E-04 | 5.77E-05 | 6.30E-03 | 6.39E-03 | 2.46E-03 |
| 2016 | 1.83E-04 | 1.55E-04 | 6.71E-05 | 6.00E-03 | 6.33E-03 | 2.92E-03 |
| 2017 | 1.72E-04 | 1.25E-04 | 4.78E-05 | 5.69E-03 | 4.99E-03 | 2.03E-03 |
| 2018 | 1.60E-04 | 1.23E-04 | 4.42E-05 | 5.26E-03 | 4.96E-03 | 1.88E-03 |

TABLE 3-57: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE BELTED KINGFISHER FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 3.38E-03 | 1.21E-03 | 8.43E-04 | 1.47E-01 | 5.34E-02 |
| 1994 | 1.71E-03 | 9.37E-04 | 5.67E-04 | 6.77E-02 | 4.05E-02 | 2.53E-02 |
| 1995 | 1.55E-03 | 1.09E-03 | 7.40E-04 | 6.09E-02 | 4.80E-02 | 3.37E-02 |
| 1996 | 1.62E-03 | 7.04E-04 | 5.03E-04 | 6.54E-02 | 2.99E-02 | 2.26E-02 |
| 1997 | 1.37E-03 | 6.14E-04 | 4.12E-04 | 5.43E-02 | 2.59E-02 | 1.84E-02 |
| 1998 | 1.11E-03 | 5.63E-04 | 3.49E-04 | 4.28E-02 | 2.38E-02 | 1.56E-02 |
| 1999 | 1.01E-03 | 4.43E-04 | 2.41E-04 | 3.92E-02 | 1.83E-02 | 1.05E-02 |
| 2000 | 9.00E-04 | 4.12E-04 | 2.28E-04 | 3.45E-02 | 1.71E-02 | 1.00E-02 |
| 2001 | 8.38E-04 | 4.35E-04 | 2.64E-04 | 3.20E-02 | 1.83E-02 | 1.18E-02 |
| 2002 | 7.96E-04 | 4.16E-04 | 2.34E-04 | 3.05E-02 | 1.75E-02 | 1.04E-02 |
| 2003 | 7.54E-04 | 3.53E-04 | 2.20E-04 | 2.90E-02 | 1.46E-02 | 9.78E-03 |
| 2004 | 6.95E-04 | 3.35E-04 | 2.20E-04 | 2.68E-02 | 1.38E-02 | 9.86E-03 |
| 2005 | 6.49E-04 | 2.81E-04 | 1.70E-04 | 2.51E-02 | 1.15E-02 | 7.55E-03 |
| 2006 | 5.79E-04 | 2.92E-04 | 1.75E-04 | 2.19E-02 | 1.21E-02 | 7.83E-03 |
| 2007 | 5.71E-04 | 2.67E-04 | 1.47E-04 | 2.19E-02 | 1.10E-02 | 6.54E-03 |
| 2008 | 5.09E-04 | 2.35E-04 | 1.26E-04 | 1.93E-02 | 9.59E-03 | 5.56E-03 |
| 2009 | 5.04E-04 | 2.55E-04 | 1.62E-04 | 1.93E-02 | 1.06E-02 | 7.27E-03 |
| 2010 | 4.85E-04 | 2.52E-04 | 1.39E-04 | 1.87E-02 | 1.05E-02 | 6.23E-03 |
| 2011 | 4.26E-04 | 1.96E-04 | 1.18E-04 | 1.64E-02 | 8.00E-03 | 5.27E-03 |
| 2012 | 3.89E-04 | 2.31E-04 | 1.32E-04 | 1.49E-02 | 9.80E-03 | 5.96E-03 |
| 2013 | 3.70E-04 | 2.10E-04 | 1.12E-04 | 1.42E-02 | 8.83E-03 | 5.03E-03 |
| 2014 | 3.49E-04 | 1.92E-04 | 1.05E-04 | 1.35E-02 | 8.07E-03 | 4.73E-03 |
| 2015 | 3.15E-04 | 1.82E-04 | 9.05E-05 | 1.21E-02 | 7.64E-03 | 4.04E-03 |
| 2016 | 2.97E-04 | 1.91E-04 | 1.41E-04 | 1.13E-02 | 8.08E-03 | 6.47E-03 |
| 2017 | 2.99E-04 | 1.49E-04 | 9.05E-05 | 1.17E-02 | 6.13E-03 | 4.09E-03 |
| 2018 | 2.73E-04 | 1.45E-04 | 7.30E-05 | 1.06E-02 | 6.03E-03 | 3.27E-03 |

**TABLE 3-58: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE GREAT BLUE HERON BASED ON 1993 DATA ON TEQ BASIS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|--|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 2.73E-08 | 8.32E-04 | 1.07E-06 | 2.91E-05 | 8.62E-04 | 1.02E-01 |
| Stillwater (168) | 4.85E-08 | 2.81E-04 | 1.87E-06 | 7.61E-05 | 3.59E-04 | 3.44E-02 |
| Federal Dam (154) | 3.39E-08 | 6.59E-05 | 4.46E-07 | 6.85E-06 | 7.32E-05 | 8.08E-03 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 2.62E-08 | 7.66E-05 | 6.22E-08 | 2.11E-06 | 7.88E-05 | 9.40E-03 |
| 137.2 | 2.62E-08 | 1.55E-04 | 1.22E-07 | 3.73E-06 | 1.59E-04 | 1.90E-02 |
| 122.4 | 1.20E-08 | 5.92E-05 | 6.77E-08 | 2.36E-06 | 6.16E-05 | 7.25E-03 |
| 113.8 | 1.20E-08 | 6.20E-05 | 6.97E-08 | 2.48E-06 | 6.46E-05 | 7.61E-03 |
| 100 | 1.20E-08 | 2.69E-05 | 3.20E-08 | 9.79E-07 | 2.79E-05 | 3.30E-03 |
| 88.9 | 7.91E-09 | 5.35E-05 | 1.61E-08 | 1.92E-06 | 5.54E-05 | 6.56E-03 |
| 58.7 | 7.91E-09 | 5.83E-05 | 4.95E-08 | 6.18E-07 | 5.90E-05 | 7.15E-03 |
| 47.3 | 7.91E-09 | 5.19E-05 | 5.61E-08 | 3.77E-06 | 5.57E-05 | 6.36E-03 |
| 25.8 | 7.91E-09 | 3.90E-05 | 1.66E-08 | 1.42E-06 | 4.05E-05 | 4.78E-03 |

TABLE 3-59: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE GREAT BLUE HERON BASED ON 1993 DATA ON TEQ BASIS

| Location | Drinking Water 95% UCL | Forage Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) | Total Concentration in Eggs (mg/Kg) |
|----------------------------|---------------------------|------------------------|---------------------------------|---------------------|---|---|
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 8.64E-08 | 1.70E-03 | 1.87E-06 | 4.26E-05 | 1.74E-03 | 2.08E-01 |
| Stillwater (168) | 1.54E-07 | 4.02E-04 | 8.77E-06 | 1.33E-04 | 5.43E-04 | 4.92E-02 |
| Federal Dam (154) | 7.27E-08 | 9.57E-05 | 7.59E-07 | 1.15E-05 | 1.08E-04 | 1.17E-02 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 2.85E-07 | 9.20E-05 | 1.53E-07 | 2.31E-06 | 9.48E-05 | 1.13E-02 |
| 137.2 | 2.85E-07 | 3.36E-04 | 4.97E-07 | 7.53E-06 | 3.45E-04 | 4.12E-02 |
| 122.4 | 1.54E-07 | 9.57E-05 | 1.70E-07 | 2.62E-06 | 9.87E-05 | 1.17E-02 |
| 113.8 | 1.54E-07 | 6.44E-05 | 2.70E-07 | 4.09E-06 | 6.89E-05 | 7.89E-03 |
| 100 | 1.54E-07 | 4.64E-05 | 2.19E-07 | 2.11E-05 | 6.79E-05 | 5.69E-03 |
| 88.9 | 3.51E-08 | 7.34E-05 | 2.85E-08 | 5.60E-06 | 7.90E-05 | 8.99E-03 |
| 58.7 | 3.51E-08 | 6.61E-05 | 4.53E-07 | 6.86E-06 | 7.34E-05 | 8.10E-03 |
| 47.3 | 3.51E-08 | 6.87E-05 | 4.12E-07 | 1.47E-05 | 8.39E-05 | 8.43E-03 |
| 25.8 | 3.51E-08 | 4.69E-05 | 2.82E-08 | 3.84E-06 | 5.08E-05 | 5.75E-03 |

TABLE 3-60: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR FEMALE GREAT BLUE HERON FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 9.25E-04 | 4.38E-04 | 2.20E-04 | 1.05E-01 | 5.09E-02 | 2.57E-02 |
| 1994 | 4.62E-04 | 3.55E-04 | 1.66E-04 | 4.84E-02 | 4.08E-02 | 1.92E-02 |
| 1995 | 4.27E-04 | 3.93E-04 | 1.98E-04 | 4.47E-02 | 4.56E-02 | 2.31E-02 |
| 1996 | 4.38E-04 | 2.61E-04 | 1.28E-04 | 4.68E-02 | 2.97E-02 | 1.47E-02 |
| 1997 | 3.28E-04 | 2.30E-04 | 1.19E-04 | 3.38E-02 | 2.61E-02 | 1.37E-02 |
| 1998 | 2.81E-04 | 2.05E-04 | 9.59E-05 | 2.87E-02 | 2.32E-02 | 1.10E-02 |
| 1999 | 2.42E-04 | 1.62E-04 | 7.06E-05 | 2.43E-02 | 1.82E-02 | 8.02E-03 |
| 2000 | 1.98E-04 | 1.50E-04 | 6.19E-05 | 1.94E-02 | 1.68E-02 | 7.04E-03 |
| 2001 | 1.93E-04 | 1.61E-04 | 7.22E-05 | 1.91E-02 | 1.81E-02 | 8.34E-03 |
| 2002 | 1.80E-04 | 1.55E-04 | 6.61E-05 | 1.78E-02 | 1.75E-02 | 7.61E-03 |
| 2003 | 1.73E-04 | 1.30E-04 | 5.85E-05 | 1.72E-02 | 1.44E-02 | 6.71E-03 |
| 2004 | 1.51E-04 | 1.25E-04 | 5.80E-05 | 1.49E-02 | 1.40E-02 | 6.69E-03 |
| 2005 | 1.47E-04 | 1.02E-04 | 4.67E-05 | 1.46E-02 | 1.13E-02 | 5.34E-03 |
| 2006 | 1.27E-04 | 1.07E-04 | 4.65E-05 | 1.23E-02 | 1.20E-02 | 5.34E-03 |
| 2007 | 1.27E-04 | 9.73E-05 | 3.86E-05 | 1.25E-02 | 1.09E-02 | 4.41E-03 |
| 2008 | 1.15E-04 | 8.55E-05 | 3.41E-05 | 1.12E-02 | 9.47E-03 | 3.88E-03 |
| 2009 | 1.11E-04 | 9.32E-05 | 4.11E-05 | 1.08E-02 | 1.05E-02 | 4.75E-03 |
| 2010 | 1.08E-04 | 9.36E-05 | 3.74E-05 | 1.07E-02 | 1.06E-02 | 4.31E-03 |
| 2011 | 1.02E-04 | 7.04E-05 | 3.02E-05 | 1.03E-02 | 7.80E-03 | 3.46E-03 |
| 2012 | 8.88E-05 | 8.47E-05 | 3.64E-05 | 8.81E-03 | 9.60E-03 | 4.23E-03 |
| 2013 | 8.58E-05 | 7.88E-05 | 3.05E-05 | 8.55E-03 | 8.90E-03 | 3.51E-03 |
| 2014 | 8.14E-05 | 7.06E-05 | 2.82E-05 | 8.17E-03 | 7.95E-03 | 3.26E-03 |
| 2015 | 7.38E-05 | 6.84E-05 | 2.62E-05 | 7.34E-03 | 7.71E-03 | 3.03E-03 |
| 2016 | 7.03E-05 | 6.80E-05 | 3.11E-05 | 6.99E-03 | 7.67E-03 | 3.63E-03 |
| 2017 | 6.64E-05 | 5.35E-05 | 2.17E-05 | 6.65E-03 | 5.96E-03 | 2.51E-03 |
| 2018 | 6.12E-05 | 5.31E-05 | 2.00E-05 | 6.10E-03 | 5.95E-03 | 2.31E-03 |

TABLE 3-61: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR FEMALE GREAT BLUE HERON FOR THE PERIOD 1993 - 2018 ON TEQ BASIS

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 1.57E-03 | 5.58E-04 | 2.89E-04 | 1.84E-01 | 6.55E-02 |
| 1994 | 7.38E-04 | 4.21E-04 | 2.05E-04 | 8.17E-02 | 4.89E-02 | 2.40E-02 |
| 1995 | 6.66E-04 | 5.02E-04 | 2.57E-04 | 7.35E-02 | 5.89E-02 | 3.04E-02 |
| 1996 | 7.11E-04 | 3.11E-04 | 1.69E-04 | 7.98E-02 | 3.57E-02 | 1.98E-02 |
| 1997 | 5.93E-04 | 2.69E-04 | 1.48E-04 | 6.58E-02 | 3.09E-02 | 1.73E-02 |
| 1998 | 4.69E-04 | 2.48E-04 | 1.22E-04 | 5.14E-02 | 2.84E-02 | 1.42E-02 |
| 1999 | 4.29E-04 | 1.90E-04 | 8.65E-05 | 4.69E-02 | 2.15E-02 | 9.96E-03 |
| 2000 | 3.79E-04 | 1.77E-04 | 7.87E-05 | 4.14E-02 | 2.00E-02 | 9.09E-03 |
| 2001 | 3.52E-04 | 1.89E-04 | 9.22E-05 | 3.82E-02 | 2.16E-02 | 1.08E-02 |
| 2002 | 3.34E-04 | 1.81E-04 | 8.29E-05 | 3.64E-02 | 2.07E-02 | 9.66E-03 |
| 2003 | 3.18E-04 | 1.51E-04 | 7.54E-05 | 3.47E-02 | 1.70E-02 | 8.78E-03 |
| 2004 | 2.93E-04 | 1.44E-04 | 7.54E-05 | 3.21E-02 | 1.62E-02 | 8.81E-03 |
| 2005 | 2.74E-04 | 1.18E-04 | 5.93E-05 | 3.00E-02 | 1.33E-02 | 6.88E-03 |
| 2006 | 2.41E-04 | 1.25E-04 | 6.01E-05 | 2.61E-02 | 1.42E-02 | 7.01E-03 |
| 2007 | 2.40E-04 | 1.14E-04 | 5.01E-05 | 2.62E-02 | 1.28E-02 | 5.81E-03 |
| 2008 | 2.12E-04 | 9.90E-05 | 4.35E-05 | 2.30E-02 | 1.11E-02 | 5.03E-03 |
| 2009 | 2.12E-04 | 1.10E-04 | 5.45E-05 | 2.31E-02 | 1.25E-02 | 6.38E-03 |
| 2010 | 2.05E-04 | 1.09E-04 | 4.81E-05 | 2.25E-02 | 1.25E-02 | 5.63E-03 |
| 2011 | 1.80E-04 | 8.27E-05 | 3.97E-05 | 1.97E-02 | 9.28E-03 | 4.62E-03 |
| 2012 | 1.64E-04 | 1.02E-04 | 4.64E-05 | 1.79E-02 | 1.17E-02 | 5.46E-03 |
| 2013 | 1.56E-04 | 9.17E-05 | 3.90E-05 | 1.70E-02 | 1.05E-02 | 4.56E-03 |
| 2014 | 1.48E-04 | 8.36E-05 | 3.64E-05 | 1.62E-02 | 9.53E-03 | 4.26E-03 |
| 2015 | 1.32E-04 | 7.91E-05 | 3.25E-05 | 1.44E-02 | 9.01E-03 | 3.80E-03 |
| 2016 | 1.25E-04 | 8.39E-05 | 4.43E-05 | 1.36E-02 | 9.62E-03 | 5.25E-03 |
| 2017 | 1.28E-04 | 6.33E-05 | 2.96E-05 | 1.41E-02 | 7.15E-03 | 3.47E-03 |
| 2018 | 1.15E-04 | 6.22E-05 | 2.55E-05 | 1.26E-02 | 7.04E-03 | 2.98E-03 |

**TABLE 3-62: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE EAGLE BASED ON 1993 DATA ON TEQ BASIS**

| Location | Drinking Water Expected | Piscivorous Fish Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) | Average Concentration in Eggs (mg/Kg) |
|----------------------------|-------------------------------|---------------------------------|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 2.06E-08 | 3.09E-03 | 3.09E-03 | 4.61E-01 |
| Stillwater (168) | 3.67E-08 | 5.49E-04 | 5.49E-04 | 8.19E-02 |
| Federal Dam (154) | 2.56E-08 | 3.98E-04 | 3.99E-04 | 5.94E-02 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.98E-08 | 3.98E-04 | 3.98E-04 | 5.94E-02 |
| 137.2 | 1.98E-08 | 1.50E-03 | 1.50E-03 | 2.24E-01 |
| 122.4 | 9.08E-09 | 3.50E-04 | 3.50E-04 | 5.22E-02 |
| 113.8 | 9.08E-09 | 3.17E-04 | 3.17E-04 | 4.72E-02 |
| 100 | 9.08E-09 | 3.64E-04 | 3.64E-04 | 5.42E-02 |
| 88.9 | 5.98E-09 | 2.34E-04 | 2.34E-04 | 3.49E-02 |
| 58.7 | 5.98E-09 | 2.72E-04 | 2.72E-04 | 4.06E-02 |
| 47.3 | 5.98E-09 | 3.12E-04 | 3.12E-04 | 4.65E-02 |
| 25.8 | 5.98E-09 | 2.20E-04 | 2.20E-04 | 3.28E-02 |

**TABLE 3-63: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR
FEMALE EAGLE BASED ON 1993 DATA ON TEQ BASIS**

| Location | Drinking Water 95% UCL | Piscivorous Fish 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) | Upper Bound Concentration in Eggs (95% UCL) (mg/Kg) |
|----------------------------|------------------------------|--------------------------------|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 6.53E-08 | 5.98E-03 | 5.98E-03 | 8.92E-01 |
| Stillwater (168) | 1.16E-07 | 7.07E-04 | 7.08E-04 | 1.05E-01 |
| Federal Dam (154) | 5.49E-08 | 7.16E-04 | 7.16E-04 | 1.07E-01 |
| <i>Lower River</i> | | | | |
| 143.5 | 2.16E-07 | 7.16E-04 | 7.16E-04 | 1.07E-01 |
| 137.2 | 2.16E-07 | 3.58E-03 | 3.58E-03 | 5.33E-01 |
| 122.4 | 1.16E-07 | 4.81E-04 | 4.82E-04 | 7.18E-02 |
| 113.8 | 1.16E-07 | 4.43E-04 | 4.43E-04 | 6.60E-02 |
| 100 | 1.16E-07 | 1.12E-03 | 1.12E-03 | 1.68E-01 |
| 88.9 | 2.66E-08 | 4.46E-04 | 4.46E-04 | 6.64E-02 |
| 58.7 | 2.66E-08 | 4.03E-04 | 4.03E-04 | 6.01E-02 |
| 47.3 | 2.66E-08 | 8.39E-04 | 8.39E-04 | 1.25E-01 |
| 25.8 | 2.66E-08 | 4.37E-04 | 4.37E-04 | 6.52E-02 |

**TABLE 3-64: SUMMARY OF ADD_{Expected} AND EGG CONCENTRATIONS FOR
FEMALE EAGLE FOR THE PERIOD 1993 - 2018 ON TEQ BASIS**

| Year | Average Dietary Dose (mg/Kg/day) | | | Average Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| 1993 | 1.71E-03 | 1.29E-03 | 6.36E-04 | 2.56E-01 | 1.93E-01 | 9.49E-02 |
| 1994 | 1.04E-03 | 8.90E-04 | 4.51E-04 | 1.55E-01 | 1.33E-01 | 6.72E-02 |
| 1995 | 1.15E-03 | 9.88E-04 | 4.87E-04 | 1.71E-01 | 1.47E-01 | 7.27E-02 |
| 1996 | 1.05E-03 | 7.94E-04 | 4.28E-04 | 1.56E-01 | 1.18E-01 | 6.38E-02 |
| 1997 | 8.16E-04 | 6.95E-04 | 3.69E-04 | 1.22E-01 | 1.04E-01 | 5.50E-02 |
| 1998 | 6.81E-04 | 6.25E-04 | 3.10E-04 | 1.02E-01 | 9.32E-02 | 4.62E-02 |
| 1999 | 5.92E-04 | 5.10E-04 | 2.36E-04 | 8.83E-02 | 7.61E-02 | 3.52E-02 |
| 2000 | 5.54E-04 | 4.82E-04 | 2.25E-04 | 8.25E-02 | 7.18E-02 | 3.36E-02 |
| 2001 | 5.17E-04 | 4.64E-04 | 2.16E-04 | 7.71E-02 | 6.91E-02 | 3.22E-02 |
| 2002 | 5.10E-04 | 4.94E-04 | 2.30E-04 | 7.60E-02 | 7.37E-02 | 3.43E-02 |
| 2003 | 4.45E-04 | 4.15E-04 | 1.96E-04 | 6.64E-02 | 6.19E-02 | 2.92E-02 |
| 2004 | 4.17E-04 | 4.04E-04 | 2.02E-04 | 6.22E-02 | 6.03E-02 | 3.02E-02 |
| 2005 | 3.70E-04 | 3.09E-04 | 1.53E-04 | 5.52E-02 | 4.61E-02 | 2.27E-02 |
| 2006 | 3.64E-04 | 3.37E-04 | 1.61E-04 | 5.43E-02 | 5.02E-02 | 2.40E-02 |
| 2007 | 3.36E-04 | 2.81E-04 | 1.25E-04 | 5.01E-02 | 4.20E-02 | 1.86E-02 |
| 2008 | 3.17E-04 | 2.54E-04 | 1.21E-04 | 4.73E-02 | 3.79E-02 | 1.80E-02 |
| 2009 | 3.21E-04 | 2.74E-04 | 1.30E-04 | 4.78E-02 | 4.09E-02 | 1.94E-02 |
| 2010 | 2.82E-04 | 2.50E-04 | 1.09E-04 | 4.20E-02 | 3.73E-02 | 1.62E-02 |
| 2011 | 2.47E-04 | 2.13E-04 | 1.05E-04 | 3.68E-02 | 3.18E-02 | 1.57E-02 |
| 2012 | 2.45E-04 | 2.16E-04 | 1.03E-04 | 3.65E-02 | 3.22E-02 | 1.53E-02 |
| 2013 | 2.36E-04 | 2.25E-04 | 9.89E-05 | 3.51E-02 | 3.35E-02 | 1.47E-02 |
| 2014 | 2.14E-04 | 2.14E-04 | 9.04E-05 | 3.19E-02 | 3.19E-02 | 1.35E-02 |
| 2015 | 1.96E-04 | 1.88E-04 | 8.34E-05 | 2.93E-02 | 2.80E-02 | 1.24E-02 |
| 2016 | 2.00E-04 | 2.11E-04 | 8.98E-05 | 2.99E-02 | 3.14E-02 | 1.34E-02 |
| 2017 | 1.72E-04 | 1.60E-04 | 7.37E-05 | 2.57E-02 | 2.38E-02 | 1.10E-02 |
| 2018 | 1.67E-04 | 1.53E-04 | 6.69E-05 | 2.49E-02 | 2.28E-02 | 9.97E-03 |

**TABLE 3-65: SUMMARY OF ADD_{95%UCL} AND EGG CONCENTRATIONS FOR
FEMALE EAGLE FOR THE PERIOD 1993 - 2018 ON TEQ BASIS**

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | | 95% UCL Egg Concentration (mg/Kg) | | |
|------|-------------------------------------|----------|----------|--------------------------------------|----------|----------|
| | 189 | 168 | 154 | 189 | 168 | 154 |
| | 1993 | 2.15E-03 | 1.58E-03 | 7.93E-04 | 3.21E-01 | 2.35E-01 |
| 1994 | 1.40E-03 | 1.07E-03 | 5.56E-04 | 2.09E-01 | 1.60E-01 | 8.28E-02 |
| 1995 | 1.54E-03 | 1.19E-03 | 5.93E-04 | 2.29E-01 | 1.77E-01 | 8.83E-02 |
| 1996 | 1.51E-03 | 9.68E-04 | 5.40E-04 | 2.25E-01 | 1.44E-01 | 8.05E-02 |
| 1997 | 1.18E-03 | 8.37E-04 | 4.62E-04 | 1.75E-01 | 1.25E-01 | 6.89E-02 |
| 1998 | 9.69E-04 | 7.60E-04 | 3.81E-04 | 1.44E-01 | 1.13E-01 | 5.67E-02 |
| 1999 | 8.79E-04 | 6.31E-04 | 2.98E-04 | 1.31E-01 | 9.41E-02 | 4.44E-02 |
| 2000 | 8.23E-04 | 5.89E-04 | 2.78E-04 | 1.23E-01 | 8.78E-02 | 4.14E-02 |
| 2001 | 7.64E-04 | 5.69E-04 | 2.69E-04 | 1.14E-01 | 8.49E-02 | 4.01E-02 |
| 2002 | 7.51E-04 | 6.05E-04 | 2.88E-04 | 1.12E-01 | 9.02E-02 | 4.30E-02 |
| 2003 | 6.64E-04 | 5.13E-04 | 2.47E-04 | 9.89E-02 | 7.64E-02 | 3.68E-02 |
| 2004 | 6.14E-04 | 4.96E-04 | 2.54E-04 | 9.15E-02 | 7.39E-02 | 3.78E-02 |
| 2005 | 5.58E-04 | 3.78E-04 | 1.91E-04 | 8.32E-02 | 5.63E-02 | 2.85E-02 |
| 2006 | 5.39E-04 | 4.13E-04 | 1.98E-04 | 8.04E-02 | 6.15E-02 | 2.94E-02 |
| 2007 | 5.07E-04 | 3.42E-04 | 1.53E-04 | 7.55E-02 | 5.10E-02 | 2.28E-02 |
| 2008 | 4.83E-04 | 3.13E-04 | 1.53E-04 | 7.20E-02 | 4.66E-02 | 2.28E-02 |
| 2009 | 4.69E-04 | 3.35E-04 | 1.64E-04 | 6.99E-02 | 5.00E-02 | 2.44E-02 |
| 2010 | 4.12E-04 | 3.07E-04 | 1.35E-04 | 6.14E-02 | 4.58E-02 | 2.01E-02 |
| 2011 | 3.75E-04 | 2.62E-04 | 1.34E-04 | 5.59E-02 | 3.91E-02 | 2.00E-02 |
| 2012 | 3.62E-04 | 2.66E-04 | 1.29E-04 | 5.40E-02 | 3.97E-02 | 1.93E-02 |
| 2013 | 3.49E-04 | 2.77E-04 | 1.25E-04 | 5.20E-02 | 4.12E-02 | 1.87E-02 |
| 2014 | 3.11E-04 | 2.62E-04 | 1.12E-04 | 4.64E-02 | 3.91E-02 | 1.67E-02 |
| 2015 | 2.88E-04 | 2.29E-04 | 1.05E-04 | 4.29E-02 | 3.41E-02 | 1.57E-02 |
| 2016 | 2.90E-04 | 2.58E-04 | 1.11E-04 | 4.32E-02 | 3.84E-02 | 1.66E-02 |
| 2017 | 2.61E-04 | 1.96E-04 | 9.71E-05 | 3.89E-02 | 2.93E-02 | 1.45E-02 |
| 2018 | 2.50E-04 | 1.89E-04 | 8.59E-05 | 3.72E-02 | 2.82E-02 | 1.28E-02 |

**TABLE 3-66: SUMMARY OF ADD_{Expected} FOR FEMALE BAT USING
1993 DATA BASED ON TRI+ CONGENERS**

| Location | Drinking Water Expected | Benthic Invertebrate Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|-------------------------------------|--|
| <i>Upper River</i> | | | |
| Thompson Island Pool (189) | 1.14E-05 | 4.47E+00 | 4.47E+00 |
| Stillwater (168) | 2.03E-05 | 7.83E+00 | 7.83E+00 |
| Federal Dam (154) | 1.42E-05 | 1.87E+00 | 1.87E+00 |
| <i>Lower River</i> | | | |
| 143.5 | 1.10E-05 | 2.60E-01 | 2.60E-01 |
| 137.2 | 1.10E-05 | 5.12E-01 | 5.12E-01 |
| 122.4 | 5.02E-06 | 2.83E-01 | 2.83E-01 |
| 113.8 | 5.02E-06 | 2.91E-01 | 2.91E-01 |
| 100 | 5.02E-06 | 1.34E-01 | 1.34E-01 |
| 88.9 | 3.31E-06 | 6.71E-02 | 6.71E-02 |
| 58.7 | 3.31E-06 | 2.07E-01 | 2.07E-01 |
| 47.3 | 3.31E-06 | 2.35E-01 | 2.35E-01 |
| 25.8 | 3.31E-06 | 6.95E-02 | 6.95E-02 |

**TABLE 3-67: SUMMARY OF ADD_{95%UCL} FOR FEMALE BAT USING
1993 DATA BASED ON TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Benthic Invertebrate 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|------------------------------------|--|
| <i>Upper River</i> | | | |
| Thompson Island Pool (189) | 3.61E-05 | 7.82E+00 | 7.82E+00 |
| Stillwater (168) | 6.43E-05 | 3.67E+01 | 3.67E+01 |
| Federal Dam (154) | 3.04E-05 | 3.17E+00 | 3.17E+00 |
| <i>Lower River</i> | | | |
| 143.5 | 1.19E-04 | 6.38E-01 | 6.38E-01 |
| 137.2 | 1.19E-04 | 2.08E+00 | 2.08E+00 |
| 122.4 | 6.43E-05 | 7.12E-01 | 7.12E-01 |
| 113.8 | 6.43E-05 | 1.13E+00 | 1.13E+00 |
| 100 | 6.43E-05 | 9.15E-01 | 9.15E-01 |
| 88.9 | 1.47E-05 | 1.19E-01 | 1.19E-01 |
| 58.7 | 1.47E-05 | 1.89E+00 | 1.89E+00 |
| 47.3 | 1.47E-05 | 1.72E+00 | 1.72E+00 |
| 25.8 | 1.47E-05 | 1.18E-01 | 1.18E-01 |

**TABLE 3-68: SUMMARY OF ADD_{Expected} FOR FEMALE BAT
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 4.47E+00 | 1.76E+00 | 7.75E-01 |
| 1994 | 5.01E+00 | 1.68E+00 | 7.32E-01 |
| 1995 | 4.64E+00 | 1.57E+00 | 6.69E-01 |
| 1996 | 4.25E+00 | 1.42E+00 | 5.71E-01 |
| 1997 | 3.88E+00 | 1.28E+00 | 4.95E-01 |
| 1998 | 3.52E+00 | 1.16E+00 | 4.30E-01 |
| 1999 | 3.21E+00 | 1.06E+00 | 3.77E-01 |
| 2000 | 2.98E+00 | 9.96E-01 | 3.37E-01 |
| 2001 | 2.82E+00 | 9.68E-01 | 3.18E-01 |
| 2002 | 2.65E+00 | 9.28E-01 | 3.03E-01 |
| 2003 | 2.45E+00 | 8.75E-01 | 2.83E-01 |
| 2004 | 2.25E+00 | 8.01E-01 | 2.55E-01 |
| 2005 | 2.11E+00 | 7.42E-01 | 2.32E-01 |
| 2006 | 2.01E+00 | 7.04E-01 | 2.15E-01 |
| 2007 | 1.88E+00 | 6.56E-01 | 1.95E-01 |
| 2008 | 1.77E+00 | 6.19E-01 | 1.82E-01 |
| 2009 | 1.67E+00 | 5.92E-01 | 1.76E-01 |
| 2010 | 1.53E+00 | 5.52E-01 | 1.65E-01 |
| 2011 | 1.37E+00 | 5.08E-01 | 1.51E-01 |
| 2012 | 1.28E+00 | 4.86E-01 | 1.43E-01 |
| 2013 | 1.20E+00 | 4.62E-01 | 1.34E-01 |
| 2014 | 1.11E+00 | 4.32E-01 | 1.22E-01 |
| 2015 | 1.05E+00 | 4.18E-01 | 1.18E-01 |
| 2016 | 9.86E-01 | 3.97E-01 | 1.09E-01 |
| 2017 | 9.27E-01 | 3.73E-01 | 9.71E-02 |
| 2018 | 9.02E-01 | 3.62E-01 | 9.35E-02 |

**TABLE 3-69: SUMMARY OF ADD_{95%UCL} FOR FEMALE BAT
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 6.74E+00 | 2.02E+00 | 9.86E-01 |
| 1994 | 6.29E+00 | 1.92E+00 | 9.31E-01 |
| 1995 | 5.83E+00 | 1.80E+00 | 8.51E-01 |
| 1996 | 5.33E+00 | 1.63E+00 | 7.26E-01 |
| 1997 | 4.87E+00 | 1.47E+00 | 6.29E-01 |
| 1998 | 4.42E+00 | 1.33E+00 | 5.47E-01 |
| 1999 | 4.03E+00 | 1.21E+00 | 4.79E-01 |
| 2000 | 3.74E+00 | 1.14E+00 | 4.29E-01 |
| 2001 | 3.54E+00 | 1.11E+00 | 4.05E-01 |
| 2002 | 3.32E+00 | 1.07E+00 | 3.86E-01 |
| 2003 | 3.07E+00 | 1.00E+00 | 3.60E-01 |
| 2004 | 2.82E+00 | 9.20E-01 | 3.25E-01 |
| 2005 | 2.65E+00 | 8.52E-01 | 2.95E-01 |
| 2006 | 2.52E+00 | 8.08E-01 | 2.74E-01 |
| 2007 | 2.36E+00 | 7.53E-01 | 2.48E-01 |
| 2008 | 2.22E+00 | 7.10E-01 | 2.32E-01 |
| 2009 | 2.10E+00 | 6.79E-01 | 2.23E-01 |
| 2010 | 1.91E+00 | 6.34E-01 | 2.09E-01 |
| 2011 | 1.72E+00 | 5.83E-01 | 1.91E-01 |
| 2012 | 1.61E+00 | 5.58E-01 | 1.82E-01 |
| 2013 | 1.50E+00 | 5.30E-01 | 1.71E-01 |
| 2014 | 1.40E+00 | 4.96E-01 | 1.55E-01 |
| 2015 | 1.32E+00 | 4.80E-01 | 1.50E-01 |
| 2016 | 1.24E+00 | 4.55E-01 | 1.38E-01 |
| 2017 | 1.16E+00 | 4.28E-01 | 1.24E-01 |
| 2018 | 1.13E+00 | 4.16E-01 | 1.19E-01 |

**TABLE 3-70: SUMMARY OF ADD_{Expected} FOR FEMALE RACCOON USING
1993 DATA BASED ON TRI+ CONGENERS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 6.05E-06 | 9.38E-02 | 7.27E-01 | 5.51E-02 | 8.76E-01 |
| Stillwater (168) | 1.08E-05 | 1.26E-02 | 1.27E+00 | 1.44E-01 | 1.43E+00 |
| Federal Dam (154) | 7.51E-06 | 7.69E-03 | 3.03E-01 | 1.30E-02 | 3.24E-01 |
| <i>Lower River</i> | | | | | |
| 143.5 | 5.81E-06 | 8.92E-03 | 4.22E-02 | 3.99E-03 | 5.52E-02 |
| 137.2 | 5.81E-06 | 1.79E-02 | 8.32E-02 | 7.05E-03 | 1.08E-01 |
| 122.4 | 2.66E-06 | 6.82E-03 | 4.60E-02 | 4.47E-03 | 5.73E-02 |
| 113.8 | 2.66E-06 | 7.24E-03 | 4.73E-02 | 4.68E-03 | 5.92E-02 |
| 100 | 2.66E-06 | 3.11E-03 | 2.18E-02 | 1.85E-03 | 2.67E-02 |
| 88.9 | 1.75E-06 | 6.18E-03 | 1.09E-02 | 3.62E-03 | 2.07E-02 |
| 58.7 | 1.75E-06 | 6.79E-03 | 3.36E-02 | 1.17E-03 | 4.16E-02 |
| 47.3 | 1.75E-06 | 6.03E-03 | 3.81E-02 | 7.14E-03 | 5.13E-02 |
| 25.8 | 1.75E-06 | 4.55E-03 | 1.13E-02 | 2.68E-03 | 1.85E-02 |

**TABLE 3-71: SUMMARY OF ADD_{95%UCL} FOR FEMALE RACCOON USING 1993 DATA
BASED ON TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|-----------------|------------------------------------|---------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 1.92E-05 | 1.43E-01 | 1.27E+00 | 8.07E-02 | 1.50E+00 |
| Stillwater (168) | 3.41E-05 | 2.32E-02 | 5.96E+00 | 2.51E-01 | 6.24E+00 |
| Federal Dam (154) | 1.61E-05 | 1.10E-02 | 5.15E-01 | 2.17E-02 | 5.48E-01 |
| <i>Lower River</i> | | | | | |
| 143.5 | 6.33E-05 | 1.06E-02 | 1.04E-01 | 4.37E-03 | 1.19E-01 |
| 137.2 | 6.33E-05 | 3.80E-02 | 3.38E-01 | 1.42E-02 | 3.90E-01 |
| 122.4 | 3.41E-05 | 1.07E-02 | 1.16E-01 | 4.96E-03 | 1.31E-01 |
| 113.8 | 3.41E-05 | 7.51E-03 | 1.83E-01 | 7.74E-03 | 1.99E-01 |
| 100 | 3.41E-05 | 5.28E-03 | 1.49E-01 | 4.00E-02 | 1.94E-01 |
| 88.9 | 7.79E-06 | 8.11E-03 | 1.94E-02 | 1.06E-02 | 3.81E-02 |
| 58.7 | 7.79E-06 | 7.65E-03 | 3.07E-01 | 1.30E-02 | 3.28E-01 |
| 47.3 | 7.79E-06 | 7.94E-03 | 2.80E-01 | 2.78E-02 | 3.16E-01 |
| 25.8 | 7.79E-06 | 5.45E-03 | 1.92E-02 | 7.25E-03 | 3.19E-02 |

**TABLE 3-72: SUMMARY OF ADD_{Expected} FOR FEMALE RACCOON
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 8.93E-01 | 3.44E-01 | 1.84E-01 |
| 1994 | 9.68E-01 | 3.27E-01 | 1.74E-01 |
| 1995 | 9.03E-01 | 3.09E-01 | 1.60E-01 |
| 1996 | 8.21E-01 | 2.76E-01 | 1.36E-01 |
| 1997 | 7.51E-01 | 2.51E-01 | 1.18E-01 |
| 1998 | 6.79E-01 | 2.24E-01 | 1.02E-01 |
| 1999 | 6.20E-01 | 2.05E-01 | 8.99E-02 |
| 2000 | 5.73E-01 | 1.92E-01 | 8.04E-02 |
| 2001 | 5.43E-01 | 1.88E-01 | 7.57E-02 |
| 2002 | 5.10E-01 | 1.80E-01 | 7.20E-02 |
| 2003 | 4.72E-01 | 1.70E-01 | 6.76E-02 |
| 2004 | 4.33E-01 | 1.56E-01 | 6.10E-02 |
| 2005 | 4.06E-01 | 1.43E-01 | 5.53E-02 |
| 2006 | 3.87E-01 | 1.37E-01 | 5.14E-02 |
| 2007 | 3.62E-01 | 1.27E-01 | 4.67E-02 |
| 2008 | 3.40E-01 | 1.19E-01 | 4.33E-02 |
| 2009 | 3.22E-01 | 1.15E-01 | 4.19E-02 |
| 2010 | 2.95E-01 | 1.07E-01 | 3.95E-02 |
| 2011 | 2.65E-01 | 9.81E-02 | 3.60E-02 |
| 2012 | 2.46E-01 | 9.44E-02 | 3.44E-02 |
| 2013 | 2.31E-01 | 8.98E-02 | 3.23E-02 |
| 2014 | 2.14E-01 | 8.39E-02 | 2.93E-02 |
| 2015 | 2.03E-01 | 8.09E-02 | 2.82E-02 |
| 2016 | 1.90E-01 | 7.73E-02 | 2.63E-02 |
| 2017 | 1.78E-01 | 7.20E-02 | 2.32E-02 |
| 2018 | 1.73E-01 | 6.99E-02 | 2.24E-02 |

**TABLE 3-73: SUMMARY OF ADD_{95%UCL} FOR FEMALE RACCOON
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose | | |
|------|----------------------------|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 1.27E+00 | 3.87E-01 | 1.85E-01 |
| 1994 | 1.18E+00 | 3.68E-01 | 1.74E-01 |
| 1995 | 1.10E+00 | 3.47E-01 | 1.61E-01 |
| 1996 | 1.00E+00 | 3.10E-01 | 1.36E-01 |
| 1997 | 9.19E-01 | 2.82E-01 | 1.18E-01 |
| 1998 | 8.32E-01 | 2.53E-01 | 1.03E-01 |
| 1999 | 7.59E-01 | 2.31E-01 | 9.01E-02 |
| 2000 | 7.02E-01 | 2.17E-01 | 8.06E-02 |
| 2001 | 6.65E-01 | 2.12E-01 | 7.58E-02 |
| 2002 | 6.25E-01 | 2.02E-01 | 7.22E-02 |
| 2003 | 5.78E-01 | 1.91E-01 | 6.77E-02 |
| 2004 | 5.31E-01 | 1.76E-01 | 6.11E-02 |
| 2005 | 4.98E-01 | 1.62E-01 | 5.54E-02 |
| 2006 | 4.74E-01 | 1.54E-01 | 5.15E-02 |
| 2007 | 4.44E-01 | 1.43E-01 | 4.68E-02 |
| 2008 | 4.17E-01 | 1.35E-01 | 4.34E-02 |
| 2009 | 3.94E-01 | 1.29E-01 | 4.23E-02 |
| 2010 | 3.61E-01 | 1.21E-01 | 3.96E-02 |
| 2011 | 3.24E-01 | 1.11E-01 | 3.61E-02 |
| 2012 | 3.02E-01 | 1.06E-01 | 3.45E-02 |
| 2013 | 2.83E-01 | 1.01E-01 | 3.23E-02 |
| 2014 | 2.63E-01 | 9.45E-02 | 2.94E-02 |
| 2015 | 2.48E-01 | 9.12E-02 | 2.82E-02 |
| 2016 | 2.33E-01 | 8.70E-02 | 2.63E-02 |
| 2017 | 2.18E-01 | 8.11E-02 | 2.33E-02 |
| 2018 | 2.12E-01 | 7.88E-02 | 2.24E-02 |

**TABLE 3-74: SUMMARY OF ADD_{Expected} FOR FEMALE MINK USING 1993 DATA
BASED ON TRI+ CONGENERS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 7.45E-06 | 1.09E+00 | 3.33E-01 | 8.44E-03 | 1.43E+00 |
| Stillwater (168) | 1.32E-05 | 1.46E-01 | 5.83E-01 | 2.21E-02 | 7.52E-01 |
| Federal Dam (154) | 9.25E-06 | 8.96E-02 | 1.39E-01 | 1.99E-03 | 2.31E-01 |
| <i>Lower River</i> | | | | | |
| 143.5 | 7.16E-06 | 1.04E-01 | 1.94E-02 | 6.12E-04 | 1.24E-01 |
| 137.2 | 7.16E-06 | 2.09E-01 | 3.81E-02 | 1.08E-03 | 2.48E-01 |
| 122.4 | 3.28E-06 | 7.94E-02 | 2.11E-02 | 6.84E-04 | 1.01E-01 |
| 113.8 | 3.28E-06 | 8.43E-02 | 2.17E-02 | 7.17E-04 | 1.07E-01 |
| 100 | 3.28E-06 | 3.63E-02 | 9.97E-03 | 2.84E-04 | 4.65E-02 |
| 88.9 | 2.16E-06 | 7.20E-02 | 5.00E-03 | 5.55E-04 | 7.75E-02 |
| 58.7 | 2.16E-06 | 7.91E-02 | 1.54E-02 | 1.79E-04 | 9.47E-02 |
| 47.3 | 2.16E-06 | 7.03E-02 | 1.75E-02 | 1.09E-03 | 8.89E-02 |
| 25.8 | 2.16E-06 | 5.30E-02 | 5.18E-03 | 4.11E-04 | 5.86E-02 |

**TABLE 3-75: SUMMARY OF ADD_{95%UCL} FOR FEMALE MINK USING 1993 DATA
BASED ON TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Forage Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|---------------------------|------------------------------------|---------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 2.36E-05 | 1.67E+00 | 5.83E-01 | 1.24E-02 | 2.26E+00 |
| Stillwater (168) | 4.20E-05 | 2.70E-01 | 2.73E+00 | 3.85E-02 | 3.04E+00 |
| Federal Dam (154) | 1.98E-05 | 1.29E-01 | 2.36E-01 | 3.33E-03 | 3.68E-01 |
| <i>Lower River</i> | | | | | |
| 143.5 | 7.80E-05 | 1.23E-01 | 4.75E-02 | 6.69E-04 | 1.71E-01 |
| 137.2 | 7.80E-05 | 4.42E-01 | 1.55E-01 | 2.18E-03 | 6.00E-01 |
| 122.4 | 4.20E-05 | 1.25E-01 | 5.30E-02 | 7.60E-04 | 1.79E-01 |
| 113.8 | 4.20E-05 | 8.75E-02 | 8.41E-02 | 1.19E-03 | 1.73E-01 |
| 100 | 4.20E-05 | 6.15E-02 | 6.82E-02 | 6.12E-03 | 1.36E-01 |
| 88.9 | 9.59E-06 | 9.46E-02 | 8.89E-03 | 1.62E-03 | 1.05E-01 |
| 58.7 | 9.59E-06 | 8.91E-02 | 1.41E-01 | 1.99E-03 | 2.32E-01 |
| 47.3 | 9.59E-06 | 9.25E-02 | 1.28E-01 | 4.26E-03 | 2.25E-01 |
| 25.8 | 9.59E-06 | 6.35E-02 | 8.78E-03 | 1.11E-03 | 7.34E-02 |

**TABLE 3-76: SUMMARY OF ADD_{Expected} FOR FEMALE MINK
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Average Dietary Dose (mg/Kg/day) | | |
|------|-------------------------------------|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 7.25E-01 | 3.16E-01 | 1.16E-01 |
| 1994 | 7.40E-01 | 2.99E-01 | 1.11E-01 |
| 1995 | 7.29E-01 | 2.89E-01 | 1.09E-01 |
| 1996 | 6.23E-01 | 2.30E-01 | 8.59E-02 |
| 1997 | 5.56E-01 | 2.28E-01 | 7.67E-02 |
| 1998 | 4.93E-01 | 1.78E-01 | 6.50E-02 |
| 1999 | 4.53E-01 | 1.69E-01 | 5.82E-02 |
| 2000 | 4.17E-01 | 1.51E-01 | 5.42E-02 |
| 2001 | 3.98E-01 | 1.58E-01 | 4.98E-02 |
| 2002 | 3.70E-01 | 1.45E-01 | 4.67E-02 |
| 2003 | 3.43E-01 | 1.37E-01 | 4.56E-02 |
| 2004 | 3.15E-01 | 1.28E-01 | 4.03E-02 |
| 2005 | 2.97E-01 | 1.14E-01 | 3.72E-02 |
| 2006 | 2.83E-01 | 1.13E-01 | 3.46E-02 |
| 2007 | 2.63E-01 | 1.01E-01 | 3.29E-02 |
| 2008 | 2.48E-01 | 9.30E-02 | 2.86E-02 |
| 2009 | 2.36E-01 | 9.47E-02 | 3.01E-02 |
| 2010 | 2.15E-01 | 8.72E-02 | 2.86E-02 |
| 2011 | 1.92E-01 | 7.65E-02 | 2.58E-02 |
| 2012 | 1.81E-01 | 8.11E-02 | 2.70E-02 |
| 2013 | 1.70E-01 | 7.38E-02 | 2.36E-02 |
| 2014 | 1.58E-01 | 7.07E-02 | 2.22E-02 |
| 2015 | 1.49E-01 | 6.65E-02 | 2.16E-02 |
| 2016 | 1.40E-01 | 6.55E-02 | 2.11E-02 |
| 2017 | 1.31E-01 | 5.67E-02 | 1.72E-02 |
| 2018 | 1.27E-01 | 5.74E-02 | 1.75E-02 |

**TABLE 3-77: SUMMARY OF ADD_{95%UCL} FOR FEMALE MINK
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|-------------------------------------|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 9.25E-01 | 3.44E-01 | 1.36E-01 |
| 1994 | 8.63E-01 | 3.26E-01 | 1.30E-01 |
| 1995 | 8.47E-01 | 3.15E-01 | 1.27E-01 |
| 1996 | 7.29E-01 | 2.53E-01 | 1.01E-01 |
| 1997 | 6.52E-01 | 2.48E-01 | 8.92E-02 |
| 1998 | 5.81E-01 | 1.97E-01 | 7.61E-02 |
| 1999 | 5.33E-01 | 1.86E-01 | 6.79E-02 |
| 2000 | 4.92E-01 | 1.66E-01 | 6.27E-02 |
| 2001 | 4.69E-01 | 1.73E-01 | 5.80E-02 |
| 2002 | 4.36E-01 | 1.60E-01 | 5.45E-02 |
| 2003 | 4.05E-01 | 1.51E-01 | 5.27E-02 |
| 2004 | 3.72E-01 | 1.40E-01 | 4.68E-02 |
| 2005 | 3.49E-01 | 1.26E-01 | 4.31E-02 |
| 2006 | 3.33E-01 | 1.24E-01 | 4.00E-02 |
| 2007 | 3.10E-01 | 1.12E-01 | 3.79E-02 |
| 2008 | 2.92E-01 | 1.03E-01 | 3.33E-02 |
| 2009 | 2.78E-01 | 1.04E-01 | 3.45E-02 |
| 2010 | 2.53E-01 | 9.61E-02 | 3.28E-02 |
| 2011 | 2.27E-01 | 8.46E-02 | 2.96E-02 |
| 2012 | 2.13E-01 | 8.88E-02 | 3.07E-02 |
| 2013 | 1.99E-01 | 8.12E-02 | 2.71E-02 |
| 2014 | 1.85E-01 | 7.75E-02 | 2.53E-02 |
| 2015 | 1.75E-01 | 7.31E-02 | 2.46E-02 |
| 2016 | 1.65E-01 | 7.17E-02 | 2.38E-02 |
| 2017 | 1.54E-01 | 6.26E-02 | 1.96E-02 |
| 2018 | 1.50E-01 | 6.31E-02 | 1.99E-02 |

**TABLE 3-78: SUMMARY OF ADD_{Expected} FOR FEMALE OTTER USING
1993 DATA BASED ON TRI+ CONGENERS**

| Location | Drinking Water Expected | Piscivorous Fish Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|---------------------------------|----------------------|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 5.97E-06 | 1.16E+01 | 5.73E-03 | 1.16E+01 |
| Stillwater (168) | 1.06E-05 | 2.06E+00 | 1.50E-02 | 2.08E+00 |
| Federal Dam (154) | 7.42E-06 | 1.50E+00 | 1.35E-03 | 1.50E+00 |
| <i>Lower River</i> | | | | |
| 143.5 | 5.74E-06 | 1.50E+00 | 4.15E-04 | 1.50E+00 |
| 137.2 | 5.74E-06 | 5.65E+00 | 7.32E-04 | 5.65E+00 |
| 122.4 | 2.63E-06 | 1.32E+00 | 4.64E-04 | 1.32E+00 |
| 113.8 | 2.63E-06 | 1.19E+00 | 4.87E-04 | 1.19E+00 |
| 100 | 2.63E-06 | 1.37E+00 | 1.92E-04 | 1.37E+00 |
| 88.9 | 1.73E-06 | 8.81E-01 | 3.76E-04 | 8.81E-01 |
| 58.7 | 1.73E-06 | 1.02E+00 | 1.21E-04 | 1.02E+00 |
| 47.3 | 1.73E-06 | 1.17E+00 | 7.41E-04 | 1.17E+00 |
| 25.8 | 1.73E-06 | 8.27E-01 | 2.79E-04 | 8.28E-01 |

**TABLE 3-79: SUMMARY OF ADD_{95%UCL} FOR FEMALE OTTER USING
1993 DATA BASED ON TRI+ CONGENERS**

| Location | Drinking Water 95% UCL | Piscivorous Fish 95% UCL | Sediment 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) |
|------------------------------|------------------------------|--------------------------------|---------------------|--|
| <i>Upper River</i> | | | | |
| Thompson's Island Pool (189) | 1.89E-05 | 2.25E+01 | 8.38E-03 | 2.25E+01 |
| Stillwater (168) | 3.37E-05 | 2.66E+00 | 2.61E-02 | 2.69E+00 |
| Federal Dam (154) | 1.59E-05 | 2.69E+00 | 2.26E-03 | 2.69E+00 |
| <i>Lower River</i> | | | | |
| 143.5 | 6.25E-05 | 2.69E+00 | 4.54E-04 | 2.69E+00 |
| 137.2 | 6.25E-05 | 1.34E+01 | 1.48E-03 | 1.34E+01 |
| 122.4 | 3.37E-05 | 1.81E+00 | 5.16E-04 | 1.81E+00 |
| 113.8 | 3.37E-05 | 1.67E+00 | 8.04E-04 | 1.67E+00 |
| 100 | 3.37E-05 | 4.22E+00 | 4.15E-03 | 4.23E+00 |
| 88.9 | 7.69E-06 | 1.68E+00 | 1.10E-03 | 1.68E+00 |
| 58.7 | 7.69E-06 | 1.52E+00 | 1.35E-03 | 1.52E+00 |
| 47.3 | 7.69E-06 | 3.15E+00 | 2.89E-03 | 3.16E+00 |
| 25.8 | 7.69E-06 | 1.64E+00 | 7.54E-04 | 1.64E+00 |

**TABLE 3-80: SUMMARY OF ADD_{Expected} FOR FEMALE OTTER
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 6.46E+00 | 4.87E+00 | 2.39E+00 |
| 1994 | 3.93E+00 | 3.35E+00 | 1.70E+00 |
| 1995 | 4.33E+00 | 3.72E+00 | 1.83E+00 |
| 1996 | 3.95E+00 | 2.99E+00 | 1.61E+00 |
| 1997 | 3.08E+00 | 2.62E+00 | 1.39E+00 |
| 1998 | 2.57E+00 | 2.35E+00 | 1.17E+00 |
| 1999 | 2.24E+00 | 1.92E+00 | 8.89E-01 |
| 2000 | 2.09E+00 | 1.81E+00 | 8.47E-01 |
| 2001 | 1.95E+00 | 1.75E+00 | 8.14E-01 |
| 2002 | 1.92E+00 | 1.86E+00 | 8.65E-01 |
| 2003 | 1.68E+00 | 1.56E+00 | 7.38E-01 |
| 2004 | 1.57E+00 | 1.52E+00 | 7.61E-01 |
| 2005 | 1.40E+00 | 1.16E+00 | 5.74E-01 |
| 2006 | 1.37E+00 | 1.27E+00 | 6.05E-01 |
| 2007 | 1.27E+00 | 1.06E+00 | 4.69E-01 |
| 2008 | 1.20E+00 | 9.56E-01 | 4.54E-01 |
| 2009 | 1.21E+00 | 1.03E+00 | 4.89E-01 |
| 2010 | 1.06E+00 | 9.42E-01 | 4.09E-01 |
| 2011 | 9.32E-01 | 8.03E-01 | 3.97E-01 |
| 2012 | 9.25E-01 | 8.13E-01 | 3.86E-01 |
| 2013 | 8.88E-01 | 8.46E-01 | 3.72E-01 |
| 2014 | 8.08E-01 | 8.06E-01 | 3.40E-01 |
| 2015 | 7.40E-01 | 7.07E-01 | 3.14E-01 |
| 2016 | 7.55E-01 | 7.93E-01 | 3.38E-01 |
| 2017 | 6.50E-01 | 6.01E-01 | 2.77E-01 |
| 2018 | 6.30E-01 | 5.77E-01 | 2.52E-01 |

**TABLE 3-81: SUMMARY OF ADD_{95%UCL} FOR FEMALE OTTER
BASED ON TRI+ PREDICTIONS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 8.10E+00 | 5.94E+00 | 2.98E+00 |
| 1994 | 5.29E+00 | 4.03E+00 | 2.09E+00 |
| 1995 | 5.79E+00 | 4.46E+00 | 2.23E+00 |
| 1996 | 5.69E+00 | 3.64E+00 | 2.03E+00 |
| 1997 | 4.43E+00 | 3.15E+00 | 1.74E+00 |
| 1998 | 3.65E+00 | 2.86E+00 | 1.43E+00 |
| 1999 | 3.31E+00 | 2.37E+00 | 1.12E+00 |
| 2000 | 3.10E+00 | 2.22E+00 | 1.05E+00 |
| 2001 | 2.88E+00 | 2.14E+00 | 1.01E+00 |
| 2002 | 2.83E+00 | 2.28E+00 | 1.08E+00 |
| 2003 | 2.50E+00 | 1.93E+00 | 9.27E-01 |
| 2004 | 2.31E+00 | 1.87E+00 | 9.54E-01 |
| 2005 | 2.10E+00 | 1.42E+00 | 7.18E-01 |
| 2006 | 2.03E+00 | 1.55E+00 | 7.43E-01 |
| 2007 | 1.91E+00 | 1.29E+00 | 5.76E-01 |
| 2008 | 1.82E+00 | 1.18E+00 | 5.75E-01 |
| 2009 | 1.77E+00 | 1.26E+00 | 6.16E-01 |
| 2010 | 1.55E+00 | 1.16E+00 | 5.07E-01 |
| 2011 | 1.41E+00 | 9.86E-01 | 5.03E-01 |
| 2012 | 1.37E+00 | 1.00E+00 | 4.86E-01 |
| 2013 | 1.31E+00 | 1.04E+00 | 4.72E-01 |
| 2014 | 1.17E+00 | 9.88E-01 | 4.22E-01 |
| 2015 | 1.09E+00 | 8.61E-01 | 3.96E-01 |
| 2016 | 1.09E+00 | 9.69E-01 | 4.19E-01 |
| 2017 | 9.84E-01 | 7.39E-01 | 3.65E-01 |
| 2018 | 9.41E-01 | 7.11E-01 | 3.23E-01 |

**TABLE 3-82: SUMMARY OF ADD_{Expected} FOR FEMALE BAT USING
1993 DATA ON A TEQ BASIS**

| Location | Drinking Water Expected | Benthic Invertebrate Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|-------------------------------------|--|
| <i>Upper River</i> | | | |
| Thompson Island Pool (189) | 7.21E-08 | 1.33E-04 | 1.33E-04 |
| Stillwater (168) | 1.28E-07 | 2.32E-04 | 2.32E-04 |
| Federal Dam (154) | 8.96E-08 | 5.54E-05 | 5.54E-05 |
| <i>Lower River</i> | | | |
| 143.5 | 6.93E-08 | 7.71E-06 | 7.78E-06 |
| 137.2 | 6.93E-08 | 1.52E-05 | 1.53E-05 |
| 122.4 | 3.18E-08 | 8.40E-06 | 8.43E-06 |
| 113.8 | 3.18E-08 | 8.64E-06 | 8.67E-06 |
| 100 | 3.18E-08 | 3.97E-06 | 4.00E-06 |
| 88.9 | 2.09E-08 | 1.99E-06 | 2.01E-06 |
| 58.7 | 2.09E-08 | 6.13E-06 | 6.15E-06 |
| 47.3 | 2.09E-08 | 6.96E-06 | 6.98E-06 |
| 25.8 | 2.09E-08 | 2.06E-06 | 2.08E-06 |

**TABLE 3-83: SUMMARY OF ADD_{95%UCL} FOR FEMALE BAT
USING 1993 DATA ON A TEQ BASIS**

| Location | Drinking Water 95% UCL | Benthic Invertebrate 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|------------------------------------|--|
| <i>Upper River</i> | | | |
| Thompson Island Pool (189) | 2.28E-07 | 2.32E-04 | 2.32E-04 |
| Stillwater (168) | 4.07E-07 | 1.09E-03 | 1.09E-03 |
| Federal Dam (154) | 1.92E-07 | 9.41E-05 | 9.43E-05 |
| <i>Lower River</i> | | | |
| 143.5 | 7.55E-07 | 1.89E-05 | 1.97E-05 |
| 137.2 | 7.55E-07 | 6.17E-05 | 6.24E-05 |
| 122.4 | 4.07E-07 | 2.11E-05 | 2.15E-05 |
| 113.8 | 4.07E-07 | 3.35E-05 | 3.39E-05 |
| 100 | 4.07E-07 | 2.71E-05 | 2.76E-05 |
| 88.9 | 9.29E-08 | 3.54E-06 | 3.63E-06 |
| 58.7 | 9.29E-08 | 5.61E-05 | 5.62E-05 |
| 47.3 | 9.29E-08 | 5.11E-05 | 5.12E-05 |
| 25.8 | 9.29E-08 | 3.50E-06 | 3.59E-06 |

**TABLE 3-84: SUMMARY OF ADD_{Expected} FOR FEMALE BAT
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 1.33E-04 | 5.22E-05 | 2.31E-05 |
| 1994 | 1.49E-04 | 4.98E-05 | 2.18E-05 |
| 1995 | 1.38E-04 | 4.67E-05 | 1.99E-05 |
| 1996 | 1.26E-04 | 4.20E-05 | 1.70E-05 |
| 1997 | 1.15E-04 | 3.80E-05 | 1.47E-05 |
| 1998 | 1.04E-04 | 3.44E-05 | 1.28E-05 |
| 1999 | 9.53E-05 | 3.14E-05 | 1.12E-05 |
| 2000 | 8.84E-05 | 2.96E-05 | 1.00E-05 |
| 2001 | 8.36E-05 | 2.88E-05 | 9.47E-06 |
| 2002 | 7.86E-05 | 2.76E-05 | 9.02E-06 |
| 2003 | 7.26E-05 | 2.60E-05 | 8.43E-06 |
| 2004 | 6.67E-05 | 2.38E-05 | 7.60E-06 |
| 2005 | 6.27E-05 | 2.20E-05 | 6.91E-06 |
| 2006 | 5.96E-05 | 2.09E-05 | 6.41E-06 |
| 2007 | 5.58E-05 | 1.95E-05 | 5.80E-06 |
| 2008 | 5.25E-05 | 1.84E-05 | 5.42E-06 |
| 2009 | 4.96E-05 | 1.76E-05 | 5.23E-06 |
| 2010 | 4.53E-05 | 1.64E-05 | 4.90E-06 |
| 2011 | 4.07E-05 | 1.51E-05 | 4.48E-06 |
| 2012 | 3.80E-05 | 1.44E-05 | 4.27E-06 |
| 2013 | 3.55E-05 | 1.37E-05 | 3.99E-06 |
| 2014 | 3.30E-05 | 1.28E-05 | 3.64E-06 |
| 2015 | 3.12E-05 | 1.24E-05 | 3.50E-06 |
| 2016 | 2.93E-05 | 1.18E-05 | 3.23E-06 |
| 2017 | 2.75E-05 | 1.11E-05 | 2.89E-06 |
| 2018 | 2.68E-05 | 1.08E-05 | 2.78E-06 |

**TABLE 3-85: SUMMARY OF ADD_{95%UCL} FOR FEMALE BAT
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 2.00E-04 | 5.99E-05 | 2.93E-05 |
| 1994 | 1.87E-04 | 5.71E-05 | 2.76E-05 |
| 1995 | 1.73E-04 | 5.36E-05 | 2.53E-05 |
| 1996 | 1.58E-04 | 4.83E-05 | 2.15E-05 |
| 1997 | 1.45E-04 | 4.36E-05 | 1.87E-05 |
| 1998 | 1.31E-04 | 3.94E-05 | 1.63E-05 |
| 1999 | 1.20E-04 | 3.60E-05 | 1.42E-05 |
| 2000 | 1.11E-04 | 3.40E-05 | 1.28E-05 |
| 2001 | 1.05E-04 | 3.30E-05 | 1.20E-05 |
| 2002 | 9.86E-05 | 3.17E-05 | 1.15E-05 |
| 2003 | 9.11E-05 | 2.98E-05 | 1.07E-05 |
| 2004 | 8.37E-05 | 2.73E-05 | 9.66E-06 |
| 2005 | 7.86E-05 | 2.53E-05 | 8.78E-06 |
| 2006 | 7.49E-05 | 2.40E-05 | 8.15E-06 |
| 2007 | 7.00E-05 | 2.24E-05 | 7.38E-06 |
| 2008 | 6.59E-05 | 2.11E-05 | 6.89E-06 |
| 2009 | 6.23E-05 | 2.02E-05 | 6.64E-06 |
| 2010 | 5.68E-05 | 1.88E-05 | 6.23E-06 |
| 2011 | 5.11E-05 | 1.73E-05 | 5.69E-06 |
| 2012 | 4.76E-05 | 1.66E-05 | 5.43E-06 |
| 2013 | 4.46E-05 | 1.57E-05 | 5.08E-06 |
| 2014 | 4.14E-05 | 1.47E-05 | 4.62E-06 |
| 2015 | 3.92E-05 | 1.43E-05 | 4.45E-06 |
| 2016 | 3.67E-05 | 1.35E-05 | 4.11E-06 |
| 2017 | 3.45E-05 | 1.27E-05 | 3.67E-06 |
| 2018 | 3.36E-05 | 1.24E-05 | 3.54E-06 |

**TABLE 3-86: SUMMARY OF ADD_{Expected} FOR FEMALE RACCOON USING
1993 DATA ON A TEQ BASIS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 3.83E-08 | 4.02E-06 | 2.16E-05 | 4.29E-05 | 6.85E-05 |
| Stillwater (168) | 6.80E-08 | 5.39E-07 | 3.78E-05 | 1.12E-04 | 1.50E-04 |
| Federal Dam (154) | 4.75E-08 | 3.30E-07 | 9.00E-06 | 1.01E-05 | 1.95E-05 |
| <i>Lower River</i> | | | | | |
| 143.5 | 3.68E-08 | 3.83E-07 | 1.25E-06 | 3.11E-06 | 4.78E-06 |
| 137.2 | 3.68E-08 | 7.68E-07 | 2.47E-06 | 5.48E-06 | 8.76E-06 |
| 122.4 | 1.68E-08 | 2.93E-07 | 1.37E-06 | 3.47E-06 | 5.15E-06 |
| 113.8 | 1.68E-08 | 3.11E-07 | 1.40E-06 | 3.64E-06 | 5.37E-06 |
| 100 | 1.68E-08 | 1.34E-07 | 6.45E-07 | 1.44E-06 | 2.24E-06 |
| 88.9 | 1.11E-08 | 2.65E-07 | 3.24E-07 | 2.82E-06 | 3.42E-06 |
| 58.7 | 1.11E-08 | 2.91E-07 | 9.97E-07 | 9.09E-07 | 2.21E-06 |
| 47.3 | 1.11E-08 | 2.59E-07 | 1.13E-06 | 5.55E-06 | 6.95E-06 |
| 25.8 | 1.11E-08 | 1.95E-07 | 3.35E-07 | 2.09E-06 | 2.63E-06 |

**TABLE 3-87: SUMMARY OF ADD_{95%UCL} FOR FEMALE RACCOON USING
1993 DATA ON A TEQ BASIS**

| Location | Drinking Water 95% UCL | Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Average Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|-----------------|------------------------------------|---------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 1.21E-07 | 6.14E-06 | 3.77E-05 | 6.27E-05 | 1.07E-04 |
| Stillwater (168) | 2.16E-07 | 9.94E-07 | 1.77E-04 | 1.96E-04 | 3.74E-04 |
| Federal Dam (154) | 1.02E-07 | 4.73E-07 | 1.53E-05 | 1.69E-05 | 3.28E-05 |
| <i>Lower River</i> | | | | | |
| 143.5 | 4.01E-07 | 4.53E-07 | 3.08E-06 | 3.40E-06 | 7.33E-06 |
| 137.2 | 4.01E-07 | 1.63E-06 | 1.00E-05 | 1.11E-05 | 2.31E-05 |
| 122.4 | 2.16E-07 | 4.60E-07 | 3.43E-06 | 3.86E-06 | 7.97E-06 |
| 113.8 | 2.16E-07 | 3.22E-07 | 5.44E-06 | 6.02E-06 | 1.20E-05 |
| 100 | 2.16E-07 | 2.27E-07 | 4.41E-06 | 3.11E-05 | 3.59E-05 |
| 88.9 | 4.93E-08 | 3.48E-07 | 5.75E-07 | 8.24E-06 | 9.22E-06 |
| 58.7 | 4.93E-08 | 3.28E-07 | 9.12E-06 | 1.01E-05 | 1.96E-05 |
| 47.3 | 4.93E-08 | 3.41E-07 | 8.31E-06 | 2.17E-05 | 3.04E-05 |
| 25.8 | 4.93E-08 | 2.34E-07 | 5.68E-07 | 5.64E-06 | 6.49E-06 |

**TABLE 3-88: SUMMARY OF ADD_{Expected} FOR FEMALE RACCOON
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 1.27E-04 | 4.27E-05 | 1.89E-05 |
| 1994 | 1.22E-04 | 4.03E-05 | 1.74E-05 |
| 1995 | 1.14E-04 | 3.88E-05 | 1.68E-05 |
| 1996 | 1.04E-04 | 3.47E-05 | 1.41E-05 |
| 1997 | 9.57E-05 | 3.16E-05 | 1.23E-05 |
| 1998 | 8.64E-05 | 2.83E-05 | 1.06E-05 |
| 1999 | 7.88E-05 | 2.59E-05 | 9.36E-06 |
| 2000 | 7.22E-05 | 2.38E-05 | 8.18E-06 |
| 2001 | 6.83E-05 | 2.33E-05 | 7.65E-06 |
| 2002 | 6.45E-05 | 2.24E-05 | 7.31E-06 |
| 2003 | 6.00E-05 | 2.13E-05 | 6.90E-06 |
| 2004 | 5.50E-05 | 1.97E-05 | 6.32E-06 |
| 2005 | 5.09E-05 | 1.78E-05 | 5.59E-06 |
| 2006 | 4.88E-05 | 1.71E-05 | 5.29E-06 |
| 2007 | 4.58E-05 | 1.59E-05 | 4.77E-06 |
| 2008 | 4.28E-05 | 1.49E-05 | 4.38E-06 |
| 2009 | 4.07E-05 | 1.42E-05 | 4.21E-06 |
| 2010 | 3.79E-05 | 1.36E-05 | 4.05E-06 |
| 2011 | 3.36E-05 | 1.23E-05 | 3.66E-06 |
| 2012 | 3.10E-05 | 1.16E-05 | 3.44E-06 |
| 2013 | 2.93E-05 | 1.13E-05 | 3.32E-06 |
| 2014 | 2.70E-05 | 1.04E-05 | 2.96E-06 |
| 2015 | 2.54E-05 | 9.99E-06 | 2.81E-06 |
| 2016 | 2.42E-05 | 9.71E-06 | 2.73E-06 |
| 2017 | 2.23E-05 | 8.93E-06 | 2.34E-06 |
| 2018 | 2.11E-05 | 8.46E-06 | 2.18E-06 |

**TABLE 3-89: SUMMARY OF ADD_{95%UCL} FOR FEMALE RACCOON
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 1.44E-04 | 4.41E-05 | 2.00E-05 |
| 1994 | 1.33E-04 | 4.16E-05 | 1.84E-05 |
| 1995 | 1.25E-04 | 4.00E-05 | 1.77E-05 |
| 1996 | 1.14E-04 | 3.59E-05 | 1.50E-05 |
| 1997 | 1.05E-04 | 3.26E-05 | 1.30E-05 |
| 1998 | 9.45E-05 | 2.93E-05 | 1.13E-05 |
| 1999 | 8.62E-05 | 2.67E-05 | 9.88E-06 |
| 2000 | 7.90E-05 | 2.46E-05 | 8.68E-06 |
| 2001 | 7.48E-05 | 2.40E-05 | 8.09E-06 |
| 2002 | 7.06E-05 | 2.31E-05 | 7.73E-06 |
| 2003 | 6.56E-05 | 2.20E-05 | 7.30E-06 |
| 2004 | 6.01E-05 | 2.04E-05 | 6.69E-06 |
| 2005 | 5.57E-05 | 1.84E-05 | 5.93E-06 |
| 2006 | 5.35E-05 | 1.77E-05 | 5.58E-06 |
| 2007 | 5.01E-05 | 1.64E-05 | 5.05E-06 |
| 2008 | 4.68E-05 | 1.54E-05 | 4.63E-06 |
| 2009 | 4.45E-05 | 1.47E-05 | 4.70E-06 |
| 2010 | 4.14E-05 | 1.40E-05 | 4.28E-06 |
| 2011 | 3.67E-05 | 1.27E-05 | 3.88E-06 |
| 2012 | 3.39E-05 | 1.20E-05 | 3.63E-06 |
| 2013 | 3.21E-05 | 1.16E-05 | 3.50E-06 |
| 2014 | 2.96E-05 | 1.08E-05 | 3.14E-06 |
| 2015 | 2.79E-05 | 1.03E-05 | 2.97E-06 |
| 2016 | 2.65E-05 | 1.00E-05 | 2.88E-06 |
| 2017 | 2.44E-05 | 9.22E-06 | 2.48E-06 |
| 2018 | 2.32E-05 | 8.73E-06 | 2.30E-06 |

**TABLE 3-90: SUMMARY OF ADD_{Expected} FOR FEMALE MINK USING 1993 DATA
ON A TEQ BASIS**

| Location | Drinking Water Expected | Forage Fish Expected | Benthic Invertebrate Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|----------------------------|-------------------------------|----------------------------|-------------------------------------|----------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 4.71E-08 | 4.69E-05 | 9.89E-06 | 6.57E-06 | 6.34E-05 |
| Stillwater (168) | 8.37E-08 | 6.28E-06 | 1.73E-05 | 1.72E-05 | 4.08E-05 |
| Federal Dam (154) | 5.85E-08 | 3.84E-06 | 4.13E-06 | 1.54E-06 | 9.57E-06 |
| <i>Lower River</i> | | | | | |
| 143.5 | 4.53E-08 | 4.46E-06 | 5.75E-07 | 4.76E-07 | 5.55E-06 |
| 137.2 | 4.53E-08 | 8.95E-06 | 1.13E-06 | 8.40E-07 | 1.10E-05 |
| 122.4 | 2.07E-08 | 3.41E-06 | 6.26E-07 | 5.32E-07 | 4.59E-06 |
| 113.8 | 2.07E-08 | 3.62E-06 | 6.44E-07 | 5.58E-07 | 4.84E-06 |
| 100 | 2.07E-08 | 1.56E-06 | 2.96E-07 | 2.20E-07 | 2.09E-06 |
| 88.9 | 1.37E-08 | 3.09E-06 | 1.48E-07 | 4.32E-07 | 3.68E-06 |
| 58.7 | 1.37E-08 | 3.40E-06 | 4.57E-07 | 1.39E-07 | 4.01E-06 |
| 47.3 | 1.37E-08 | 3.02E-06 | 5.19E-07 | 8.50E-07 | 4.40E-06 |
| 25.8 | 1.37E-08 | 2.27E-06 | 1.54E-07 | 3.20E-07 | 2.76E-06 |

**TABLE 3-91: SUMMARY OF ADD_{95%UCL} FOR FEMALE MINK USING 1993 DATA
ON A TEQ BASIS**

| Location | Drinking Water 95% UCL | Forage Fish 95% UCL | Benthic Invertebrate 95% UCL | Sediment 95% UCL | Total Upper Bound Daily Dose _{95%UCL} (mg/Kg/day) |
|----------------------------|------------------------------|---------------------------|------------------------------------|---------------------|--|
| <i>Upper River</i> | | | | | |
| Thompson Island Pool (189) | 1.49E-07 | 7.15E-05 | 1.73E-05 | 9.61E-06 | 9.86E-05 |
| Stillwater (168) | 2.66E-07 | 1.16E-05 | 8.11E-05 | 2.99E-05 | 1.23E-04 |
| Federal Dam (154) | 1.26E-07 | 5.52E-06 | 7.01E-06 | 2.59E-06 | 1.52E-05 |
| <i>Lower River</i> | | | | | |
| 143.5 | 4.93E-07 | 5.28E-06 | 1.41E-06 | 5.21E-07 | 7.70E-06 |
| 137.2 | 4.93E-07 | 1.90E-05 | 4.59E-06 | 1.70E-06 | 2.58E-05 |
| 122.4 | 2.66E-07 | 5.36E-06 | 1.57E-06 | 5.91E-07 | 7.79E-06 |
| 113.8 | 2.66E-07 | 3.75E-06 | 2.50E-06 | 9.22E-07 | 7.44E-06 |
| 100 | 2.66E-07 | 2.64E-06 | 2.02E-06 | 4.76E-06 | 9.69E-06 |
| 88.9 | 6.07E-08 | 4.06E-06 | 2.64E-07 | 1.26E-06 | 5.64E-06 |
| 58.7 | 6.07E-08 | 3.82E-06 | 4.18E-06 | 1.54E-06 | 9.61E-06 |
| 47.3 | 6.07E-08 | 3.97E-06 | 3.81E-06 | 3.32E-06 | 1.12E-05 |
| 25.8 | 6.07E-08 | 2.72E-06 | 2.61E-07 | 8.64E-07 | 3.91E-06 |

**TABLE 3-92: SUMMARY OF ADD_{Expected} FOR FEMALE MINK
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Average Dietary Dose (mg/Kg/day) | | |
|------|-------------------------------------|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 4.18E-05 | 1.67E-05 | 6.40E-06 |
| 1994 | 4.08E-05 | 1.58E-05 | 6.05E-06 |
| 1995 | 3.99E-05 | 1.54E-05 | 5.99E-06 |
| 1996 | 3.45E-05 | 1.25E-05 | 4.76E-06 |
| 1997 | 3.10E-05 | 1.22E-05 | 4.24E-06 |
| 1998 | 2.76E-05 | 9.76E-06 | 3.60E-06 |
| 1999 | 2.54E-05 | 9.21E-06 | 3.22E-06 |
| 2000 | 2.32E-05 | 8.21E-06 | 2.94E-06 |
| 2001 | 2.21E-05 | 8.51E-06 | 2.71E-06 |
| 2002 | 2.07E-05 | 7.89E-06 | 2.55E-06 |
| 2003 | 1.92E-05 | 7.48E-06 | 2.47E-06 |
| 2004 | 1.76E-05 | 6.96E-06 | 2.21E-06 |
| 2005 | 1.65E-05 | 6.22E-06 | 2.02E-06 |
| 2006 | 1.58E-05 | 6.12E-06 | 1.89E-06 |
| 2007 | 1.47E-05 | 5.53E-06 | 1.77E-06 |
| 2008 | 1.38E-05 | 5.09E-06 | 1.56E-06 |
| 2009 | 1.32E-05 | 5.12E-06 | 1.60E-06 |
| 2010 | 1.21E-05 | 4.76E-06 | 1.53E-06 |
| 2011 | 1.08E-05 | 4.19E-06 | 1.38E-06 |
| 2012 | 1.01E-05 | 4.33E-06 | 1.41E-06 |
| 2013 | 9.46E-06 | 4.01E-06 | 1.27E-06 |
| 2014 | 8.76E-06 | 3.80E-06 | 1.17E-06 |
| 2015 | 8.25E-06 | 3.58E-06 | 1.13E-06 |
| 2016 | 7.82E-06 | 3.54E-06 | 1.12E-06 |
| 2017 | 7.25E-06 | 3.09E-06 | 9.09E-07 |
| 2018 | 6.97E-06 | 3.07E-06 | 9.06E-07 |

**TABLE 3-93: SUMMARY OF ADD_{95%UCL} FOR FEMALE MINK
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|-------------------------------------|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 4.89E-05 | 1.77E-05 | 7.04E-06 |
| 1994 | 4.56E-05 | 1.67E-05 | 6.65E-06 |
| 1995 | 4.45E-05 | 1.62E-05 | 6.54E-06 |
| 1996 | 3.86E-05 | 1.33E-05 | 5.24E-06 |
| 1997 | 3.48E-05 | 1.28E-05 | 4.63E-06 |
| 1998 | 3.11E-05 | 1.04E-05 | 3.97E-06 |
| 1999 | 2.85E-05 | 9.77E-06 | 3.53E-06 |
| 2000 | 2.61E-05 | 8.75E-06 | 3.22E-06 |
| 2001 | 2.49E-05 | 9.02E-06 | 2.98E-06 |
| 2002 | 2.33E-05 | 8.39E-06 | 2.81E-06 |
| 2003 | 2.16E-05 | 7.95E-06 | 2.71E-06 |
| 2004 | 1.98E-05 | 7.39E-06 | 2.43E-06 |
| 2005 | 1.85E-05 | 6.62E-06 | 2.21E-06 |
| 2006 | 1.77E-05 | 6.50E-06 | 2.07E-06 |
| 2007 | 1.65E-05 | 5.89E-06 | 1.93E-06 |
| 2008 | 1.55E-05 | 5.43E-06 | 1.71E-06 |
| 2009 | 1.48E-05 | 5.43E-06 | 1.79E-06 |
| 2010 | 1.36E-05 | 5.07E-06 | 1.67E-06 |
| 2011 | 1.21E-05 | 4.48E-06 | 1.51E-06 |
| 2012 | 1.13E-05 | 4.59E-06 | 1.53E-06 |
| 2013 | 1.06E-05 | 4.26E-06 | 1.38E-06 |
| 2014 | 9.85E-06 | 4.04E-06 | 1.28E-06 |
| 2015 | 9.29E-06 | 3.81E-06 | 1.23E-06 |
| 2016 | 8.79E-06 | 3.75E-06 | 1.21E-06 |
| 2017 | 8.15E-06 | 3.29E-06 | 9.92E-07 |
| 2018 | 7.85E-06 | 3.26E-06 | 9.85E-07 |

**TABLE 3-94: SUMMARY OF ADD_{Expected} FOR FEMALE OTTER USING
1993 DATA ON A TEQ BASIS**

| Location | Drinking Water Expected | Piscivorous Fish Expected | Sediment Expected | Total Average Daily Dose _{Expected} (mg/Kg/day) |
|------------------------------|-------------------------------|---------------------------------|----------------------|--|
| <i>Upper River</i> | | | | |
| Thompson's Island Pool (189) | 3.78E-08 | 4.98E-04 | 4.46E-06 | 5.03E-04 |
| Stillwater (168) | 6.71E-08 | 8.86E-05 | 1.16E-05 | 1.00E-04 |
| Federal Dam (154) | 4.69E-08 | 6.43E-05 | 1.05E-06 | 6.54E-05 |
| <i>Lower River</i> | | | | |
| 143.5 | 3.63E-08 | 6.43E-05 | 3.23E-07 | 6.46E-05 |
| 137.2 | 3.63E-08 | 2.42E-04 | 5.70E-07 | 2.43E-04 |
| 122.4 | 1.66E-08 | 5.65E-05 | 3.61E-07 | 5.69E-05 |
| 113.8 | 1.66E-08 | 5.11E-05 | 3.78E-07 | 5.15E-05 |
| 100 | 1.66E-08 | 5.87E-05 | 1.50E-07 | 5.88E-05 |
| 88.9 | 1.10E-08 | 3.78E-05 | 2.93E-07 | 3.81E-05 |
| 58.7 | 1.10E-08 | 4.39E-05 | 9.45E-08 | 4.40E-05 |
| 47.3 | 1.10E-08 | 5.03E-05 | 5.77E-07 | 5.08E-05 |
| 25.8 | 1.10E-08 | 3.55E-05 | 2.17E-07 | 3.57E-05 |

**TABLE 3-95: SUMMARY OF ADD_{95%UCL} FOR FEMALE OTTER USING 1993 DATA
ON A TEQ BASIS**

| Location | Drinking Water 95% UCL | Piscivorous Fish 95% UCL | Sediment 95% UCL | Total Average Daily Dosage _{95%UCL} (mg/Kg/day) |
|------------------------------|------------------------------|--------------------------------|---------------------|--|
| <i>Upper River</i> | | | | |
| Thompson's Island Pool (189) | 1.20E-07 | 9.65E-04 | 6.52E-06 | 9.71E-04 |
| Stillwater (168) | 2.13E-07 | 1.14E-04 | 2.03E-05 | 1.35E-04 |
| Federal Dam (154) | 1.01E-07 | 1.15E-04 | 1.76E-06 | 1.17E-04 |
| <i>Lower River</i> | | | | |
| 143.5 | 3.95E-07 | 1.15E-04 | 3.53E-07 | 1.16E-04 |
| 137.2 | 3.95E-07 | 5.77E-04 | 1.15E-06 | 5.78E-04 |
| 122.4 | 2.13E-07 | 7.76E-05 | 4.01E-07 | 7.83E-05 |
| 113.8 | 2.13E-07 | 7.14E-05 | 6.25E-07 | 7.23E-05 |
| 100 | 2.13E-07 | 1.81E-04 | 3.23E-06 | 1.85E-04 |
| 88.9 | 4.87E-08 | 7.19E-05 | 8.57E-07 | 7.28E-05 |
| 58.7 | 4.87E-08 | 6.51E-05 | 1.05E-06 | 6.61E-05 |
| 47.3 | 4.87E-08 | 1.35E-04 | 2.25E-06 | 1.38E-04 |
| 25.8 | 4.87E-08 | 7.05E-05 | 5.86E-07 | 7.12E-05 |

**TABLE 3-96: SUMMARY OF ADD_{Expected} FOR FEMALE OTTER
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total Average Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 2.87E-04 | 2.12E-04 | 1.04E-04 |
| 1994 | 1.78E-04 | 1.47E-04 | 7.42E-05 |
| 1995 | 1.95E-04 | 1.62E-04 | 8.00E-05 |
| 1996 | 1.78E-04 | 1.31E-04 | 7.02E-05 |
| 1997 | 1.40E-04 | 1.15E-04 | 6.05E-05 |
| 1998 | 1.17E-04 | 1.03E-04 | 5.09E-05 |
| 1999 | 1.02E-04 | 8.44E-05 | 3.89E-05 |
| 2000 | 9.52E-05 | 7.96E-05 | 3.70E-05 |
| 2001 | 8.90E-05 | 7.67E-05 | 3.55E-05 |
| 2002 | 8.76E-05 | 8.15E-05 | 3.77E-05 |
| 2003 | 7.68E-05 | 6.87E-05 | 3.22E-05 |
| 2004 | 7.18E-05 | 6.69E-05 | 3.32E-05 |
| 2005 | 6.39E-05 | 5.13E-05 | 2.51E-05 |
| 2006 | 6.28E-05 | 5.58E-05 | 2.64E-05 |
| 2007 | 5.79E-05 | 4.67E-05 | 2.05E-05 |
| 2008 | 5.47E-05 | 4.22E-05 | 1.98E-05 |
| 2009 | 5.51E-05 | 4.54E-05 | 2.13E-05 |
| 2010 | 4.86E-05 | 4.15E-05 | 1.78E-05 |
| 2011 | 4.26E-05 | 3.54E-05 | 1.73E-05 |
| 2012 | 4.21E-05 | 3.58E-05 | 1.68E-05 |
| 2013 | 4.04E-05 | 3.72E-05 | 1.62E-05 |
| 2014 | 3.68E-05 | 3.54E-05 | 1.48E-05 |
| 2015 | 3.37E-05 | 3.11E-05 | 1.37E-05 |
| 2016 | 3.43E-05 | 3.48E-05 | 1.47E-05 |
| 2017 | 2.96E-05 | 2.65E-05 | 1.21E-05 |
| 2018 | 2.87E-05 | 2.54E-05 | 1.10E-05 |

**TABLE 3-97: SUMMARY OF ADD_{95%UCL} FOR FEMALE OTTER
ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | Total 95% UCL Dietary Dose (mg/Kg/day) | | |
|------|---|----------|----------|
| | 189 | 168 | 154 |
| 1993 | 3.58E-04 | 2.58E-04 | 1.30E-04 |
| 1994 | 2.37E-04 | 1.76E-04 | 9.10E-05 |
| 1995 | 2.58E-04 | 1.95E-04 | 9.70E-05 |
| 1996 | 2.53E-04 | 1.59E-04 | 8.82E-05 |
| 1997 | 1.98E-04 | 1.38E-04 | 7.55E-05 |
| 1998 | 1.64E-04 | 1.25E-04 | 6.23E-05 |
| 1999 | 1.49E-04 | 1.04E-04 | 4.88E-05 |
| 2000 | 1.39E-04 | 9.69E-05 | 4.55E-05 |
| 2001 | 1.29E-04 | 9.37E-05 | 4.40E-05 |
| 2002 | 1.27E-04 | 9.95E-05 | 4.71E-05 |
| 2003 | 1.12E-04 | 8.44E-05 | 4.03E-05 |
| 2004 | 1.04E-04 | 8.16E-05 | 4.14E-05 |
| 2005 | 9.44E-05 | 6.24E-05 | 3.13E-05 |
| 2006 | 9.12E-05 | 6.80E-05 | 3.23E-05 |
| 2007 | 8.57E-05 | 5.65E-05 | 2.51E-05 |
| 2008 | 8.16E-05 | 5.17E-05 | 2.50E-05 |
| 2009 | 7.92E-05 | 5.53E-05 | 2.68E-05 |
| 2010 | 6.97E-05 | 5.06E-05 | 2.21E-05 |
| 2011 | 6.34E-05 | 4.33E-05 | 2.19E-05 |
| 2012 | 6.11E-05 | 4.39E-05 | 2.11E-05 |
| 2013 | 5.88E-05 | 4.55E-05 | 2.05E-05 |
| 2014 | 5.25E-05 | 4.32E-05 | 1.83E-05 |
| 2015 | 4.87E-05 | 3.77E-05 | 1.72E-05 |
| 2016 | 4.89E-05 | 4.24E-05 | 1.82E-05 |
| 2017 | 4.40E-05 | 3.24E-05 | 1.58E-05 |
| 2018 | 4.21E-05 | 3.11E-05 | 1.40E-05 |

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TABLE 4-1
COMMON EFFECTS OF PCB EXPOSURE IN ANIMALS

Hepatotoxicity

Hepatomegaly; bile duct hyperplasia, proliferation of smooth ER
Focal necrosis; fatty degeneration
Induction of microsomal enzymes; implications for hormone imbalances, pancreas and reproductive effects
Depletion of fat soluble vitamins (predominantly vitamin A)
Porphyria

Immunotoxicity

Atrophy of lymphoid tissues
Reduction in circulating leukocytes and lymphocytes
Suppressed antibody responses
Enhanced susceptibility to viruses
Suppression of natural killer cells

Neurotoxicity

Impaired behavioral responses
Alterations in catecholamine levels
Depressed spontaneous motor activity
Developmental deficits
Numbness in extremities

Reproduction

Increased abortion; low birth weights
Decreased survival and mating success
Increased length of estrus
Embryo and fetal mortality
Gross teratogenic effects
Biochemical, neurological, and functional changes following *in utero* exposure (mammals)
Decreased libido, decreased sperm numbers and motility

Gastrointestinal

Gastric hyperplasia
Ulceration and necrosis

Respiratory

Chronic bronchitis
Decreased vital capacity

Dermal Toxicity

Chloracne
Hyperplasia and hyperkeratosis of epithelium
Edema

Mutagenic Effects

Commercial mixtures are weakly mutagenic

Carcinogenic Effects

Preneoplastic changes
Neoplastic changes
Promotion considered main contribution
Attenuation of other carcinogens under certain conditions

Source: Hansen, L. G.. 1987. Environmental Toxicology of Polychlorinated Biphenyls in Environmental Toxin Series 1. eds. Safe, S. and Hutzinger, O., p. 32.

TABLE 4-2
WORLD-HEALTH ORGANIZATION FOR TOXIC EQUIVALENCY FACTORS (TEFs) FOR HUMANS,
MAMMALS, FISH, AND BIRDS

| Congener | Toxic Equivalency Factor | | |
|--------------------------------|--------------------------|-----------|---------|
| | Humans/Mammals | Fish | Birds |
| Non-ortho PCBs | | | |
| 3,4,4',5-TetraCB (81) | 0.0001 | 0.0005 | 0.1 |
| 3,3',4,4'-TetraCB (77) | 0.0001 | 0.0001 | 0.05 |
| 3,3',4,4',5-PentaCB (126) | 0.1 | 0.005 | 0.1 |
| 3,3',4,4',5,5'-HexaCB (169) | 0.01 | 0.00005 | 0.001 |
| Mono-ortho PCBs | | | |
| 2,3,3',4,4'-PentaCB (105) | 0.0001 | <0.000005 | 0.0001 |
| 2,3,4,4',5-PentaCB (114) | 0.0005 | <0.000005 | 0.0001 |
| 2,3',4,4',5-PentaCB (118) | 0.0001 | <0.000005 | 0.00001 |
| 2',3,4,4',5-PentaCB (123) | 0.0001 | <0.000005 | 0.00001 |
| 2,3,3',4,4',5-HexaCB (156) | 0.0005 | <0.000005 | 0.0001 |
| 2,3,3',4,4',5'-HexaCB (157) | 0.0005 | <0.000005 | 0.0001 |
| 2,3',4,4',5,5'-HexaCB (167) | 0.00001 | <0.000005 | 0.00001 |
| 2,3,3',4,4',5,5'-HeptaCB (189) | 0.0001 | <0.000005 | 0.00001 |

Notes: CB = chlorinated biphenyls

Reference: van den Berg, et al. (1998). Toxic Equivalency Factors (TEFs) for PCBs, PCDDs, PCDFs for Humans and Wildlife. Environmental Health Perspectives, 106:12, 775-791.

**TABLE 4-3
SELECTED SEDIMENT SCREENING GUIDELINES: PCBs**

| | Total PCBs | Aroclor 1254 | Aroclor 1248 | Aroclor 1016 | Aroclor 1260 |
|---|-----------------------|-------------------------|-------------------------|-------------------------|-------------------------|
| <i>Hudson River Sediment Effect Concentrations (mg/kg, or ppm)</i> | | | | | |
| <i>(MacDonald Env. Sci., 1999)</i> | | | | | |
| <i>(Estuarine, freshwater, and saltwater)</i> | | | | | |
| Threshold Effect Concentration | 0.04 | | | | |
| Mid-range Effect Concentration | 0.4 | | | | |
| Extreme Effect Concentration | 1.7 | | | | |
| <i>NYSDEC (1998) (Freshwater) (mg/kg organic carbon)</i> | | | | | |
| Benthic Aquatic Life Acute Toxicity | 2760.8 | | | | |
| Benthic Aquatic Life Chronic Toxicity | 19.3 | | | | |
| Wildlife Bioaccumulation | 1.4 | | | | |
| <i>NYSDEC (1998) (Saltwater) (mg/kg organic carbon)</i> | | | | | |
| Benthic Aquatic Life Acute Toxicity | 13803.3 | | | | |
| Benthic Aquatic Life Chronic Toxicity | 41.4 | | | | |
| Wildlife Bioaccumulation | 1.4 | | | | |
| <i>Ontario Ministry of the Environment Sediment Guidelines (Freshwater)</i> | | | | | |
| <i>(Persaud et al., 1993)</i> | | | | | |
| No Effect Level (mg/kg) | 0.01 | | | | |
| Lowest Effect Level (mg/kg) | 0.07 | 0.06 | 0.03 | 0.007 | 0.005 |
| Severe Effect Level (mg/kg organic carbon) | 530 | 34 | 150 | 53 | 24 |
| <i>Long et al. (1995) Sediment Guidelines (ug/kg)</i> | | | | | |
| <i>(Marine and Estuarine)</i> | | | | | |
| Effects-Range-Low | 22.7 | | | | |
| Effects-Range-Median | 180 | | | | |
| <i>Ingersoll et al. (1996) Sediment Guidelines (ug/kg, or ppb)</i> | | | | | |
| <i>(Freshwater)</i> | | | | | |
| <i>(Derived from 28-day Hyalella azteca data)</i> | | | | | |
| Effects-Range-Low | 50 | | | | |
| Effects-Range-Median | 730 | | | | |
| Threshold Effect Level | 32 | | | | |
| Probable Effect Level | 240 | | | | |
| No Effect Concentration | 190 | | | | |
| <i>Washington State Dep't of Ecology 1997 Sediment Guidelines</i> | | | | | |
| <i>(Freshwater) (ug/kg, or ppb)¹</i> | | | | | |
| Apparent Effects Threshold (Microtox) | 21 | 7.3 | | | |
| Apparent Effects Threshold (<i>Hyalella azteca</i>) | 820 | 350 | | 100 | |
| Probable Apparent Effects Threshold (Microtox) | 21 | 7.3 | 21 | | |
| Probable Apparent Effects Threshold (<i>Hyalella azteca</i>) | 450 | 240 | | 100 | |
| Lowest Apparent Effects Threshold (between Microtox and <i>H. azteca</i>) | 21 | 7.3 | 21 | | |
| <i>Florida Department of Environmental Protection (ug/kg, or ppb)</i> | | | | | |
| <i>(MacDonald, D.D., et al., 1996) (Marine and Estuarine)</i> | | | | | |
| Threshold Effect Level | 21.6 | | | | |
| Probable Effect Level | 189 | | | | |
| <i>Jones et al. (1997) (ug/kg, or ppb)</i> | | | | | |
| <i>EqP-derived; recommended TOC adjustment</i> | | | | | |
| Secondary Chronic Value | | 810 | 1000 | | 450000 |
| <i>Smith et al. (1996) (ug/kg, or ppb)</i> | | | | | |
| Threshold Effect Level | 34.1 | | | | |
| Probable Effect Level | 277 | | | | |

Note: All values are dry weight unless noted.

Please note that for Washington state values, the Aroclor 1016 column becomes Aroclor 1242. This applies only to this one set of values.

¹ Some values also available in mg/kg organic carbon

TABLE 4-4
TOXICITY ENDPOINTS FOR BENTHIC INVERTEBRATES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs, AROCLORS, AND DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EXPOSURE MEDIA | PCB TYPE | EXPOSURE DURATION | EFFECT LEVEL | EFFECT CONC, WHOLE BODY CONC. (mg/kg wet wt) | EFFECT ENDPOINT | REFERENCE |
|--|-----------------|--------------|-------------------|-------------------|--|--------------------------------------|--|
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1248 | 2 months | LD ₅₀ | 552 | Mortality | Nebeker and Puglisi (1974) |
| Amphipod (<i>Hyalella azteca</i>) | Water | PCB 52 | > or = 10 weeks | LD ₁₀₀ | 180 | Mortality | Borgmann et al. (1990) |
| Amphipod (<i>Hyalella azteca</i>) | Water | Aroclor 1242 | > or = 10 weeks | LD ₁₀₀ | 100 | Mortality | Borgmann et al. (1990) |
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1242 | 2 months | LD ₅₀ | 316 | Mortality | Nebeker and Puglisi (1974) |
| Cladoceran (<i>Daphnia magna</i>) | Model ecosystem | 2,3,7,8-TCDD | 33 days | EL (no effect) | 1570 | Mortality | Isensee and Jones (1975) |
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1248 | 2 months | LOAEL | 552 | Reproduction reduced by at least 50% | Nebeker and Puglisi (1974) |
| Snail (<i>Physa</i> spp.) | Water | 2,3,7,8-TCDD | 33 days | EL (no effect) | 502 | Mortality | Isensee and Jones (1975) Isensee (1978) |
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1242 | 2 months | EL (effect) | 316 | No reproduction | Nebeker and Puglisi (1974) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 153 | 35 days | LOAEL | 126 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 153 | 35 days | LOAEL | 126 | Weight loss | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 15 | 35 days | LOAEL | 119 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 15 | 35 days | LOAEL | 119 | Weight loss | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 47 | 35 days | LOAEL | 113 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 47 | 35 days | LOAEL | 113 | Weight loss | Fisher et al. (1998) |
| Grass shrimp (<i>Palaemonetes pugio</i>) | Water | Aroclor 1254 | 7 days | LOAEL | 65 | Mortality (60%) | Nimmo et al. (1974) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 1 | 35 days | LOAEL | 64 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 1 | 35 days | LOAEL | 64 | Weight loss | Fisher et al. (1998) |
| Grass shrimp (<i>Palaemonetes pugio</i>) | Water | Aroclor 1254 | 16 days | LOAEL | 27 | Mortality (45%) | Nimmo et al. (1974) |
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1248 | 2 months | NOAEL | 127 | Reproduction | Nebeker and Puglisi (1974) |
| Amphipod (<i>Gammarus pseudolimnaeus</i>) | Water | Aroclor 1242 | 2 months | NOAEL | 76 | Reproduction | Nebeker and Puglisi (1974) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 153 | 35 days | NOAEL | 65 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 153 | 35 days | NOAEL | 65 | Weight loss | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 15 | 35 days | NOAEL | 63.1 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 15 | 35 days | NOAEL | 63.1 | Weight loss | Fisher et al. (1998) |
| Amphipod (<i>Hyalella azteca</i>) | Water | PCB 52 | > or = 10 weeks | NOAEL | 54 | Mortality | Borgmann et al. (1990) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 47 | 35 days | NOAEL | 49.3 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 47 | 35 days | NOAEL | 49.3 | Weight loss | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 1 | 35 days | NOAEL | 33.2 | Mortality | Fisher et al. (1998) |
| Oligochaete (<i>Lumbriculus variegatus</i>) | Algae (Food) | PCB 1 | 35 days | NOAEL | 33.2 | Weight loss | Fisher et al. (1998) |

TABLE 4-4
TOXICITY ENDPOINTS FOR BENTHIC INVERTEBRATES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs, AROCLORS, AND DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EXPOSURE MEDIA | PCB TYPE | EXPOSURE DURATION | EFFECT LEVEL | EFFECT CONC, WHOLE BODY CONC. (mg/kg wet wt) | EFFECT ENDPOINT | REFERENCE |
|---|----------------|--------------|-------------------|--------------|--|-----------------|------------------------|
| Amphipod <i>(Hyalella azteca)</i> | Water | Aroclor 1242 | > or = 10 weeks | NOAEL | 30 | Mortality | Borgmann et al. (1990) |
| Grass shrimp <i>(Palaemonetes pugio)</i> | Water | Aroclor 1254 | 16 days | NOAEL | 18 | Mortality | Nimmo et al. (1974) |
| Grass shrimp <i>(Palaemonetes pugio)</i> | Water | Aroclor 1255 | 7 days | NOAEL | 5.4 | Mortality | Nimmo et al. (1974) |

**TABLE 4-5
TOXICITY ENDPOINTS FOR FISH - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs AND AROCLORS**

| SPECIES | EXPOSURE MEDIA | PCB TYPE | EXPOSURE DURATION | EFFECT LEVEL | EFFECT CONCENTRATION WHOLE BODY CONCENTRATION mg/kg wet wt. | EFFECT ENDPOINT | REFERENCE |
|--|--|--------------|--|--------------|---|-------------------------------------|--------------------------|
| Laboratory studies | | | | | | | |
| Lake trout (<i>Salvelinus namaycush</i>) | Water | PCB-153 | 15 days | LD100 | 7.6 | Fry mortality | Broyles and Noveck, 1979 |
| Chinook salmon (<i>Oncorhynchus</i>) | Water | PCB-153 | 15 days | LD100 | 3.6 | Fry mortality | Broyles and Noveck, 1979 |
| Adult Fathead Minnow (<i>Pimephales promelas</i>) | Water | Aroclor 1254 | 9 months | LOAEL | 999 | Adult mortality | Nebeker et al., 1974 |
| Adult Fathead Minnow (<i>Pimephales promelas</i>) | Water | Aroclor 1254 | 9 months | LOAEL | 429 | Spawning | Nebeker et al., 1974 |
| Adult Minnow (<i>Phoxinus phoxinus</i>) | Diet | Clophen A50 | 40 days; studied for 300 days | LOAEL | 170 | Egg hatchability | Bengtsson, B., 1980 |
| Brook trout fry (<i>Salvelinus fontinalis</i>) | Water | Aroclor 1254 | 118 days | LOAEL | 125 | Fry mortality | Mauck et al., 1978 |
| Brook trout fry (<i>Salvelinus fontinalis</i>) | Water | Aroclor 1254 | 21 days | EL-effect | 32.8 in muscle | Egg hatchability | Freeman and Ilder, 1974 |
| Brook trout fry (<i>Salvelinus fontinalis</i>) | Water | Aroclor 1254 | 21 days | EL-effect | 77.9 in eggs | Egg hatchability | Freeman and Ilder, 1974 |
| Juvenile Spot (<i>Leiostomus xanthurus</i>) | Water | Aroclor 1254 | 20 days | LOAEL | 46 | Adult mortality | Hansen et al., 1971 |
| Adult pinfish (<i>Lagodon rhomboides</i>) | Water | Aroclor 1016 | 42 days | LOAEL | 42 | Adult mortality | Hansen et al., 1974 |
| Killifish (<i>Fundulus heteroclitus</i>) | Single intraperitoneal injection into adults | PCB mixture | Single injection, 40 d of observation | LOAEL | 19 (nominal dose) | Adult female mortality | Black et al., 1998a |
| Lake trout fry (<i>Salmo gairdneri</i>) | Water | Aroclor 1254 | 48 days | EL-effect | 4.5 | Fry mortality | Mac and Seelye, 1981 |
| Killifish (<i>Fundulus heteroclitus</i>) | Single intraperitoneal injection into adults | PCB mixture | Single injection, 40 days of observation | LOAEL | 3.8 (nominal dose) | Egg production and food consumption | Black et al., 1998a |
| Adult Fathead Minnow (<i>Pimephales promelas</i>) | Water | Aroclor 1242 | 9 months | NOAEL | 436 | Adult mortality | Nebeker et al., 1974 |
| Adult Fathead Minnow (<i>Pimephales promelas</i>) | Water | Aroclor 1254 | 9 months | NOAEL | 429 | Egg hatchability | Nebeker et al., 1974 |
| Adult pinfish (<i>Lagodon rhomboides</i>) | Water | Aroclor 1016 | 42 days | NOAEL | 170 | Adult mortality | Hansen et al., 1974 |
| Adult Fathead Minnow (<i>Pimephales promelas</i>) | Water | Aroclor 1254 | 9 months | NOAEL | 105 | Spawning | Nebeker et al., 1974 |
| Brook trout fry (<i>Salvelinus fontinalis</i>) | Water | Aroclor 1254 | 118 days | NOAEL | 71 | Fry mortality | Mauck et al., 1978 |
| Juvenile Spot (<i>Leiostomus xanthurus</i>) | Water | Aroclor 1254 | Lab Stu | NOAEL | 27 | Adult mortality | Hansen et al., 1971 |
| Adult Minnow (<i>Phoxinus phoxinus</i>) | Diet | Clophen A50 | 40 days; studied for 300 days | NOAEL | 15 | Egg hatchability | Bengtsson, B., 1980 |
| Killifish (<i>Fundulus heteroclitus</i>) | Single intraperitoneal injection into adults | PCB mixture | Single injection, 40 days of observation | NOAEL | 3.8 (nominal dose) | Adult female mortality | Black et al., 1998a |
| Killifish (<i>Fundulus heteroclitus</i>) | Single intraperitoneal injection into adults | PCB mixture | Single injection, 40 days of observation | NOAEL | 0.76 (nominal dose) | Egg production and food consumption | Black et al., 1998a |

**TABLE 4-6
TOXICITY ENDPOINTS FOR FISH - FIELD STUDIES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs AND AROCLORS**

| SPECIES | FIELD COMPONENT | CONTAMINANT TYPE | EFFECT LEVEL | EFFECT CONCENTRATION mg/kg wet wt (or as noted below) | EFFECT ENDPOINT | REFERENCE |
|---|---|------------------------------|---------------|---|--|--------------------------------|
| Field studies | | | | | | |
| Arctic charr (<i>Salvelinus alpinus</i>) | Adult fish and eggs collected from Lake Geneva | PCBs DDT | EL-effect | 10 to 78 mg/kg lipid | Embryomortality | Monod, 1985 |
| Winter flounder (<i>Pseudopleuronectes americanus</i>) | Adult and eggs collected from New Bedford Harbor | PCBs | EL-effect | 39.6 mg/kg dry wt in eggs | Growth rate of larvae | Black et al., 1988b |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | PCBs | LOAEL | 29.2 mg/kg dry wt in liver | Embryo and larval survival | Black et al., 1998b |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | PCBs | LOAEL | 20.8 mg/kg dry wt in liver | Adult female mortality | Black et al., 1998b |
| English sole (<i>Parophrys vetulus</i>) | Fish collected from Puget Sound | PCBs, PAHs | EL-effect | Approx. 10 mg/kg in liver | Increased fecundity | Johnson et al., 1997 |
| Striped bass (<i>Morone saxatilis</i>) | Eggs from hatcheries. Larvae fed naturally contaminated | PCBs, HCB, pesticides | EL-effect | 0.1 to 10 in eggs | Larval mortality | Westin et al., 1985 |
| Chinook salmon (<i>Oncorhynchus tshawytscha</i>) | Adult fish and eggs collected from Lake Michigan | PCBs, pesticides | EL-effect | 2.8 to 9.9 A-1254 in eggs | Hatching success | Giesy et al., 1986 |
| Chinook salmon (<i>Oncorhynchus tshawytscha</i>) | Adult fish and eggs collected from Lake Michigan | PCBs | EL-effect | 2.75 to 5.75 in eggs | Hatching success | Ankley et al., 1981 |
| Rainbow trout (<i>Salmo gairdneri</i>) | Adult fish and eggs hatchery | PCBs, DDT | EL-effect | 2.7 in eggs | Embryomortality | Hogan and Braun, 1975 |
| English sole (<i>Parophrys vetulus</i>) | Adults and eggs collected from Puget Sound | PCBs | LOAEL | 2.56 in liver | Production of normal larvae | Casillas et al., 1991 |
| Lake trout (<i>Salvelinus namaycush</i>) | Adult fish and eggs collected from Great Lakes | PCBs | EL-effect | 0.25 to 7.77 in eggs | Egg mortality and percent of normal fry hatching | Mac et al., 1993 |
| Chinook salmon (<i>Oncorhynchus tshawytscha</i>) | Adult fish and eggs collected from Lake Michigan | PCBs, pesticides | EL-effect | 0.322 to 2.6 A-1260 in eggs | Hatching success | Giesy et al., 1986 |
| Starry flounder (<i>Platichthys stellatus</i>) | Adult fish and eggs collected from area of San Francisco Bay | PCBs, HCB, Phthalates | EL-effect | about 50 to 200 in eggs | Hatching success | Spies and Rice, 1988 |
| Redbreast sunfish (<i>Lepomis auritus</i>) | Adult fish collected from East Tennessee stream | PCBs, PAHs, metals, chlorine | EL-effect | 0.95 | Fecundity, clutch size, growth | Adams et al., 1989, 1990, 1992 |
| Baltic herring (<i>Clupea harengus</i>) | Adult fish and eggs collected from Baltic Sea | PCBs, pesticides | EL-effect | > 0.120 in ovaries | Hatching success | Hansen et al., 1985 |
| Baltic flounder (<i>Platichthys flesus</i>) | Adult fish and eggs collected from Baltic Sea | PCBs, pesticides, metals | EL-effect | > 0.120 in ovaries | Hatching success | von Westernhagen et al., 1981 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | PCBs | NOAEL | 9.5 mg/kg dry wt in liver | Embryo and larval mortality | Black et al., 1998b |
| Striped bass (<i>Morone saxatilis</i>) | Eggs from Hudson River fish. Larvae fed naturally contaminated food | PCBs | EL-no effect | 3.1 in post yolk sac larvae | Larval mortality | Westin et al., 1983 |
| Winter flounder (<i>Pseudopleuronectes americanus</i>) | Adult and eggs collected from New Bedford Harbor | PCBs | EL-no effect | 1.08 mg/kg dry wt in eggs | Growth rate of larvae | Black et al., 1988b |
| English sole (<i>Parophrys vetulus</i>) | Adults and eggs collected from Puget Sound | PCBs | NOAEL | 0.09 in liver | Production of normal larvae | Casillas et al., 1991 |
| Redbreast sunfish (<i>Lepomis auritus</i>) | Fish from an East Tennessee stream | PCBs, PAHs, metals, chlorine | EL-no effect | 0.5 | Fecundity, clutch size, growth | Adams et al., 1989, 1990, 1992 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | PCBs | NOAEL | 0.461 mg/kg dry wt in liver | Adult female mortality | Black et al., 1998b |
| Arctic charr (<i>Salvelinus alpinus</i>) | Adult fish and eggs collected from Lake Geneva | PCBs DDT | EL- no effect | 0.1 to 0.31 in eggs | Embryomortality | Monod, 1985 |

**TABLE 4-7
TOXICITY ENDPOINTS FOR FISH - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)**

| SPECIES | EXPOSURE MEDIA | EFFECT LEVEL | TISSUE | CONTAMINANT TYPE | EFFECT CONC. (ug/kg ww) | LIPID CONTENT OF EGG (g lipid/gww egg) | TEF | EFFECT CONC. DIOXIN EQUIVALENTS (ug TEQ/kg lipid) | EFFECT ENDPOINT | REFERENCE |
|--|----------------|--------------|--------|------------------|-------------------------|--|-------|---|----------------------------|--|
| Laboratory studies^a | | | | | | | | | | |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LD50 | Embryo | 2,3,7,8-TCDD | 25.7 | 0.024 | 1 | 1071 | Early life stage mortality | Olivieri and Cooper, 1997 ^b |
| Zebrafish (<i>Danio danio</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 2.61 | 0.017 | 1 | 154 | Early life stage mortality | Elonen et al., 1998 |
| Zebrafish (<i>Danio danio</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 2.5 | 0.017 | 1 | 147 | Early life stage mortality | Henry et al., 1997 |
| White sucker (<i>Catostomus commersoni</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 1.89 | 0.025 | 1 | 76 | Early life stage mortality | Elonen et al., 1998 |
| Northern Pike (<i>Esox lucius</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 2.46 | 0.042 | 1 | 59 | Early life stage mortality | Elonen et al., 1998 |
| Medaka (<i>Oryzias latipes</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 1.11 | 0.029 | 1 | 38 | Early life stage mortality | Elonen et al., 1998 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.539 | 0.024 | 1 | 22 | Early life stage mortality | Elonen et al., 1998 |
| Lake herring (<i>Coregonus artedii</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.902 | 0.066 | 1 | 14 | Early life stage mortality | Elonen et al., 1998 |
| Channel catfish (<i>Ictalurus punctatus</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.644 | 0.048 | 1 | 13 | Early life stage mortality | Elonen et al., 1998 |
| Rainbow Trout (<i>Salmo gairdneri</i>) - Erwin strain | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.439 | 0.087 | 1 | 5.0 | Early life stage mortality | Walker et al., 1992 |
| Rainbow Trout (<i>Salmo gairdneri</i>) - Erwin strain | Injection | LD50 | Egg | 2,3,7,8-TCDD | 0.421 | 0.087 | 1 | 4.8 | Early life stage mortality | Walker et al., 1992 |
| Brook Trout (<i>Salvenius fontinalis</i>) | Water | LD100 | Egg | 2,3,7,8-TCDD | 0.324 | 0.068 | 1 | 4.8 | Early life stage mortality | Walker and Peterson, 1994 |
| Rainbow Trout (<i>Salmo gairdneri</i>) - Erwin strain | Egg injection | LD50 | Egg | 2,3,7,8-TCDD | 0.409 | 0.087 | 1 | 4.7 | Early life stage mortality | Zabel & Peterson, 1996 |
| Rainbow Trout (<i>Salmo gairdneri</i>) | Egg injection | LD50 | Egg | 2,3,7,8-TCDD | 0.374 | 0.087 | 1 | 4.3 | Early life stage mortality | Walker and Peterson, 1991 |
| Rainbow Trout (<i>Salmo gairdneri</i>) | Egg injection | LD50 | Egg | PCB 126 | 74 | 0.087 | 0.005 | 4.3 | Early life stage mortality | Walker and Peterson, 1991 |
| Brook Trout (<i>Salvenius fontinalis</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.200 | 0.068 | 1 | 2.9 | Early life stage mortality | Walker and Peterson, 1994 |
| Rainbow Trout (<i>Salmo gairdneri</i>) | Egg injection | LD50 | Egg | 2,3,7,8-TCDD | 0.242 | 0.087 | 1 | 2.8 | Early life stage mortality | Zabel & Peterson, 1996 |
| Lake trout (<i>Salvenius namaycush</i>) | Water | LD50 | Egg | PCB 126 | 29 | 0.08 | 0.005 | 1.8 | Early life stage mortality | Zabel et al., 1995 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LD50 | Embryo | 2,3,7,8-TCDD | 0.026 | 0.024 | 1 | 1.1 | Early life stage mortality | Olivieri and Cooper, 1997 |
| Lake trout (<i>Salvenius namaycush</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.085 | 0.08 | 1 | 1.1 | Early life stage mortality | Zabel et al., 1995 |
| Lake trout (<i>Salvenius namaycush</i>) | Water | LD50 | Egg | 2,3,7,8-TCDD | 0.065 | 0.08 | 1 | 0.8 | Early life stage mortality | Walker et al., 1992 |
| Lake trout (<i>Salvenius namaycush</i>) | Injection | LD50 | Egg | 2,3,7,8-TCDD | 0.047 | 0.08 | 1 | 0.6 | Early life stage mortality | Walker et al., 1992 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LD100 | Larvae | 2,3,7,8-TCDD | 163 | Not reported for larvae | 1 | | Early life stage mortality | Olivieri and Cooper, 1997 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LD50 | Larvae | 2,3,7,8-TCDD | 70.9 | Not reported for larvae | 1 | | Early life stage mortality | Olivieri and Cooper, 1997 |

**TABLE 4-7
TOXICITY ENDPOINTS FOR FISH - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)**

| SPECIES | EXPOSURE MEDIA | EFFECT LEVEL | TISSUE | CONTAMINANT TYPE | EFFECT CONC. (ug/kg ww) | LIPID CONTENT OF EGG (g lipid/gww egg) | TEF | EFFECT CONC. DIOXIN EQUIVALENTS (ug TEQ/kg lipid) | EFFECT ENDPOINT | REFERENCE |
|---|-------------------|--------------|--------|------------------|-------------------------|--|-----|---|----------------------------|---------------------------|
| Zebrafish (<i>Danio danio</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 2 | 0.017 | 1 | 118 | Early life stage mortality | Elonen et al., 1998 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LOAEL | Embryo | 2,3,7,8-TCDD | 2.46 | 0.024 | 1 | 103 | Early life stage mortality | Olivieri and Cooper, 1997 |
| White sucker (<i>Catostomus commersoni</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 1.22 | 0.025 | 1 | 49 | Early life stage mortality | Elonen et al., 1998 |
| Northern Pike (<i>Esox lucius</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 1.8 | 0.042 | 1 | 43 | Early life stage mortality | Elonen et al., 1998 |
| Medaka (<i>Oryzias latipes</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.949 | 0.029 | 1 | 33 | Early life stage mortality | Elonen et al., 1998 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.435 | 0.024 | 1 | 18 | Early life stage mortality | Elonen et al., 1998 |
| Channel catfish (<i>Ictalurus punctatus</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.855 | 0.048 | 1 | 18 | Early life stage mortality | Elonen et al., 1998 |
| Lake herring (<i>Coregonus artedii</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.27 | 0.066 | 1 | 4.1 | Early life stage mortality | Elonen et al., 1998 |
| Rainbow Trout (<i>Salmo gairderi</i>) | Injection | LOAEL | Egg | 2,3,7,8-TCDD | 0.291 | 0.087 | 1 | 3.3 | Early life stage mortality | Walker et al., 1992 |
| Rainbow Trout (<i>Salmo gairderi</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.279 | 0.087 | 1 | 3.2 | Early life stage mortality | Walker et al., 1992 |
| Brook Trout (<i>Salvenius fontinalis</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.185 | 0.068 | 1 | 2.7 | Early life stage mortality | Walker and Peterson, 1994 |
| Lake trout (<i>Salvelinus namaycush</i>) | Injection | LOAEL | Egg | 2,3,7,8-TCDD | 0.058 | 0.08 | 1 | 0.7 | Early life stage mortality | Walker et al., 1992 |
| Lake trout (<i>Salvelinus namaycush</i>) | Injection | LOAEL | Egg | 2,3,7,8-TCDD | 0.055 | 0.08 | 1 | 0.7 | | Walker et al., 1994 |
| Lake trout (<i>Salvelinus namaycush</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.055 | 0.08 | 1 | 0.7 | Early life stage mortality | Walker et al., 1992 |
| Lake trout (<i>Salvelinus namaycush</i>) | Maternal transfer | LOAEL | Egg | 2,3,7,8-TCDD | 0.05 | 0.08 | 1 | 0.6 | | Walker et al., 1994 |
| Lake trout (<i>Salvelinus namaycush</i>) | Water | LOAEL | Egg | 2,3,7,8-TCDD | 0.04 | 0.08 | 1 | 0.5 | | Walker et al., 1994 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | LOAEL | Larvae | 2,3,7,8-TCDD | 20 | Not reported for larvae | 1 | | Early life stage mortality | Olivieri and Cooper, 1997 |
| White sucker (<i>Catostomus commersoni</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.848 | 0.025 | 1 | 34 | Early life stage mortality | Elonen et al., 1998 |
| Northern Pike (<i>Esox lucius</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 1.19 | 0.042 | 1 | 28 | Early life stage mortality | Elonen et al., 1998 |
| Zebrafish (<i>Danio danio</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.424 | 0.017 | 1 | 25 | Early life stage mortality | Elonen et al., 1998 |
| Medaka (<i>Oryzias latipes</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.455 | 0.029 | 1 | 16 | Early life stage mortality | Elonen et al., 1998 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.235 | 0.024 | 1 | 9.8 | Early life stage mortality | Elonen et al., 1998 |
| Channel catfish (<i>Ictalurus punctatus</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.385 | 0.048 | 1 | 8.0 | Early life stage mortality | Elonen et al., 1998 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | NOAEL | Embryo | 2,3,7,8-TCDD | 0.13 | 0.024 | 1 | 5.4 | Early life stage mortality | Olivieri and Cooper, 1997 |
| Lake herring (<i>Coregonus artedii</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.175 | 0.066 | 1 | 2.7 | Early life stage mortality | Elonen et al., 1998 |
| Rainbow Trout (<i>Salmo gairderi</i>) | Injection | NOAEL | Egg | 2,3,7,8-TCDD | 0.291 | 0.087 | 1 | 3.3 | Early life stage mortality | Walker et al., 1992 |
| Brook Trout (<i>Salvenius fontinalis</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.135 | 0.068 | 1 | 2.0 | Early life stage mortality | Walker and Peterson, 1994 |

**TABLE 4-7
TOXICITY ENDPOINTS FOR FISH - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)**

| SPECIES | EXPOSURE MEDIA | EFFECT LEVEL | TISSUE | CONTAMINANT TYPE | EFFECT CONC. (ug/kg ww) | LIPID CONTENT OF EGG (g lipid/gww egg) | TEF | EFFECT CONC. DIOXIN EQUIVALENTS (ug TEQ/kg lipid) | EFFECT ENDPOINT | REFERENCE |
|--|-------------------|--------------|--------|------------------|-------------------------|--|-----|---|----------------------------|---------------------------|
| Lake trout (<i>Salvelinus namaycush</i>) | Injection | NOAEL | Egg | 2,3,7,8-TCDD | 0.044 | 0.08 | 1 | 0.55 | Early life stage mortality | Walker et al., 1992 |
| Lake trout (<i>Salvelinus namaycush</i>) | Injection | NOAEL | Egg | 2,3,7,8-TCDD | 0.044 | 0.08 | 1 | 0.55 | | Walker et al., 1994 |
| Lake trout (<i>Salvelinus namaycush</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.034 | 0.08 | 1 | 0.43 | Early life stage mortality | Walker et al., 1992 |
| Lake trout (<i>Salvelinus namaycush</i>) | Water | NOAEL | Egg | 2,3,7,8-TCDD | 0.034 | 0.08 | 1 | 0.43 | | Walker et al., 1994 |
| Lake trout (<i>Salvelinus namaycush</i>) | Maternal transfer | NOAEL | Egg | 2,3,7,8-TCDD | 0.023 | 0.08 | 1 | 0.29 | | Walker et al., 1994 |
| Fathead minnow (<i>Pimephales promelas</i>) | Water | NOAEL | Larvae | 2,3,7,8-TCDD | 3.59 | Not reported for larvae | 1 | | Early life stage mortality | Olivieri and Cooper, 1997 |

Notes:

^a No relevant field studies were found.

^b Fathead minnow embryo is assumed to have same lipid content as reported for eggs.

**TABLE 4-8
TOXICITY ENDPOINTS FOR FISH - FIELD STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)**

| SPECIES | EXPOSURE MEDIA | EFFECT LEVEL | TISSUE | CONTAMINANT TYPE | EFFECT CONC. (ug/kg ww, unless noted differently) | LIPID CONTENT OF EGG (g lipid/gww egg) | EFFECT CONC. (ug/kg lipid) | TEF | EFFECT CONC. DIOXIN EQUIVALENTS (ug TEQ/kg lipid) | EFFECT ENDPOINT | REFERENCE |
|--|--|--------------|--------|------------------|---|--|----------------------------|-----|---|----------------------------|--------------------------|
| Rainbow Trout - Arlee strain (<i>Salmo gairdneri</i>) | Egg injection or extract from field-collected fish | LD50 | Eggs | TEQs | 0.514 | 0.087 | 5.9 | 1 | 5.9 | Embryomortality | Wright and Tillitt, 1999 |
| Rainbow Trout - Erwin strain (<i>Salmo gairdneri</i>) | Egg injection or extract from field-collected fish | LD50 | Eggs | TEQs | 0.206 | 0.087 | 2.4 | 1 | 2.4 | Embryomortality | Wright and Tillitt, 1999 |
| Rainbow Trout - Lake Superior (<i>Salmo gairdneri</i>) | Egg injection or extract from field-collected fish | LD50 | Eggs | TEQs | 1.43 | 0.087 | 16.4 | 1 | 16.4 | Embryomortality | Wright and Tillitt, 1999 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | LOAEL | Liver | TEQs | 1.56 ug/kg dry wt | Not available | Not available | 1 | Not available | Embryo and larval survival | Black et al., 1998 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | LOAEL | Liver | TEQs | 0.543 ug/kg dry wt | Not available | Not available | 1 | Not available | Adult female mortality | Black et al., 1998 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | NOAEL | Liver | TEQs | 0.132 ug/kg dry wt | Not available | Not available | 1 | Not available | Embryo and larval survival | Black et al., 1998 |
| Lake trout (<i>Salvelinus namaycush</i>) | Fish collected from Lake Ontario | EL-no effect | Eggs | TEQs | 0.011 | 0.08 | 0.1 | 1 | 0.1 | Early life stage mortality | Guiney et al., 1996 |
| Killifish (<i>Fundulus heteroclitus</i>) | Fish collected from New Bedford Harbor | NOAEL | Liver | TEQs | 0.00572 ug/kg dry wt | Not available | Not available | 1 | Not available | Adult female mortality | Black et al., 1998 |

**TABLE 4-9
TOXICITY ENDPOINTS FOR AVIANS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS**

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|--|---------------------------|----------------------------|--------------|--------------|----------------------------|------------------------------|--|-----------------------------|
| Laboratory studies | | | | | | | | |
| Mallard Duck (<i>Anas platyrhynchos</i>) | | 5 day | LD50 | Aroclor 1254 | 853 | 8122 | Mortality | Hill et al., 1975 |
| Japanese Quail (<i>Coturnix coturnix</i>) | | 5 day | LD50 | Aroclor 1254 | 759 | 6737 | Mortality | Hill et al., 1975 |
| Bobwhite Quail (<i>Colinus virginianus</i>) | | 5 day | LD50 | Aroclor 1254 | 141 | 1516 | Mortality | Hill et al., 1975 |
| Brown-headed Cowbird (<i>Molothrus ater</i>) | Diet | 7 days | EL-effect | Aroclor 1254 | 333 | 1500 | Mortality | Stickel et al., 1984 |
| Red-winged Blackbird (<i>Agelaius phoeniceus</i>) | Diet | 6 days | EL-effect | Aroclor 1254 | 321 | 1500 | Mortality | Stickel et al., 1984 |
| Japanese Quail (<i>Coturnix coturnix</i>) | Oral by syringe | 7 days | LOAEL | Aroclor 1260 | 100 | 888 | Weight loss | Vos et al., 1971 |
| Mallard Duck (<i>Anas platyrhynchos</i>) | Diet | 12 weeks | EL-effect | Aroclor1242 | 16 | 150 | Decreased weight gain in hens, eggshell thinning | Haseltine and Prouty, 1980 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Drinking water | 6 weeks | EL-effect | Aroclor 1254 | 3.5 | 50 | Hatching success | Tumasonis et al., 1973 |
| Ring-Necked Pheasant (<i>Phasianus colchicus</i>) | Diet, in gelatin capsules | Once per week for 17 weeks | LOAEL | Aroclor 1254 | 2.9 | 50 | Egg production | Dahlgren et al., 1972 |
| Ring-Necked Pheasant (<i>Phasianus colchicus</i>) | Diet | Not available | LOAEL | Aroclor 1254 | 2.9 | 50 | Female fertility | Roberts et al., 1978 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor 1242 | 1.4 | 20 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor 1248 | 1.4 | 20 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor 1254 | 1.4 | 20 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor1242 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor 1254 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | LOAEL | Aroclor1248 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Ringed Turtle Dove (<i>Streptopelia risoria</i>) | Diet | 3 months | EL-effect | Aroclor 1254 | 1.1 | 10 | Hatching success | Peakall et al, 1972 |
| Ringed Turtle Dove (<i>Streptopelia risoria</i>) | Diet | | LOAEL | Aroclor 1254 | 1.1 | 10 | Hatching success | Peakall and Peakall, 1973 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 6 weeks | LOAEL | Aroclor 1242 | 0.7 | 10 | Hatching success | Britton and Huston, 1973 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | LOAEL | Aroclor 1242 | 0.7 | 10 | Hatching success | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | LOAEL | Aroclor 1248 | 0.7 | 10 | Hatching success | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | LOAEL | Aroclor 1248 | 0.7 | 10 | Hatching success | Scott, 1977 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | | LOAEL | Aroclor 1254 | 0.3 | 5 | Fertility and egg production | Platonow and Reinhart, 1973 |

**TABLE 4-9
TOXICITY ENDPOINTS FOR AVIANS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS**

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|---|---------------------------|----------------------------|--------------|--------------|----------------------------|------------------------------|---|----------------------------|
| Laboratory studies | | | | | | | | |
| European Starling (<i>Sternus vulgaris</i>) | Diet | 4 days | EL-effect | Aroclor 1254 | Not available | 1,500 | Mortality | Stickel et al., 1984 |
| Common Grackle (<i>Quiscalus quiscula</i>) | Diet | 8 days | EL-effect | Aroclor 1254 | Not available | 1,500 | Mortality | Stickel et al., 1984 |
| Mallard Duck (<i>Anas platyrhynchos</i>) | Diet | 12 weeks | EL-no effect | Aroclor1242 | 16 | 150 | Reproduction success, hatching success, survival and growth of chicks | Haseltine and Prouty, 1980 |
| Japanese Quail (<i>Coturnix coturnix</i>) | Diet | 14 weeks | EL-no effect | Aroclor 1254 | 5.6 | 50 | Mortality and growth rates of adults | Chang and Stokstad, 1975 |
| Mallard Duck (<i>Anas platyrhynchos</i>) | Diet | Approx. 1 month | EL-no effect | Aroclor 1254 | 2.6 | 25 | Reproduction success | Custer and Heinz, 1980 |
| Japanese Quail (<i>Coturnix coturnix</i>) | Diet | Not reported | NOAEL | Aroclor1248 | 2.3 | 20 | Hatching success | Scott, 1977 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | NOAEL | Aroclor 1016 | 1.4 | 20 | Egg production | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | NOAEL | Aroclor 1254 | 1.4 | 20 | Egg production | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | EL-no effect | Aroclor1221 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | EL-no effect | Aroclor1232 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | EL-no effect | Aroclor1268 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | EL-no effect | Aroclor 5442 | 1.4 | 20 | Hatching success | Cecil et al., 1974 |
| Ring-Necked Pheasant (<i>Phasianus colchicus</i>) | Diet, in gelatin capsules | Once per week for 17 weeks | NOAEL | Aroclor 1254 | 0.7 | 12.5 | Egg production | Dahlgren et al., 1972 |
| Screech Owl (<i>Otus asio</i>) | Diet | > 8 weeks | EL-no effect | Aroclor1248 | 0.4 | 3 | Egg production, hatching success, fledging success | McLane and Hughes, 1980 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 6 weeks | NOAEL | Aroclor1242 | 0.3 | 5 | Hatching success | Britton and Huston, 1973 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | NOAEL | Aroclor 1242 | 0.3 | 5 | Hatching success | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | NOAEL | Aroclor1248 | 0.3 | 5 | Hatching success | Lillie et al., 1975 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor 1242 | 0.1 | 2 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor 1248 | 0.1 | 2 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor 1254 | 0.1 | 2 | Egg production, hatching success, chick growth | Lillie et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor1242 | 0.1 | 2 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor1248 | 0.1 | 2 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 9 weeks | NOAEL | Aroclor 1254 | 0.1 | 2 | Hatching success | Cecil et al., 1974 |
| Domestic Chicken (<i>Gallus domesticus</i>) | Diet | 8 weeks | NOAEL | Aroclor1248 | 0.1 | 1 | Hatching success | Scott, 1977 |

TABLE 4-10
TOXICITY ENDPOINTS FOR AVIANS - FIELD STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS

| SPECIES | FIELD COMPONENT | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE FOOD CONC. | EFFECT ENDPOINT | REFERENCE |
|--|--|--------------|------------------|----------------------------|----------------------|---------------------------------|--------------------------------|
| Field studies | | | | | | | |
| Tree Swallow (<i>Tachycineta bicolor</i>) | Populations in Fox River and Green Bay, Lake Michigan, studied | NOAEL | PCBs, DDE | 0.55 | up to 0.61 | Clutch and egg success | Custer et al., 1998 |
| Tree Swallow (<i>Tachycineta bicolor</i>) | Populations along Hudson River studied | NOAEL | PCBs | 16.1 | up to 17.9 | Growth, mortality, reproduction | US EPA Phase 2 Database (1998) |

TABLE 4-11
TOXICITY ENDPOINTS FOR AVIANS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE DIOXIN EQUIVALENTS (ug/kg/day) | EFFECT ENDPOINT | REFERENCE |
|---|-----------------|-------------------|-------------------|------------------|---|-----------------------------|----------------------|
| Laboratory studies^a | | | | | | | |
| Ringed turtle dove (<i>Streptopelia risoria</i>) | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | > 810 | Mortality | Hudson et al., 1984 |
| Mallard (<i>Anas platyrhynchos</i>) | Oral | Single dose | | 2,3,7,8-TCDD | > 108 | Mortality | Hudson et al., 1984 |
| Chicken (<i>Gallus domesticus</i>) | Oral | 21 days | LD ₁₀₀ | 2,3,7,8-TCDD | 25 - 50 | Mortality | Greig et al., 1973 |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Intraperitoneal | Single dose | LD ₇₅ | 2,3,7,8-TCDD | 25 | Mortality | Nosek et al., 1992 |
| Northern bobwhite quail (<i>Colinus virginianus</i>) | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | 15 | Mortality | Hudson et al., 1984 |
| Chicken (<i>Gallus domesticus</i>) | Oral | 21 days | LOAEL | 2,3,7,8-TCDD | 1.0 | Mortality | Schwetz et al., 1973 |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Intraperitoneal | 10 weeks | LOAEL | 2,3,7,8-TCDD | 0.14 | Fertility, embryo mortality | Nosek et al., 1992 |
| Chicken (<i>Gallus domesticus</i>) | Oral | 21 days | NOAEL | 2,3,7,8-TCDD | 0.1 | Mortality | Schwetz et al., 1973 |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Intraperitoneal | 10 weeks | NOAEL | 2,3,7,8-TCDD | 0.014 | Fertility, embryo mortality | Nosek et al., 1992 |

Notes:

^a No relevant field studies were found.

Note units of ug/kg/day.

TABLE 4-12
TOXICITY ENDPOINTS FOR AVIANS - FIELD STUDIES
EFFECTIVE DIETARY DOSES OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | FIELD COMPONENT | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE DIOXIN EQUIVALENTS (ug/kg/day) | EFFECTIVE FOOD CONC. (ug/kg) | EFFECT ENDPOINT | REFERENCE |
|--|--|--------------|------------------|---|------------------------------|---------------------------------|-------------------------------|
| Field studies | | | | | | | |
| Tree Swallow (<i>Tachycineta bicolor</i>) | Populations along Hudson River studied | EL-no effect | TEQs | 4.9 | up to 5.41 | Growth, mortality, reproduction | US EPA Phase 2 Database, 1998 |
| Tree Swallow (<i>Tachycineta bicolor</i>) | Populations in Fox River and Green Bay, Lake Michigan, studied | EL-no effect | TEQs, DDE | 0.08 | up to 0.091 | Clutch and egg success | Custer et al., 1998 |

TABLE 4-13
TOXICITY ENDPOINTS FOR AVIAN EGGS - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs AND AROCLORS

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE EGG CONC. (mg/kg egg) | EFFECT ENDPOINT | REFERENCE |
|---|----------------|-------------------|--------------|--------------|---------------------------------|--|-----------------------------|
| Laboratory studies | | | | | | | |
| Chicken (<i>Gallus domesticus</i>) | Drinking water | 6 weeks | EL-effect | Aroclor 1254 | > 10-15 ppm in yolk | Deformities | Tumasonis et al., 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | LOAEL | Aroclor 1260 | 10 | Growth rate of chicks | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | LOAEL | Aroclor 1254 | 6.7 | Growth and mortality of embryos | Gould et al., 1997 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | LOAEL | Aroclor 1242 | 5 | Hatching success | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | LOAEL | Aroclor 1254 | 5 | Hatching success | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | LOAEL | Aroclor 1242 | 5 | Growth rate of chicks | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | | | LOAEL | | 5 | Egg production and hatching success | Platanow and Reinhart, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Diet | 6 weeks | LOAEL | Aroclor 1242 | 3.7 | Hatching success | Britton and Huston, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Diet | 4 weeks | LOAEL | Aroclor 1248 | 2.21 | Hatching success | Scott, 1977 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1260 | 10 | Hatching success | Carlson and Duby, 1973 |
| Screech owl (<i>Otus asio</i>) | Diet of hens | > 8 weeks | NOAEL | Aroclor 1248 | 7.1 | Egg production, hatching success, and fledging success | McLane and Hughes, 1980 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1260 | 5 | Growth rate of chicks | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1242 | 2.5 | Hatching success | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1254 | 2.5 | Hatching success | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1242 | 2.5 | Growth rate of chicks | Carlson and Duby, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Diet | 6 weeks | NOAEL | Aroclor 1242 | 1.7 | Hatching success | Britton and Huston, 1973 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | | NOAEL | Aroclor 1254 | 0.67 | Growth and mortality of embryos | Gould et al., 1997 |
| Chicken (<i>Gallus domesticus</i>) | Diet | 4 weeks | NOAEL | Aroclor 1248 | 0.33 | Hatching success | Scott, 1977 |

**TABLE 4-14
TOXICITY ENDPOINTS FOR AVIAN EGGS - FIELD STUDIES
EFFECTIVE CONCENTRATIONS OF TOTAL PCBs AND AROCLORS**

| SPECIES | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE EGG CONC. (mg/kg egg) | EFFECT ENDPOINT | REFERENCE |
|--|------------------|-----------------------------------|---------------------------------|--|---|
| Field studies | | | | | |
| Bald eagle (<i>Haliaeetus leucocephalus</i>) | EL--Effect level | PCBs, Pesticides | 20-54 | Reproductive success | Clark et al., 1988 |
| Double-crested cormorant (<i>Phalacrocorax auritus</i>) | EL-Effect level | PCBs, Pesticides, Hg | 23.8 | Hatching success and fledging success | Weseloh et al., 1983 |
| Caspian tern (<i>Hydropogne caspia</i>) | EL-Effect level | PCBs, Pesticides | 4.2 - 18 | Increased rate of embryo deformities | Yamashita et al., 1993 |
| Forster's tern (<i>Sterna forsteri</i>) | LOAEL | PCBs, Pesticides, Dioxins, Furans | 22.2 | Hatching success | Kubiak et al., 1989 |
| Common tern (<i>Sterna hirundo</i>) | LOAEL | PCBs, Pesticides, Hg | 7 | Hatching success | Becker et al., 1993 |
| Common tern (<i>Sterna hirundo</i>) | LOAEL | PCBs, Pesticides, Hg | 9.8 | Hatching success | Hoffman et al., 1993 |
| Bald eagle (<i>Haliaeetus leucocephalus</i>) | LOAEL | PCBs, Pesticides, Hg | 3 - 5.6 | 10 % reduction in reproductive success | Wiemeyer et al., 1984, 1993 |
| Bald eagle (<i>Haliaeetus leucocephalus</i>) | EL- No Effect | PCBs, TEQs, Pesticides | 33.2 - 64 in yolk sac | Hatching success | Elliott et al., 1996 Secord and McCarty, 1997, McCarty and Secord, 1999, U.S. EPA Phase 2 Database Release 4.1b. |
| Tree swallow (<i>Tachycineta bicolor</i>) | NOAEL | PCBs | 26.7 | Reproductive output | U.S. EPA Phase 2 Database Release 4.1b. |
| Common tern (<i>Sterna hirundo</i>) | NOAEL | PCBs, Pesticides, Hg | 6.7 | Hatching success | Hoffman et al., 1993 |
| Common tern (<i>Sterna hirundo</i>) | NOAEL | PCBs, Pesticides, Hg | 5.2 | Hatching success | Becker et al., 1993 |
| Forster's tern (<i>Sterna forsteri</i>) | NOAEL | PCBs, Pesticides, Dioxins, Furans | 4.5 | Hatching success | Kubiak et al., 1989 |
| Tree swallow (<i>Tachycineta bicolor</i>) | NOAEL | PCBs, DDE | 3.24 in eggs and pippers | Clutch success, egg success | Custer et al., 1998 |
| Bald eagle (<i>Haliaeetus leucocephalus</i>) | NOAEL | PCBs, Pesticides, Hg | < 3 | Reproductive success | Wiemeyer et al., 1984, 1993 |

**TABLE 4-15
TOXICITY ENDPOINTS FOR AVIAN EGGS - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)**

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE EGG CONC. (ug/kg egg) | TEF | EFFECTIVE EGG CONC. DIOXIN EQUIVALENTS (ug-TEQ/kg-egg) | EFFECT ENDPOINT | REFERENCE |
|--|----------------|--------------------|--------------|------------------|---------------------------------|--------|--|------------------|-------------------------|
| Laboratory studies | | | | | | | | | |
| American kestrel (<i>Falco sparverius</i>) | Egg injection | 18 days | LD50 | PCB 77 | 316 | 0.05 | 16 | Embryo mortality | Hoffman et al., 1998 |
| cormorant (<i>Phalacrocorax auritus</i>) | Egg injection | 21 days | LD50 | PCB 126 | 158 | 0.1 | 16 | Embryo mortality | Powell et al., 1997 |
| Common tern (<i>Sterna hirundo</i>) | Egg injection | 18 days | LD50 | PCB 126 | 104 | 0.1 | 10 | Embryo mortality | Hoffman et al., 1998 |
| American kestrel (<i>Falco sparverius</i>) | Egg injection | 20 days | LD50 | PCB 126 | 65 | 0.1 | 7 | Embryo mortality | Hoffman et al., 1998 |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Egg injection | 28 days | LD50 | 2,3,7,8-TCDD | 1.35 | 1 | 1 | Embryo mortality | Nosek et al., 1993 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LD50 | PCB 105 | 5592 | 0.0001 | 1 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LD50 | PCB 77 | 8.8 | 0.05 | 0.4 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | LD50 | PCB 126 | 2.3 | 0.1 | 0.2 | Embryo mortality | Powell et al., 1996a |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | LD50 | 2,3,7,8-TCDD | 0.15 | 1 | 0.2 | Embryo mortality | Powell et al., 1996a |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 20 days | LD50 | PCB 77 | 2.6 | 0.05 | 0.1 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 18 days | LD50 | PCB 126 | 0.4 | 0.1 | 0.04 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LD50 | PCB 126 | 0.6 | 0.1 | 0.1 | Embryo mortality | Powell et al., 1996b |
| cormorant (<i>Phalacrocorax auritus</i>) | Egg injection | 21 days | LOAEL | PCB 126 | 800 | 0.1 | 80 | Embryo mortality | Powell et al., 1997 |
| American kestrel (<i>Falco sparverius</i>) | Egg injection | 20 days | LOAEL | PCB 126 | 233 | 0.1 | 23 | Embryo mortality | Hoffman et al., 1998 |
| American kestrel (<i>Falco sparverius</i>) | Egg injection | 20 days | LOAEL | PCB 77 | 100 | 0.05 | 5 | Embryo mortality | Hoffman et al., 1998 |
| Common tern (<i>Sterna hirundo</i>) | Egg injection | 18 days | LOAEL | PCB 126 | 44 | 0.1 | 4 | Embryo mortality | Hoffman et al., 1998 |
| Double-crested cormorant | Egg injection | 21 days | LOAEL | 2,3,7,8-TCDD | 4 | 1 | 4 | Embryo mortality | Powell et al., 1997 |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Egg injection | 21 days | LOAEL | 2,3,7,8-TCDD | 1 | 1 | 1.0 | Embryo mortality | Nosek et al., 1993 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LOAEL | PCB 105 | 8100 | 0.0001 | 1 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LOAEL | PCB 77 | 9 | 0.05 | 0.5 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 18 days | LOAEL | PCB 77 | 6 | 0.05 | 0.3 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | LOAEL | 2,3,7,8-TCDD | 0.16 | 1 | 0.2 | Embryo mortality | Powell et al., 1996a |
| Pidgeon (<i>Columba livia</i>) | Egg injection | Embryonic Day 3 | EL-Effect | 2,3,7,8-TCDD | 1 | 1 | 1.0 | Hatchability | Janz and Bellward, 1996 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | LOAEL | PCB 126 | 0.9 | 0.1 | 0.09 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 18 days | LOAEL | PCB 126 | 0.5 | 0.1 | 0.05 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | LOAEL | PCB 126 | 0.2 | 0.1 | 0.02 | Embryo mortality | Powell et al., 1996a |
| Double-crested cormorant | Egg injection | 21 days | NOAEL | PCB 126 | 400 | 0.1 | 40 | Embryo mortality | Powell et al., 1997 |
| Great Blue Heron (<i>Ardea herodias</i>) | Egg injection | Embryonic Day 9 | EL-No effect | 2,3,7,8-TCDD | 2 | 1 | 2 | Hatchability | Janz and Bellward, 1996 |
| American kestrel (<i>Falco sparverius</i>) | Egg injection | 20 days | NOAEL | PCB 126 | 23 | 0.1 | 2 | Embryo mortality | Hoffman et al., 1998 |
| Double-crested cormorant | Egg injection | 21 days | NOAEL | 2,3,7,8-TCDD | 1 | 1 | 1 | Embryo mortality | Powell et al., 1997 |

TABLE 4-15
TOXICITY ENDPOINTS FOR AVIAN EGGS - LABORATORY STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE EGG CONC. (ug/kg egg) | TEF | EFFECTIVE EGG CONC. DIOXIN EQUIVALENTS (ug TEQ/kg egg) | EFFECT ENDPOINT | REFERENCE |
|--|----------------|-------------------|--------------|------------------|---------------------------------|--------|--|------------------|-------------------------|
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | NOAEL | PCB 105 | 2700 | 0.0001 | 0.3 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | NOAEL | PCB 77 | 3 | 0.05 | 0.2 | Embryo mortality | Powell et al., 1996b |
| Ring-necked pheasant (<i>Phasianus colchicus</i>) | Egg injection | 28 days | NOAEL | 2,3,7,8-TCDD | 0.1 | 1 | 0.1 | Embryo mortality | Nosek et al., 1993 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | NOAEL | 2,3,7,8-TCDD | 0.08 | 1 | 0.1 | Embryo mortality | Powell et al., 1996a |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 18 days | NOAEL | PCB 77 | 1.2 | 0.05 | 0.1 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus gallus</i>) | Egg injection | Embryonic Day 4 | EL-No effect | 2,3,7,8-TCDD | 0.1 | 1 | 0.1 | Hatchability | Janz and Bellward, 1996 |
| Chicken (<i>Gallus gallus</i>) | Egg injection | 18 days | NOAEL | PCB 126 | 0.3 | 0.1 | 0.03 | Embryo mortality | Hoffman et al., 1998 |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 18 days | NOAEL | PCB 126 | 0.3 | 0.1 | 0.03 | Embryo mortality | Powell et al., 1996b |
| Chicken (<i>Gallus domesticus</i>) | Egg injection | 24 days | NOAEL | PCB 126 | 0.1 | 0.1 | 0.01 | Embryo mortality | Powell et al., 1996a |

TABLE 4-16
TOXICITY ENDPOINTS FOR AVIAN EGGS - FIELD STUDIES
EFFECTIVE CONCENTRATIONS OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE EGG CONC. DIOXIN EQUIVALENTS (ug TEQ/kg egg) | EFFECT ENDPOINT | REFERENCE |
|---|-----------------|------------------|--|---|--|
| Field studies | | | | | |
| Osprey (<i>Pandion haliaeetus</i>) | EL-Effect level | TCDD | 29 - 162 | Growth rate of chicks | Woodford, et al., 1998 |
| Bald eagle (<i>Haliaeetus leucocephalus</i>) | EL-Effect level | TEQs, DDE | 0.51-1.2 | Reproductive success | Clark et al., 1998 |
| Great blue heron (<i>Ardea herodias</i>) | LOAEL | TEQs | 0.5 | Growth rate | Sanderson et al., 1994 |
| Great blue heron (<i>Ardea herodias</i>) | EL-Effect level | TEQs, pesticides | 0.5 | Growth rate | Hart et al., 1991 |
| Cormorant (<i>Phalacrocorax auritus</i>) | EL-effect level | TEQ | 0.035 - 0.344 | Egg mortality | Tillitt et al., 1992 |
| Great blue heron (<i>Ardea herodias</i>) | EL-Effect level | TEQs, pesticides | 0.23 | Reproductive success | Elliott et al., 1989 |
| Forster's tern (<i>Sterna forsteri</i>) | EL-Effect | TEQs, pesticides | 2.20 | Hatching success, growth rate of chicks | Kubiak et al., 1989 |
| Forster's tern (<i>Sterna forsteri</i>) | EL-Effect level | TEQ | 0.21 | Hatching success | Tillitt et al., 1993 |
| Wood duck (<i>Aix sponsa</i>) | LOAEL | TEQs, pesticides | 0.02 | Nest success, hatching success, duckling production | White and Seginak, 1994; White and Hoffman, 1995 |
| Tree swallow (<i>Tachycineta bicolor</i>) | NOAEL | TEQs | 13 | Reproductive success | US EPA Phase 2 Database (1998) |
| Tree swallow (<i>Tachycineta bicolor</i>) | EL-No effect | TEQs | 0.589 in pippers | Reproductive success | Custer et al., 1998 |
| Great blue heron (<i>Ardea herodias</i>) | NOAEL | TEQs | 0.3 | Reduced body weight | Sanderson et al., 1994 |
| Great blue heron (<i>Ardea herodias</i>) | NOAEL | TEQs | 0.24 | Growth rate | Hart et al., 1991 |
| Forster's tern (<i>Sterna forsteri</i>) | EL-no effect | TEQs, pesticides | 0.2 | Hatchability, growth rate of chicks | Kubiak et al., 1989 |
| Great blue heron (<i>Ardea herodias</i>) | EL-No effect | TEQs, pesticides | 0.079 | Reproductive success | Elliott et al., 1989 |
| Osprey (<i>Pandion haliaeetus</i>) | EL-no effect | TCDD, TEQs | ND - 23.8 | Growth rate of chicks | Woodford et al., 1998 |
| Osprey (<i>Pandion haliaeetus</i>) | EL-no effect | TEQs | 0.136 | Embryo survival | Woodford et al., 1998 |
| Foster's tern (<i>Sterna forsteri</i>) | EL-no effect | TEQs | 0.023 | Hatching success | Tillitt et al., 1993 |
| Wood duck (<i>Aix sponsa</i>) | NOAEL | TEQs, pesticides | 0.005 | Nest success, hatching success, duckling production | White and Seginak, 1994; White and Hoffman, 1995 |

**TABLE 4-17
TOXICITY ENDPOINTS FOR OTHER MAMMALS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS**

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE DOSE (mg/kg/day) | FOOD INGESTION RATE (kg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|---|----------------------------------|--------------------------------|------------------|--------------|----------------------------|---------------------------------|------------------------------|---|-----------------------------------|
| Laboratory studies^a | | | | | | | | | |
| Osborne-Mendel Rat | Oral-gavage | 2.5 wk, 2 d per week | LD ₅₀ | Aroclor 1254 | 1530 | 0.099 | | Mortality | Garthoff et al., 1981 (ATSDR) |
| Osborne-Mendel Rat | Oral-gavage | 2.5 wk, 2 d per week | LD ₅₀ | Aroclor 1254 | 1530 | 0.099 | | Mortality | Garthoff et al., 1981 (ATSDR) |
| Wistar Rat | Diet | From mating to weaning of pups | LD ₅₀ | Aroclor 1254 | 22 | 0 | 269 | 2 day postnatal mortality of offspring | Overmann et al., 1987 |
| Juvenile Male Rat | Single intraperitoneal injection | Observed after 14 days | LOAEL | Aroclor 1248 | 2000 | | | Growth rate of juveniles | Harris et al., 1993 |
| Juvenile Male Rat | Single intraperitoneal injection | Observed after 14 days | LOAEL | Aroclor 1232 | 2000 | | | Growth rate of juveniles | Harris et al., 1993 |
| Sherman Rat | Diet | 8 months | LOAEL | Aroclor 1260 | 72.4 | 0.08 | | Mortality | Kimbrough et al., 1972 (ATSDR) |
| Raccoon (<i>Procyon lotor</i>) | Diet | 8 days | EL-effect | Aroclor 1254 | 50 | | | Decreased weight gain | Montz et al., 1982 |
| Osborne-Mendel Rat | Diet | During pregnancy and lactation | LOAEL | Not reported | 49.471 | 0.080 | 500 | Reduced litter size | Collins & Capen, 1980 |
| Balb/c Mouse | Oral | 6 months | LOAEL | Aroclor 1254 | 48.75 | 0.18 | | Mortality | (ATSDR) |
| Adult Female Rat | Oral | Day 1,3,5,7 and 9 of lactation | LOAEL | Aroclor 1254 | 32 | 0.08 | | Reduced growth rate of offspring | Sager & Girard, 1994 |
| Wistar Rat | Oral-gavage | 1 month | LOAEL | Aroclor 1254 | 30 | 0.08 | | Decreased litter size, survival of weanlings | Brezner et al., 1984 (ATSDR) |
| White-footed Mouse (<i>Peromyscus leucopus</i>) | Diet | 12 weeks | EL-effect | Aroclor 1254 | 17 | | 10 | Reduced growth rate reproduction in second generation | Linzey, 1988 (Golub) |
| Wistar Rat | Diet | 42 days | LOAEL | Aroclor 1254 | 13.5 | 0.08 | | Neonatal death | Overmann, 1987 (ATSDR) |
| Mouse | Diet | 108 days | LOAEL | Aroclor 1254 | 12.5 | 0.18 | | Decreased conception | Welsch, 1975 (ATSDR) |
| Rabbit | Oral-gavage | 28 days | LOAEL | Aroclor 1254 | 12.5 | 0.034 | | Fetal death | Villeneuve et al., 1971 (ATSDR) |
| Pig | Diet | 91 days | LOAEL | Aroclor 1242 | 9.2 | | | Decreased weight gain | Hansen et al., 1976 |
| New Zealand White Rabbit | Diet | > 4 weeks | LOAEL | Aroclor 1248 | 8.9 | 0.0 | 250 | Reduced growth rate in offspring | Thomas and Hinsdill, 1980 (Golub) |
| Osborne-Mendel Rat | Diet | During pregnancy and lactation | LOAEL | Not reported | 4.947 | 0.080 | 50 | Reduced growth rate of offspring | Collins & Capen, 1980 |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 2 months | LOAEL | Aroclor 1248 | 4.3 | 0.2 | | Decreased conception | Allen et al., 1974a (ATSDR) |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 2 months | LOAEL | Aroclor 1248 | 4.3 | 0.2 | | Abortion | Allen et al., 1974a (ATSDR) |
| Fischer Rat | Diet | 105 weeks | LOAEL | Aroclor 1254 | 2.5 | 0.08 | | Decreased survival | NCI, 1978 (ATSDR) |
| Guinea Pig | Oral-gavage | Gestational day 18-60 | LOAEL | Clophen A50 | 2.5 | | | Fetal death | Lundkvist, 1990 (ATSDR) |
| Sherman Rat | Diet | Multigenerational | LOAEL | Aroclor 1254 | 1.5 | 0.08 | 20 | Decreased litter size | Linder et al., 1974 |
| Wistar Rat | Diet | 52 weeks | LOAEL | Aroclor 1254 | 1 | 0.08 | | Decreased growth rate | Phillips et al., 1972 |
| Oldfield Mouse (<i>Peromyscus polionotus</i>) | Diet | 12 months | EL-effect | Aroclor 1254 | 0.68 | 0.01 | 5 | Decreased offspring born per mated pair, birth weight, % survival of offspring to weaning | McCoy et al., 1995 |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 38 weeks | LOAEL | Aroclor 1254 | 0.2 | 0.2 | | No conception, abortion | Arnold et al., 1990 (ATSDR) |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 7 months | LOAEL | Aroclor 1248 | 0.2 | 0.2 | | Decreased conception | Barsotti et al., 1976 (ATSDR) |
| Wistar Rat | Diet | From mating to weaning of pups | LOAEL | Aroclor 1254 | 0.2 | 0.08 | 2.5 | Reduced growth rate in offspring | Overmann et al., 1987 |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 2 months | LOAEL | Aroclor 1242 | 0.12 | 0.2 | | No weight gain | Becker et al., 1979 (ATSDR) |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 1.5 years | LOAEL | Aroclor 1248 | 0.12 | 0.2 | 5 | Reduced birth weight | Allen and Barsotti, 1976 (Golub) |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 18 months | LOAEL | Aroclor 1248 | 0.1 | 0.2 | | Infant mortality | Allen et al., 1980 (ATSDR) |
| Cynomolgus Monkey | Diet | 238 days | LOAEL | Aroclor 1254 | 0.1 | | | 100% fetal death | Truelove et al., 1982 (ATSDR) |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | 18.2 | LOAEL | Aroclor 1248 | 0.08 | 0.2 | | Decreased birth weight | Levin et al., 1988 (ATSDR) |

TABLE 4-17
TOXICITY ENDPOINTS FOR OTHER MAMMALS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE DOSE (mg/kg/day) | FOOD INGESTION RATE (kg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|--|----------------------------------|--------------------------------|--------------|--------------|----------------------------|---------------------------------|------------------------------|----------------------------------|---------------------------------------|
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | > 8 months | LOAEL | Aroclor 1016 | 0.04 | 0.2 | 1 | Reduced birth weight | Barsotti and Van Miller, 1984 (Golub) |
| Swine | Diet | Throughout gestation | EL-effect | Aroclor 1242 | Not available | | 20 | Decreased litter size | Hansen et al., 1975 (Golub) |
| Juvenile Male Rat | Single intraperitoneal injection | Observed after 14 days | NOAEL | Aroclor 1248 | 480 | | | Growth rate of juveniles | Harris et al., 1993 |
| Juvenile Male Rat | Single intraperitoneal injection | Observed after 14 days | NOAEL | Aroclor 1232 | 480 | | | Growth rate of juveniles | Harris et al., 1993 |
| Wistar Rat | Diet | 52 weeks | NOAEL | Aroclor 1254 | 10 | 0.08 | | Decreased growth rate | Phillips et al., 1972 |
| Rabbit | Oral-gavage | 28 days | NOAEL | Aroclor 1254 | 10 | 0.034 | | Fetal death | Villeneuve et al., 1971 (ATSDR) |
| Adult Female Rat | Oral | Day 1,3,5,7 and 9 of lactation | NOAEL | Aroclor 1254 | 8 | 0.099 | | Growth rate of offspring | Sager & Girard, 1994 |
| New Zealand White Rabbit | Diet | > 4 weeks | NOAEL | Aroclor 1248 | 3.6 | 0.034 | 100 | Reduced growth rate in offspring | Thomas and Hinsdill, 1980 (Golub) |
| Sherman Rat | Diet | Multigenerational | NOAEL | Aroclor 1254 | 0.32 | 0.08 | 5 | Decreased litter size | Linder et al., 1974 |
| Osborne-Mendel Rat | Diet | During pregnancy and lactation | NOAEL | Aroclor 1254 | 0.059 | 0.08 | 50 | Reduced litter size | Collins & Capen, 1980 |
| Rhesus Monkey (<i>Macaca mulatta</i>) | Diet | > 8 months | NOAEL | Aroclor 1016 | 0.01 | 0.2 | 0.25 | Reduced birth weight | Barsotti and Van Miller, 1984 (Golub) |
| Wistar Rat | Diet | From mating to weaning of pups | NOAEL | Aroclor 1254 | 0.0016 | 0.08 | 0.02 | Reduced growth rate in offspring | Overmann et al., 1987 |

Notes:

*No relevant field studies were found.

Dose to rhesus monkey calculated using food ingestion rate of 0.2 kg/day and body weight of 5 kg (Sample et al., 1996)

TABLE 4-18
TOXICITY ENDPOINTS FOR OTHER MAMMALS - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE DIOXIN EQUIVALENTS (ug TEQ/kg/day)* | EFFECT ENDPOINT | REFERENCE |
|--|----------------|-----------------------------|------------------|------------------|--|-------------------------|-----------------------------|
| Laboratory studies | | | | | | | |
| Hamster | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | 1,160 - 5,050 | Mortality | Kociba and Schwetz, |
| Mouse | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | 114 -284 | Mortality | Kociba and Schwetz, |
| Dog | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | about 100 - 200 | Mortality | Kociba and Schwetz, |
| Rabbit | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | 115 | Mortality | Schwetz et al., 1973 |
| Rhesus monkey (<i>Macaca mulatta</i>) | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | approx. 70 | Mortality | Kociba and Schwetz, 1982 |
| Rat | Oral | Single dose | LD ₅₀ | 2,3,7,8-TCDD | 22 - 45 | Mortality | Schwetz et al., 1973 |
| Guinea pig | Oral | Single dose | LD ₅₂ | 2,3,7,8-TCDD | 0.6 - 2.1 | Mortality | Schwetz et al., 1973 |
| Rat | | Gestation days 6 to | LOAEL | 2,3,7,8-TCDD | 0.25 | Litter size, pup weight | Khera and Ruddick, 1973 |
| Rat | | 2 years | LOAEL | 2,3,7,8-TCDD | 0.1 | Female mortality | Kociba et al., 1978 |
| Rat | | 3 generations | LOAEL | 2,3,7,8-TCDD | 0.01 | Reproductive capacity | Murray et al., 1979 |
| Rhesus monkey (<i>Macaca mulatta</i>) | | 7 months | LOAEL | 2,3,7,8-TCDD | 0.0021 | Number of births | Allen et al., 1979 |
| Rhesus monkey (<i>Macaca mulatta</i>) | | 7 - 48 months, maternal | LOAEL | 2,3,7,8-TCDD | 0.00059 | Reproductive | Bowman et al., 1989b |
| Rat | | Gestation days 6 to | NOAEL | 2,3,7,8-TCDD | 0.125 | Litter size, pup weight | Khera and Ruddick, 1973 |
| Rat | | 2 years | NOAEL | 2,3,7,8-TCDD | 0.01 | Female mortality | Kociba et al., 1978 |
| Rat | | 3 generations | NOAEL | 2,3,7,8-TCDD | 0.001 | Reproductive capacity | Murray et al. 1979 |
| Rhesus monkey (<i>Macaca mulatta</i>) | | 7 to 48 months, maternal | NOAEL | 2,3,7,8-TCDD | 0.00012 | Reproductive | Bowman et al., 1989 |

TABLE 4-19
TOXICITY ENDPOINTS FOR MINK - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS

| SPECIES | EXPOSURE MEDIA | EXPOSURE DURATION | EFFECT LEVEL | PCB TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|--------------------------------|----------------|-------------------|--------------|--------------------------|----------------------------|------------------------------|---|---|
| Laboratory studies | | | | | | | | |
| Mink (<i>Mustela vision</i>) | Diet | 4 weeks | LD50 | Aroclor 1254 | 11.5 | 84 | Adult mortality | Hornshaw (1984), as cited in Aulerich et al. (1986) |
| Mink (<i>Mustela vision</i>) | Diet | 4 weeks | LD50 | Aroclor 1254 | 10.8 | 79 | Adult mortality | Aulerich et al. (1986) |
| Mink (<i>Mustela vision</i>) | Diet | 4 weeks | LD50 | Aroclor 1254 | 6.4 | 47 | Adult mortality | Hornshaw et al. (1986) |
| Mink (<i>Mustela vision</i>) | Diet | 4 weeks | LD50 | Aroclor 1254 (weathered) | 6.4 | 47 | Adult mortality | Aulerich et al. (1986) |
| Mink (<i>Mustela vision</i>) | Diet | 9 months | LD50 | Aroclor 1254 | 0.9 | 6.6 | Mortality | Ringer et al. (1981) |
| Mink (<i>Mustela vision</i>) | Diet | 8 months | EL-effect | Aroclor 1016 | 2.7 | 20 | Reduced birth weight and growth rate of | Bleavins et al., 1980 |
| Mink (<i>Mustela vision</i>) | Diet | 8 months | EL-effect | Aroclor 1016 | 2.7 | 20 | Adult mortality | Bleavins et al., 1980 |
| Mink (<i>Mustela vision</i>) | Diet | 4 weeks | LOAEL | Aroclor 1254 | 1.4 | 10 | Reduced weight gain in juveniles | Hornshaw et al. (1986) |
| Mink (<i>Mustela vision</i>) | Diet | 8 months | LOAEL | Aroclor 1242 | 1.4 | 10 | Adult mortality | Bleavins et al., 1980 |
| Mink (<i>Mustela vision</i>) | Diet | 3 months | EL-effect | Clophen A-50 | 2 | Not reported | Decreased number of kits born alive | Kihlstrom et al., 1992 |
| Mink (<i>Mustela vision</i>) | Diet | 3 months | EL-effect | Aroclor 1254 | 2 | Not reported | Decreased number of kits born alive | Kihlstrom et al., 1992 |
| Mink (<i>Mustela vision</i>) | Diet | 8 months | LOAEL | Aroclor 1242 | 0.7 | 5 | Reduced reproduction | Bleavins et al., 1980 |
| Mink (<i>Mustela vision</i>) | Diet | 4 months | LOAEL | Aroclor 1254 | 0.7 | 5 | Decreased number of kits born alive | Aulerich and Ringer |
| Mink (<i>Mustela vision</i>) | Diet | 105 days | LOAEL | Aroclor 1254 (weathered) | 0.5 | 3.57 | Adult mortality | Platanow & Karstad (1973) |
| Mink (<i>Mustela vision</i>) | Diet | 66 days | LOAEL | Not reported | 0.5 | 3.3 | Decreased number of kits born alive | Jensen et al. (1977) |
| Mink (<i>Mustela vision</i>) | Diet | 4 months | EL-effect | Aroclor 1254 | 0.3 | 2.5 | Decreased number of kits born alive | Aulerich et al. (1985) |
| Mink (<i>Mustela vision</i>) | Diet | 6 months | EL-effect | Aroclor 1254 | 0.1 | 1 | Reduced growth rates of kits | Wren et al., 1987 |
| Mink (<i>Mustela vision</i>) | Diet | 160 days | LOAEL | Aroclor 1254 (weathered) | 0.09 | 0.64 | Reduced number of kits born alive | Platanow & Karstad (1973) |
| Mink (<i>Mustela vision</i>) | Diet | 8 months | NOAEL | Aroclor 1242 | 0.9 | 5 | Adult mortality | Bleavins et al., 1980 |
| Mink (<i>Mustela vision</i>) | Diet | 4 months | NOAEL | Aroclor 1254 | 0.1 | 1 | Decreased number of kits born alive | Aulerich & Ringer (1977) |

TABLE 4-20
TOXICITY ENDPOINTS FOR MINK - FIELD STUDIES
EFFECTIVE DIETARY DOSES OF TOTAL PCBs AND AROCLORS

| SPECIES | FIELD COMPONENT | STUDY DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE FOOD CONC. (mg/kg) | EFFECT ENDPOINT | REFERENCE |
|-------------------------------|--|---|--------------|--------------------|----------------------------|------------------------------|--|----------------------|
| Field studies | | | | | | | | |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Mink were fed prior to and throughout the reproductive period | LOAEL | PCBs, TEQs, others | 0.13 | N/A | Reproductive success, growth/survival of offspring | Heaton et al. (1995) |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Mink fed prior to breeding and over two generations | LOAEL | PCBs, pesticides | 0.08 | 0.5 | Kit survival | Restum et al., 1998 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Mink fed prior to breeding and over two generations | LOAEL | PCBs, pesticides | 0.04 | 0.25 | Reduced growth rate of kits | Restum et al., 1998 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Mink fed prior to breeding and over two generations | LOAEL | PCBs, pesticides | 0.04 | 0.25 | Kit survival | Restum et al., 1998 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Mink were fed prior to and throughout the reproductive period | NOAEL | PCBs, TEQs, others | 0.004 | N/A | Reproductive success, growth/survival of offspring | Heaton et al. (1995) |

TABLE 4-21
TOXICITY ENDPOINTS FOR MINK - LABORATORY STUDIES
EFFECTIVE DIETARY DOSES OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | FIELD COMPONENT | STUDY DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE (mg/kg/day) | EFFECTIVE DOSE DIOXIN EQUIVALENTS (ug TEQ/kg/day) | EFFECT ENDPOINT | REFERENCE |
|--|-----------------|----------------|------------------|------------------|----------------------------|---|-----------------|------------------------|
| Laboratory studies | | | | | | | | |
| Mink kits (<i>Mustela vison</i>) | Intraperitoneal | 12 days | LD ₅₃ | 2,3,7,8-TCDD | < 0.01 | < 0.01 | Mortality | Aulerich et al., 1988 |
| Mink males (<i>Mustela vison</i>) | Oral | Single dose | LD ₅₁ | 2,3,7,8-TCDD | 4.2 | 4.2 | Mortality | Hochstein et al., 1988 |

TABLE 4-22
TOXICITY ENDPOINTS FOR MINK - FIELD STUDIES
EFFECTIVE DIETARY DOSES OF DIOXIN TOXIC EQUIVALENTS (TEQs)

| SPECIES | FIELD COMPONENT | STUDY DURATION | EFFECT LEVEL | CONTAMINANT TYPE | EFFECTIVE DOSE DIOXIN EQUIVALENTS (ug TEQ/kg/day) | EFFECT ENDPOINT | REFERENCE |
|-------------------------------|--|---|--------------|---------------------------|---|----------------------------------|----------------------|
| Field studies | | | | | | | |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | LOAEL | TEQs, pesticides | 0.0036 | Growth rate of kits | Heaton et al. (1995) |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | LOAEL | TEQs (chemically derived) | 0.00224 | Growth and survival rate of kits | Tillitt et al., 1996 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | LOAEL | TEQs (bioassay derived) | 0.00027 | Growth and survival rate of kits | Tillitt et al., 1996 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | NOAEL | TEQs (bioassay derived) | 0.00344 | Growth and survival rate of kits | Tillitt et al., 1996 |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | NOAEL | TEQs, pesticides | 0.00025 | Growth rate of kits | Heaton et al. (1995) |
| Mink (<i>Mustela vison</i>) | Fed contaminated carp from Saginaw Bay, MI | Fed prior to and throughout breeding period | NOAEL | TEQs (chemically derived) | 0.00008 | Growth and survival rate of kits | Tillitt et al., 1996 |

TABLE 4-23
TAXONOMY OF STUDIED ORGANISMS

| <i>Phylum</i> | <i>Class</i> | <i>Subclass</i> | <i>Order</i> | <i>Family</i> | <i>Genus</i> | <i>Species</i> | <i>Common name</i> |
|---------------|--------------|-----------------|------------------|-------------------|-----------------------|---------------------------|-----------------------------|
| Chordata | Mammalia | | Carnivora | Mustelidae | <i>Lutra</i> | <i>canadensis</i> | River Otter |
| Chordata | Mammalia | | Carnivora | Mustelidae | <i>Mustela</i> | <i>vision</i> | Mink |
| Chordata | Mammalia | | Carnivora | Procyonidae | <i>Procyon</i> | <i>lotor</i> | Raccoon |
| Chordata | Mammalia | | Chiroptera | Vespertilionidae | <i>Myotis</i> | <i>lucifugus</i> | Little Brown Bat |
| Chordata | Mammalia | | Lagomorphus | Leporidae | [<i>Sylvilagus</i>] | [<i>transitionalis</i>] | Rabbit [Eastern Cottontail] |
| Chordata | Mammalia | | Rodentia | Muridae | [<i>Peromyscus</i>] | [<i>polionotus</i>] | Mouse [Oldfield Mouse] |
| Chordata | Mammalia | | Rodentia | Muridae | [<i>Rattus</i>] | [<i>rattus</i>] | Rat |
| | | | | | | | Birds |
| Chordata | Aves | | Anseriformes | Anatidae | <i>Aix</i> | <i>sponsa</i> | Wood Duck |
| Chordata | Aves | | Anseriformes | Anatidae | <i>Anas</i> | <i>platyrhynchos</i> | Mallard Duck |
| Chordata | Aves | | Charadriiformes | Laridae | <i>Hydropogone</i> | <i>caspia</i> | Caspian tern |
| Chordata | Aves | | Charadriiformes | Laridae | <i>Sterna</i> | <i>hirundo</i> | Common tern |
| Chordata | Aves | | Charadriiformes | Laridae | <i>Sterna</i> | <i>forsteri</i> | Forster's tern |
| Chordata | Aves | | Ciconiiformes | Ardeidae | <i>Ardea</i> | <i>herodias</i> | Great Blue Heron |
| Chordata | Aves | | Coraciiformes | Alcedinidae | <i>Ceryle</i> | <i>alcyon</i> | Kingfisher |
| Chordata | Aves | | Falconiiformes | Accipitridae | <i>Haliaeetus</i> | <i>leucocephalus</i> | Bald Eagle |
| Chordata | Aves | | Falconiiformes | Falconidae | <i>Falco</i> | <i>sparverius</i> | American Kestrel |
| Chordata | Aves | | Falconiiformes | Pandionidae | <i>Pandion</i> | <i>haliaeetus</i> | Osprey |
| Chordata | Aves | | Galliformes | Phasianidae | <i>Colinus</i> | <i>virginianus</i> | Northern Bobwhite |
| Chordata | Aves | | Galliformes | Phasianidae | <i>Coturnix</i> | <i>coturnix</i> | Japanese Quail |
| Chordata | Aves | | Galliformes | Phasianidae | <i>Gallus</i> | <i>domesticus</i> | Domestic Chicken |
| Chordata | Aves | | Galliformes | Phasianidae | <i>Phasianus</i> | <i>colchicus</i> | Ring-Necked Pheasant |
| Chordata | Aves | | Passeriformes | Hirundinidae | <i>Tachycineta</i> | <i>bicolor</i> | Tree Swallow |
| Chordata | Aves | | Passeriformes | Icteridae | <i>Agelaius</i> | <i>phoeniceus</i> | Red-Winged Blackbird |
| Chordata | Aves | | Passeriformes | Icteridae | <i>Molothrus</i> | <i>ater</i> | Brown-Headed Cowbird |
| Chordata | Aves | | Passeriformes | Icteridae | <i>Quiscalus</i> | <i>quiscula</i> | Common Grackle |
| Chordata | Aves | | Passeriformes | Sturnidae | <i>Sturnus</i> | <i>vulgaris</i> | European Starling |
| Chordata | Aves | | Pelecaniformes | Phalacrocoracidae | <i>Phalacrocorax</i> | <i>auritus</i> | Double-Crested Cormorant |
| Chordata | Aves | | Strigiformes | Strigidae | <i>Otus</i> | <i>asio</i> | Screech Owl |
| | | | | | | | Fish |
| Chordata | Pisces | Actinopterygii | Acipenseriformes | Acipenseridae | <i>Acipenser</i> | <i>brevirostrum</i> | Shortnose Sturgeon |
| Chordata | Pisces | Actinopterygii | Beloniformes | Adrianichthyidae | <i>Oryzias</i> | <i>latipes</i> | Medaka |
| Chordata | Pisces | Actinopterygii | Clupeiformes | Clupeidae | <i>Clupea</i> | <i>harengus</i> | Baltic Herring |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Catostomidae | <i>Catostomus</i> | <i>commersoni</i> | White sucker |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Cyprinidae | <i>Danio</i> | <i>danio</i> | Zebrafish |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Cyprinidae | <i>Notropis</i> | <i>hudsonius</i> | Spottail Shiner |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Cyprinidae | <i>Phoxinus</i> | <i>phoxinus</i> | Minnow |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Cyprinidae | <i>Pimephalus</i> | <i>promelas</i> | Fathead Minnow |
| Chordata | Pisces | Actinopterygii | Cypriniformes | Cyprinodontidae | <i>Fundulus</i> | <i>heterocliuis</i> | Killifish |
| Chordata | Pisces | Actinopterygii | Perciformes | Centrarchidae | <i>Lepomis</i> | <i>gibbosus</i> | Pumpkinseed |
| Chordata | Pisces | Actinopterygii | Perciformes | Centrarchidae | <i>Lepomis</i> | <i>auritus</i> | Redbreast Sunfish |
| Chordata | Pisces | Actinopterygii | Perciformes | Centrarchidae | <i>Micropterus</i> | <i>salmoides</i> | Largemouth Bass |
| Chordata | Pisces | Actinopterygii | Perciformes | Moronidae | <i>Morone</i> | <i>americana</i> | White Perch |
| Chordata | Pisces | Actinopterygii | Perciformes | Moronidae | <i>Morone</i> | <i>saxatilis</i> | Striped Bass |
| Chordata | Pisces | Actinopterygii | Perciformes | Percidae | <i>Perca</i> | <i>flavescens</i> | Yellow Perch |
| Chordata | Pisces | Actinopterygii | Perciformes | Sciaenidae | <i>Leiostomus</i> | <i>xanthurus</i> | Spot |

**TABLE 4-23
TAXONOMY OF STUDIED ORGANISMS**

| <i>Phylum</i> | <i>Class</i> | <i>Subclass</i> | <i>Order</i> | <i>Family</i> | <i>Genus</i> | <i>Species</i> | <i>Common name</i> |
|---------------|--------------|-----------------|-------------------|----------------|---------------------------|--------------------|--------------------|
| Chordata | Pisces | Actinopterygii | Perciformes | Sparidae | <i>Lagodon</i> | <i>rhomboides</i> | Pinfish |
| Chordata | Pisces | Actinopterygii | Pleuronectiformes | Pleuronectidae | <i>Parophrys</i> | <i>vetulus</i> | English Sole |
| Chordata | Pisces | Actinopterygii | Pleuronectiformes | Pleuronectidae | <i>Platichthys</i> | <i>flesus</i> | Baltic Flounder |
| Chordata | Pisces | Actinopterygii | Pleuronectiformes | Pleuronectidae | <i>Platichthys</i> | <i>stellatus</i> | Starry Flounder |
| Chordata | Pisces | Actinopterygii | Pleuronectiformes | Pleuronectidae | <i>Pseudopleuronectes</i> | <i>americanus</i> | Winter Flounder |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Esocidae | <i>Esox</i> | <i>lucius</i> | Northern Pike |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Coregonus</i> | <i>artedii</i> | Lake Herring |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Oncorhynchus</i> | <i>tshawytscha</i> | Chinook Salmon |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Salmo</i> | <i>gairdneri</i> | Rainbow Trout |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Salvelinus</i> | <i>alpinus</i> | Arctic Charr |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Salvelinus</i> | <i>fontinalis</i> | Brook Trout |
| Chordata | Pisces | Actinopterygii | Salmoniformes | Salmonidae | <i>Salvelinus</i> | <i>namaycush</i> | Lake Trout |
| Chordata | Pisces | Actinopterygii | Siluriformes | Ictaluridae | <i>Ictalurus</i> | <i>nebulosus</i> | Brown Bullhead |
| Chordata | Pisces | Actinopterygii | Siluriformes | Ictaluridae | <i>Ictalurus</i> | <i>punctatus</i> | Channel Catfish |

**TABLE 4-24
STANDARD ANIMAL BODY WEIGHTS AND FOOD INTAKE RATES**

| <i>Animal</i> | <i>Body Weight (kg)</i> | <i>Food Ing. Rate (g/d)</i> | <i>Food Ingestion Rate (kg/d)</i> | <i>Food factor (kg/kg body wt/d)</i> |
|---|-----------------------------|---------------------------------|---------------------------------------|--|
| MAMMALS | | | | |
| Mink | 1 | | 0.137 | 0.137 |
| Mouse | 0.03 | | 0.0055 | 0.180 |
| | 0.028 | | | |
| Mean Mouse | 0.029 | | | |
| Mouse, Oldfield | 0.014 | 1.9 | 0.0019 | |
| Rabbit | 3.8 | | 0.135 | 0.034 |
| Rhesus Monkey | 5 | | 0.2 | 0.040 |
| Rat | 0.35 | | 0.028 | 0.080 |
| | 0.435 | | | |
| | 0.303 | | | |
| | 0.273 | | 0.0375 | 0.137 |
| | 0.365 | | | |
| | 0.26 | | | |
| Mean Rat | 0.331 | | 0.03275 | 0.099 |
| BIRDS | | | | |
| Blackbird, Red-Winged | 0.064 | | 0.0137 | 0.214 |
| Chicken, Domestic--adult | 1.6 | | 0.11 | 0.069 |
| | 1.5 | | 0.106 | 0.071 |
| Mean Chicken, Domestic--adult | 1.55 | | 0.108 | 0.070 |
| Chickens, Domestic--chick | 0.121 | | 0.0126 | 0.104 |
| | 0.534 | | 0.044 | 0.082 |
| Mean Chicken, Domestic--chick | 0.3275 | | 0.0283 | 0.086 |
| Cowbird, Brown-headed | 0.049 | | 0.01087 | 0.222 |
| Dove, Ringed | 0.155 | | 0.017 | 0.110 |
| Duck, Mallard--adult | 1 | | 0.1 | 0.100 |
| | 1.153 | | 0.11 | 0.095 |
| | 1.15 | 115 | 0.115 | 0.100 |
| | 1 | | 0.128 | 0.128 |
| | 1.17 | | 0.121 | 0.103 |
| Mean Duck, Mallard--adult | 1.0946 | | 0.1148 | 0.105 |
| Duck, Mallard--duckling | 0.782 | 78.2 | 0.0782 | 0.100 |
| Kestrel, American | 0.13 | | 0.01 | 0.077 |
| Owl, Screech | 0.181 | 25 | 0.025 | 0.138 |
| Pheasant, Ring-necked | 1 | | 0.0582 | 0.058 |
| Quail, Japanese | 0.15 | | 0.0169 | 0.113 |
| Quail, Japanese--3 months | 0.072 | | | |
| | | | | |
| Note: All values are from Toxicological Benchmarks for Wildlife:1996 Revision (USEPA, 1996) unless otherwise noted. | | | | |

TABLE 4-25
TOXICITY REFERENCE VALUES FOR FISH
DIETARY DOSES AND EGG CONCENTRATIONS OF TOTAL PCBs AND DIOXIN TOXIC EQUIVALENTS (TEQs)

| TRVs | | Pumpkinseed (<i>Lepomis gibbosus</i>) | Spottail Shiner (<i>Notropis hudsonius</i>) | Brown Bullhead (<i>Ictalurus nebulosus</i>) | Yellow Perch (<i>Perca flavescens</i>) | White Perch (<i>Morone americana</i>) | Largemouth Bass (<i>Micropterus salmoides</i>) | Striped Bass (<i>Morone saxatilis</i>) | Shortnose Sturgeon (<i>Acipenser brevirostrum</i>) | References |
|--|-------|--|---|--|---|--|--|---|--|--|
| <i>Tissue Concentration</i> | | | | | | | | | | |
| Lab-based TRVs for PCBs (mg/kg wet wt.) | LOAEL | 17 | 170 | 17 | 17 | 17 | 17 | 17 | 17 | Bengtsson (1980) |
| | NOAEL | 1.5 | 15 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 | |
| Field-based TRVs for PCBs (mg/kg wet wt.) | LOAEL | NA | NA | NA | NA | NA | NA | NA | NA | White perch and striped bass: Westin et al. (1983) Pumpkinseed and Largemouth bass: Adams et al. (1989, 1990, 1992) |
| | NOAEL | 0.5 | NA | NA | NA | 3.1 | 0.5 | 3.1 | NA | |
| <i>Egg Concentration</i> | | | | | | | | | | |
| Lab-based TRV for TEQs (ug/kg lipid) from salmonids | LOAEL | 0.6 | Not derived | 18 | 0.6 | 0.6 | 0.6 | 0.6 | 0.6 | Brown Bullhead: Elonen et al. (1998) All others: Walker et al. (1994) |
| | NOAEL | 0.29 | Not derived | 8.0 | 0.29 | 0.29 | 0.29 | 0.29 | 0.29 | |
| Lab-based TRV for TEQs (ug/kg lipid) from non-salmonids | LOAEL | 10.3 | 103 | Not derived | 10.3 | 10.3 | 10.3 | 10.3 | 10.3 | Oliveri and Cooper (1997) |
| | NOAEL | 0.54 | 5.4 | Not derived | 0.54 | 0.54 | 0.54 | 0.54 | 0.54 | |
| Field-based TRVs for TEQs (ug/kg lipid) | LOAEL | NA | NA | NA | NA | NA | NA | NA | NA | |
| | NOAEL | NA | NA | NA | NA | NA | NA | NA | NA | |

Note:

^a Pumpkinseed (*Lepomis gibbosus*) and spottail shiner (*Notropis hudsonius*)

Units vary for PCBs and TEQ.

NA = Not available

Selected TRVs are **bolded and italicized**.

TABLE 4-26
TOXICITY REFERENCE VALUES FOR BIRDS
DIETARY DOSES AND EGG CONCENTRATIONS OF TOTAL PCBs AND DIOXIN TOXIC EQUIVALENTS (TEQs)

| TRVs | | Tree Swallow (<i>Tachycineta bicolor</i>) | Mallard Duck (<i>Anas platyrhynchos</i>) | Belted Kingfisher (<i>Ceryle alcyon</i>) | Great Blue Heron (<i>Ardea herodias</i>) | Bald Eagle (<i>Haliaeetus leucocephalus</i>) | References |
|---------------------------------------|-------|--|---|---|--|---|---|
| <i>Dietary Dose</i> | | | | | | | |
| Lab-based TRVs for PCBs (mg/kg/day) | LOAEL | 0.07 | 2.6 | 0.07 | 0.07 | 0.07 | Mallard: Custer and Heinz (1980) |
| | NOAEL | 0.01 | 0.26 | 0.01 | 0.01 | 0.01 | All others: Scott (1977) |
| Field-based TRVs for PCBs (mg/kd/day) | LOAEL | NA | NA | NA | NA | NA | Tree Swallow: US EPA Phase 2 Database (1998) |
| | NOAEL | 16.1 | NA | NA | NA | NA | |
| <i>Egg Concentration</i> | | | | | | | |
| Lab-based TRVs for PCBs (mg/kg egg) | LOAEL | 2.21 | 2.21 | 2.21 | 2.21 | 2.21 | Scott (1977) |
| | NOAEL | 0.33 | 0.33 | 0.33 | 0.33 | 0.33 | |
| Field-based TRVs for PCBs (mg/kg egg) | LOAEL | NA | NA | NA | NA | NA | Bald Eagle: Wiemeyer (1984, 1993) |
| | NOAEL | 26.7 | NA | NA | NA | 3.0 | Tree Swallow: US EPA Phase 2 Database (1998) |
| Lab-based TRVs for TEQs (ug/kg egg) | LOAEL | 0.02 | 0.02 | 0.02 | NA | 0.02 | Great Blue Heron: Janz and Bellward (1996) |
| | NOAEL | 0.01 | 0.01 | 0.01 | 2 | 0.01 | Others: Powell et al. (1996a) |
| Field-based TRVs for TEQs (ug/kg egg) | LOAEL | NA | 0.02 | NA | 0.5 | NA | Mallard: White and Segniak (1994); White and Hoffman (1995) |
| | NOAEL | 13 | 0.005 | NA | 0.3 | NA | Great Blue Heron: Sanderson et al. (1994) Tree Swallow: US EPA Phase 2 Database (1998) |

Note: Units vary for PCBs and TEQ.
NA = Not Available
Selected TRVs are **bolded and italicized**.

TABLE 4-27
TOXICITY REFERENCE VALUES FOR MAMMALS
DIETARY DOSES OF TOTAL PCBs AND DIOXIN TOXIC EQUIVALENTS (TEQs)

| TRVs | | Little Brown Bat (<i>Myotis lucifugus</i>) | Raccoon (<i>Procyon lotor</i>) | Mink (<i>Mustela vison</i>) | Otter (<i>Lutra canadensis</i>) | References |
|---------------------------------------|-------|---|-------------------------------------|----------------------------------|--------------------------------------|--|
| Lab-based TRVs for PCBs (mg/kg/day) | LOAEL | <i>0.15</i> | <i>0.15</i> | 0.07 | 0.07 | Mink and otter: Aulerich and Ringer (1977) |
| | NOAEL | <i>0.032</i> | <i>0.032</i> | 0.01 | 0.01 | Raccoon and bat: Linder et al. (1984) |
| Field-based TRVs for PCBs (mg/kg/day) | LOAEL | NA | NA | <i>0.13</i> | <i>0.13</i> | Heaton et al. (1995) |
| | NOAEL | NA | NA | <i>0.004</i> | <i>0.004</i> | |
| Lab-based TRVs for TEQs (ug/kg/day) | LOAEL | <i>0.001</i> | <i>0.001</i> | 0.001 | 0.001 | Murray et al. (1979) |
| | NOAEL | <i>0.0001</i> | <i>0.0001</i> | 0.0001 | 0.0001 | |
| Field-based TRVs for TEQs (ug/kg/day) | LOAEL | NA | NA | <i>0.00224</i> | <i>0.00224</i> | Tillitt et al. (1996) |
| | NOAEL | NA | NA | <i>0.00008</i> | <i>0.00008</i> | |

Note: Units vary for PCBs and TEQ.

Note: TRVs for raccoon and bat are based on multi-generational studies to which interspecies uncertainty factors are applied.

NA = Not Available

Final selected TRVs are ***bolded and italicized***.

TABLE 4-28: WILDLIFE SURVEY RESULTS Amphibians

Hudson River

New York

| Information Source | Date | Contact | Response | Contact Information | Data Available | Information/Findings |
|--|---------------------|-----------------------|-----------|---|---|---|
| Amphibians | | | | | | |
| Amphibian Expert | 1-Jun-99 | Email | Yes | Thomas Palmer, frog consultant for Wellesley Project; Ophis@world.std.com | He doesn't know anything about PCB effects on frogs; posted message on amphibian web page | Recommended the following website: http://cciw.ca/green-lane/herptox/ |
| NYSDEC - Amphibian and Reptile Atlas Project | 3-Jun-99 | Email | No | herps@gw.dec.state.ny.us; http://www.dec.state.ny.us/website/dfwmr/wildlife/herp/index.html | | |
| NYS Department of Environmental Conservation - Endangered Species Unit | 8-Jun-99 | WWW | No | www.dec.state.ny.us/website/dfwmr/wildlife/endspec/enspamphib.html | Brief summaries, listed by species, for NY state. | <i>Eurycea longicauda</i> (Longtail Salamander): nocturnal salamander which occupies shallow rocky streams and moist forested areas. Found in Cattaraugus County and mid Hudson Valley. Very few in NY. Status: Special Concern. |
| NYS Department of Environmental Conservation | 8-Jun-99 | WWW | No | www.dec.state.ny.us/website/dfwmr/wildlife/herp/atproj.html | 10 year survey documenting geographic distribution of herpetofauna in NY state. | Common frogs and toads abundant, snapping turtles abundant, some box turtles present. |
| NYSDEC | 16-Jun-99 | Call | Yes | Mark Brown (518) 623-3671 | Familiar with the area regarding mammals, birds, and herps. Good source. See General Info page. | Reports snapping and painted turtles, red back and two-line salamanders. Frogs: bull, spring peepers, gray tree, northern leopard, and pickeral. American toad. Garter and water snakes (none are poisonous). Currently working on a herp survey. |
| Ndakinna Wilderness Project | 6/3/1999 6/16/99 | Email Call Call | No Yes | Jim Brushek (518) 583-9980x3, 23 Middle Grove Road, Greenfield Center, NY 12833; Received address from Saratoga County Information - Annamaria Dalton (annamaria@spa.net) | Professional Tracker | Common amphibians present in strong numbers. Box, snapper, and painted turtles. Some snakes which he could not identify. |

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TABLE 5-1
BENTHIC INVERTEBRATES COLLECTED AT TI POOL STATIONS

| Taxa in Rank Order | Common Name | Mean % of Total Ind. Collected |
|------------------------------|-----------------------------|---------------------------------------|
| <i>Caecidotea racovitzai</i> | Isopod (sowbug) | 34.6 |
| Chironomidae ¹ | Midges | ~30.2 |
| Oligochaeta | Aquatic worms | 14.3 |
| <i>Gammarus fasciatus</i> | Amphipod | 10.3 |
| <i>Pisidium</i> sp. | Pill Clam | 5.0 |
| <i>Canthocamptes</i> sp. | Harpacticoid copepod | 1.5 |
| Nematoda | Nematods (worms) | 1.1 |
| <i>Phylocentropus</i> sp. | Caddis fly larvae | <1.0 |
| <i>Dubiraphia</i> sp. | Beetle larvae | <1.0 |
| <i>Menetus</i> sp. | Caddis fly larvae | <1.0 |
| <i>Valvata</i> sp. | Snail | <1.0 |
| <i>Sialis</i> sp. | Alderfly larvae | <1.0 |
| <i>Oecetis</i> sp. | Caddisfly larvae | <1.0 |
| <i>Probezzia</i> sp. | Biting midges | <1.0 |
| <i>Enallagma</i> sp. | Damselfly nymph | <1.0 |
| Chydoridae | Water fleas (Cladoceran) | <1.0 |
| Acariformes | Mites | <1.0 |
| <i>Amnicola</i> sp. | Snail | <1.0 |
| <i>Mystacides</i> sp. | Caddisfly larvae | <1.0 |
| <i>Diaphanosoma</i> sp. | Water fleas (Cladoceran) | <1.0 |
| Ceratopogonidae | Biting midges | <1.0 |
| <i>Helobdella fusca</i> | Leech | <1.0 |
| Arthropoda | Arthropods | <1.0 |
| <i>Eukiefferiella</i> sp. | Biting Midges | <1.0 |

**TABLE 5-1
BENTHIC INVERTEBRATES COLLECTED AT TI POOL STATIONS**

| Taxa in Rank Order | Common Name | Mean % of Total Ind. Collected |
|-----------------------------|-----------------------------|---------------------------------------|
| Turbellaria | Flatworms | <1.0 |
| <i>Dugesia tigrina</i> | Flatworm | <1.0 |
| <i>Bithynia tentaculata</i> | Snail | <1.0 |
| Trichoptera | Caddisfly larvae | <1.0 |
| <i>Chydorus</i> sp. | Water fleas (Cladoceran) | <1.0 |
| <i>Caenis</i> sp. | Mayfly nymph | <1.0 |
| <i>Physa</i> sp. | Snail | <1.0 |
| <i>Helobdella</i> sp. | Leech | <1.0 |
| <i>Mesocyclops</i> sp. | Cyclopoid copepods | <1.0 |
| <i>Orthotrichia</i> sp. | Caddis fly larvae | <1.0 |
| Aeschnidae | Dragonfly nymph | <1.0 |
| <i>Hexagenia</i> sp. | Mayfly nymph | <1.0 |
| Hirudinea | Leeches | <1.0 |
| <i>Neureclipsis</i> sp. | Caddisfly larvae | <1.0 |
| <i>Culicoides</i> sp. | Mosquito larvae | <1.0 |
| Corixidae | Water boatman | <1.0 |
| <i>Neoperla</i> sp. | Stonefly nymph | <1.0 |
| Caenidae | Mayfly nymph | <1.0 |
| <i>Donacia</i> sp. | Beetle | <1.0 |
| Hemiptera | True bugs | <1.0 |
| <i>Molanna</i> sp. | Caddisfly larvae | <1.0 |
| Copepoda | Copepods | <1.0 |
| Insecta | Insects | <1.0 |
| Baetidae | Mayfly nymph | <1.0 |
| <i>Macronychus</i> sp. | Riffle beetle | <1.0 |
| Tipulidae | Cranefly larvae | <1.0 |

TABLE 5-1
BENTHIC INVERTEBRATES COLLECTED AT TI POOL STATIONS

| Taxa in Rank Order | Common Name | Mean % of Total Ind. Collected |
|---|-------------------------|---------------------------------------|
| <i>Cymatia</i> sp. | Water boatman | <1.0 |
| <i>Notonecta</i> sp. | Water boatman | <1.0 |
| Talitridae | Amphipod | <1.0 |
| <i>Baetis</i> sp. | Mayfly nymph | <1.0 |
| <i>Dromogomphus</i> sp. | Dragonfly nymph | <1.0 |
| <i>Oxyethira</i> sp. | Caddis fly larvae | <1.0 |
| Diptera | Flies and midges | <1.0 |
| <i>Atherix</i> sp. | Snipe fly | <1.0 |
| Tabanidae | Horsefly larvae | <1.0 |
| <i>Elliptio</i> sp. | Eastern elliptio mussel | <1.0 |
| <p>Notes: Taxa are listed in order of absolute abundance. Mean Percent of individuals is based on the mean of Stations 3 to 7. ¹ Chironomidae were primarily composed of Chironominae, <i>Procladius</i> sp., <i>Tanytarsus</i> sp., <i>Dicrotendipes</i> sp., <i>Polypedilum</i> sp., <i>Clinotanypus</i> sp., <i>Tribelos jucundus</i>, and Tanypodinae.</p> | | |

TABLE 5 -2
RELATIVE ABUNDANCE OF FIVE DOMINANT TAXANOMIC GROUPS AT TI POOL STATIONS

| Group/Taxa | Station 3 | | Station 4 | | Station 5 | | Station 6 | | Station 7 | |
|---|---------------------------------|--------------|---------------------------------|--------------|---------------------------------|--------------|---------------------------------|--------------|---------------------------------|--------------|
| | Abundance ind/m ² | Percent |
| Total Dominant Isopoda <i>Caecidotea racovitzai</i> | 653 | 5.6% | 3245 | 24.6% | 14256 | 50.9% | 2347 | 15.2% | 7286 | 60.9% |
| Total Dominant Chironomids | 3775 | 32.3% | 3959 | 30.1% | 7619 | 27.2% | 3277 | 21.3% | 1561 | 13.0% |
| Unidentified Chironomidae | 1398 | 12.0% | 122 | 0.9% | 2232 | 8.0% | 293 | 1.9% | 398 | 3.3% |
| Unidentified Chironominae | 510 | 4.4% | 1490 | 11.3% | 374 | 1.3% | 1378 | 8.9% | 41 | 0.3% |
| <i>Procladius</i> sp. | 479 | 4.1% | 204 | 1.5% | 1474 | 5.3% | 128 | 0.8% | 296 | 2.5% |
| <i>Tanytarsus</i> sp. | 255 | 2.2% | 0 | 0.0% | 1409 | 5.0% | 26 | 0.2% | 0 | 0.0% |
| <i>Dicrotendipes</i> sp. | 479 | 4.1% | 337 | 2.6% | 560 | 2.0% | 38 | 0.2% | 204 | 1.7% |
| <i>Polypedilum</i> sp. | 82 | 0.7% | 102 | 0.8% | 396 | 1.4% | 281 | 1.8% | 224 | 1.9% |
| <i>Clinotanytus</i> sp. | 51 | 0.4% | 133 | 1.0% | 200 | 0.7% | 332 | 2.2% | 194 | 1.6% |
| <i>Tribelos jucundus</i> | 0 | 0.0% | 867 | 6.6% | 0 | 0.0% | 0 | 0.0% | 0 | 0.0% |
| Unidentified Tanypodinae | 112 | 1.0% | 571 | 4.3% | 131 | 0.5% | 38 | 0.2% | 0 | 0.0% |
| <i>Tribelos</i> sp. | 214 | 1.8% | 51 | 0.4% | 194 | 0.7% | 128 | 0.8% | 204 | 1.7% |
| <i>Chironomus</i> sp. | 41 | 0.3% | 41 | 0.3% | 650 | 2.3% | 0 | 0.0% | 0 | 0.0% |
| <i>Cricotopus trifascia</i> | 102 | 0.9% | 41 | 0.3% | 0 | 0.0% | 306 | 2.0% | 0 | 0.0% |
| Unidentified Orthoclaadiinae | 51 | 0.4% | 0 | 0.0% | 0 | 0.0% | 332 | 2.2% | 0 | 0.0% |
| Total Dominant Oligochaeta Unidentified Oligochaeta | 2918 | 25.0% | 2245 | 17.0% | 2681 | 9.6% | 3584 | 23.3% | 71 | 0.6% |
| Total Dominant Amphipoda <i>Gammarus fasciatus</i> | 1030 | 8.8% | 1102 | 8.4% | 682 | 2.4% | 3176 | 20.6% | 2296 | 19.2% |
| Total Dominant Pelecypoda <i>Pisidium</i> sp. | 1245 | 10.6% | 1581 | 12.0% | 49 | 0.2% | 1097 | 7.1% | 0 | 0.0% |
| Subtotals | 9621 | 82.3% | 12132 | 92.1% | 25287 | 90.4% | 13482 | 87.5% | 11214 | 93.7% |
| Total Abundance (all taxa) | 11691 | | 13172 | | 27983 | | 15407 | | 11968 | |

**TABLE 5-3
SUMMARY OF INFAUNA AND TOTAL BENTHOS INDICES – TI POOL**

| Station | Simpson Diversity D_s | | Simpson Dominance I | | Evenness Distribution | | Species Richness | | Abundance No. Ind./Sq M | |
|-----------------------|----------------------------|---------------|------------------------|---------------|-----------------------|---------------|------------------|---------------|----------------------------|---------------|
| | Infauna | Total Benthos | Infauna | Total Benthos | Infauna | Total Benthos | Infauna | Total Benthos | Infauna | Total Benthos |
| 3 | 0.84 | 0.87 | 0.16 | 0.13 | 0.88 | 0.90 | 25 | 27 | 10,008 | 11,691 |
| 4 | 0.79 | 0.83 | 0.21 | 0.17 | 0.84 | 0.87 | 19 | 21 | 8,825 | 13,172 |
| 5 | 0.81 | 0.69 | 0.19 | 0.31 | 0.87 | 0.73 | 17 | 19 | 13,044 | 27,983 |
| 6 | 0.78 | 0.84 | 0.22 | 0.16 | 0.82 | 0.88 | 22 | 24 | 9,884 | 15,407 |
| 7 | 0.84 | 0.57 | 0.16 | 0.43 | 0.95 | 0.61 | 12 | 14 | 2,387 | 11,968 |
| TI Pool Grand Mean | 0.81 | 0.76 | 0.19 | 0.24 | 0.87 | 0.80 | 19 | 21 | 8,830 | 16,044 |

Notes: Total benthos equals the sum of infaunal and epibenthic macroinvertebrates

**TABLE 5-4
RELATIVE PERCENT ABUNDANCE OF MACROINVERTEBRATES -- LOWER HUDSON RIVER**

| Station 12 | | Station 14 | | Station 15 | | Station 17 | | Station 18 | |
|-----------------------------|-------|-----------------------------|-------|----------------------------|-------|---------------------------|-------|---------------------------------|-------|
| Species/Group | % | Species/Group | % | Species/Group | % | Species/Group | % | Species/Group | % |
| Oligochaeta | 42.4% | Chironominae Indet. | 36.1% | Oligochaeta | 22.0% | <i>Hobsonia florida</i> | 36.1% | Oligochaeta | 18.4% |
| Chironominae Indet. | 12.9% | <i>Dicrotendipes sp.</i> | 10.5% | Chydoridae | 17.3% | Oligochaeta | 32.8% | <i>Cyathura polita</i> | 16.5% |
| Chironomidae Indet. | 10.3% | <i>Procladius sp.</i> | 10.2% | <i>Coelotanypus sp.</i> | 14.0% | <i>Gammarus fasciatus</i> | 11.3% | <i>Hobsonia florida</i> | 14.2% |
| <i>Procladius sp.</i> | 8.0% | <i>Polypedilum sp.</i> | 9.0% | Nematoda | 7.3% | <i>Clinotanypus sp.</i> | 6.3% | <i>Hydrobia minuta</i> | 11.5% |
| <i>Polypedilum sp.</i> | 7.1% | <i>Clinotanypus sp.</i> | 6.4% | <i>Clinotanypus sp.</i> | 6.0% | Nematoda | 3.3% | Isopoda | 10.8% |
| <i>Pisidium sp.</i> | 2.9% | Oligochaeta | 4.1% | <i>Polypedilum sp.</i> | 5.3% | <i>Cyathura polita</i> | 2.0% | <i>Clinotanypus sp.</i> | 10.0% |
| <i>Tribelos sp.</i> | 2.9% | <i>Gammarus fasciatus</i> | 2.6% | Acariformes | 4.0% | <i>Coelotanypus sp.</i> | 2.0% | <i>Gammarus fasciatus</i> | 9.7% |
| <i>Cryptotendipes sp.</i> | 2.9% | <i>Pisidium sp.</i> | 2.3% | <i>Dicrotendipes sp.</i> | 4.0% | <i>Procladius sp.</i> | 1.7% | Ostracoda | 4.5% |
| <i>Tanytarsus sp.</i> | 2.3% | Chironomidae Indet. | 2.3% | <i>Cladotanytarsus sp.</i> | 3.3% | Pelecypoda | 1.3% | <i>Neanthes succinea</i> | 1.3% |
| <i>Chironomus sp.</i> | 1.9% | <i>Ammicola limosa</i> | 1.9% | <i>Ammicola sp.</i> | 3.3% | <i>Neanthes succinea</i> | 1.0% | Pelecypoda | 1.3% |
| <i>Gammarus fasciatus</i> | 1.0% | <i>Cladotanytarsus sp.</i> | 1.5% | <i>Synorthocladius sp.</i> | 2.7% | Bryozoa | 0.7% | <i>Procladius sp.</i> | 1.0% |
| Acariformes | 0.6% | <i>Orthotrichia sp.</i> | 1.1% | <i>Pisidium sp.</i> | 2.7% | <i>Balanus improvisus</i> | 0.7% | <i>Rhithropanopeus harrisii</i> | 0.5% |
| Tanypodinae Indet. | 0.6% | Nematoda | 1.1% | <i>Tribelos sp.</i> | 2.0% | Isopoda | 0.3% | <i>Coelotanypus sp.</i> | 0.3% |
| <i>Clinotanypus sp.</i> | 0.6% | Gastropoda | 1.1% | Cyclopoida | 1.3% | Orthoclaadiinae | 0.3% | | |
| Coleoptera | 0.3% | <i>Cricotopus bicinctus</i> | 1.1% | <i>Gammarus fasciatus</i> | 1.3% | <i>Dicrotendipes sp.</i> | 0.3% | | |
| <i>Bithynia tentaculata</i> | 0.3% | <i>Tanytarsus sp.</i> | 1.1% | Hydroptilidae | 1.3% | | | | |
| <i>Valvata sp.</i> | 0.3% | <i>Trianaodes sp.</i> | 0.8% | <i>Cyathura polita</i> | 0.7% | | | | |
| Nematoda | 0.3% | Orthoclaadiinae Indet. | 0.8% | <i>Hydroptila sp.</i> | 0.7% | | | | |
| <i>Cyathura polita</i> | 0.3% | <i>Chironomus sp.</i> | 0.8% | <i>Chironomus sp.</i> | 0.7% | | | | |
| Ostracoda | 0.3% | Acariformes | 0.4% | | | | | | |
| Leptoceridae | 0.3% | <i>Dugesia tigrina</i> | 0.4% | | | | | | |
| Ceratopogonidae | 0.3% | <i>Diaphanosoma sp.</i> | 0.4% | | | | | | |
| Hemiptera | 0.3% | <i>Hydroptila sp.</i> | 0.4% | | | | | | |
| <i>Nilothauma sp.</i> | 0.3% | <i>Probezzia sp.</i> | 0.4% | | | | | | |
| <i>Cryptochironomus sp.</i> | 0.3% | <i>Bithynia tentaculata</i> | 0.4% | | | | | | |
| | | Tanypodinae Indet. | 0.4% | | | | | | |
| | | <i>Synorthocladius sp.</i> | 0.4% | | | | | | |
| | | <i>Tribelos sp.</i> | 0.4% | | | | | | |
| | | <i>Djalmabatista sp.</i> | 0.4% | | | | | | |
| | | <i>Labrundinia sp.</i> | 0.4% | | | | | | |
| | | <i>Coelotanypus sp.</i> | 0.4% | | | | | | |
| | | <i>Synorthocladius sp.</i> | 0.4% | | | | | | |
| | | <i>Cryptotendipes sp.</i> | 0.4% | | | | | | |

TABLE 5-5
SUMMARY OF DIVERSITY INDICES AND ABUNDANCE DATA – LOWER HUDSON RIVER

| Station | D_s | I | D_{max} | E_s | Species Richness | Abundance ind/m² | Biomass mg/m² |
|-------------------------------|----------------------|----------|------------------------|----------------------|-------------------------|------------------------------------|---------------------------------|
| Station 12 Stockport Flats | 0.70 | 0.30 | 0.92 | 0.76 | 14 | 5,289 | 63 |
| Station 14 Tivoli Bays | 0.82 | 0.18 | 0.95 | 0.86 | 16 | 4,524 | 126 |
| Station 15 Esopus Meadows | 0.86 | 0.14 | 0.93 | 0.93 | 11 | 2,551 | 65 |
| Station 17 Iona Island | 0.71 | 0.29 | 0.90 | 0.79 | 9 | 5,136 | 365 |
| Station 18 Piermont Pier | 0.84 | 0.16 | 0.90 | 0.93 | 9 | 6,480 | 291 |
| Grand Mean | 0.79 | 0.21 | 0.92 | 0.85 | 12 | 4,796 | 182 |

**TABLE 5-6
SELECTED SEDIMENT SCREENING GUIDELINES: PCBs**

| | Total PCBs | Aroclor 1254 | Aroclor 1248 | Aroclor 1016 | Aroclor 1260 | Aroclor 1242 |
|--|------------|--------------|--------------|--------------|--------------|--------------|
| Sediment Guidelines/Effect Levels | | | | | | |
| Hudson River Sediment Effect Concentrations (NOAA, 1999) - mg/kg (ppm) | | | | | | |
| Threshold Effect Concentration | 0.04 | | | | | |
| Mid-range Effect Concentration | 0.4 | | | | | |
| Extreme Effect Concentration | 1.7 | | | | | |
| NYSDEC (1998) Freshwater (µg/g OC) | | | | | | |
| Benthic Aquatic Life Acute Toxicity | 2760.8 | | | | | |
| Benthic Aquatic Life Chronic Toxicity | 19.3 | | | | | |
| Wildlife Bioaccumulation | 1.4 | | | | | |
| Ontario Ministry of the Environment Freshwater Guidelines (Persaud et al., 1993) | | | | | | |
| No Effect Level (µg/g) | 0.01 | | | | | |
| Lowest Effect Level (µg/g) | 0.07 | 0.06 | 0.03 | 0.007 | 0.005 | |
| Severe Effect Level (µg/g OC) | 530 | 34 | 150 | 53 | 24 | |
| Long et al. (1995) Marine & Estuaries- ppb | | | | | | |
| Effects-Range-Low | 22.7 | | | | | |
| Effects-Range-Median | 180 | | | | | |
| Ingersoll et al. (1996) Freshwater Guidelines based on <i>Hyalloella azteca</i> - ppb | | | | | | |
| Effects-Range-Low | 50 | | | | | |
| Effects-Range-Median | 730 | | | | | |
| Threshold Effect Level | 32 | | | | | |
| Probable Effect Level | 240 | | | | | |
| No Effect Concentration | 190 | | | | | |
| Washington State (1997) Freshwater - ppb | | | | | | |
| Probable Apparent Effects Threshold - Microtox | 21 | 7.3 | 21 | | | |
| PAET - <i>Hyalloella azteca</i> | 450 | 240 | | | | 100 |
| Apparent Effects Threshold - Microtox | 21 | 7.3 | | | | |
| AET - <i>Hyalloella azteca</i> | 820 | 350 | | | | 100 |
| Apparent Effects Threshold - Microtox mg/kg OC | 2.6 | 0.73 | | | | |
| AET - <i>Hyalloella azteca</i> mg/kg OC | | 18 | | | | |
| Jones et al. (1997) ppb; Eq-P-derived assuming 1% OC Recommended TOC adjustment | | | | | | |
| Secondary Chronic Values | | 810 | 1000 | | 4500000 | |

Notes: All values are provided in dry weight unless noted
Mean PCB conc.Upper Hudson benthic stations: 9.292 - 29.320 ppm
Mean PCB conc.Lower Hudson benthic stations: 0.367 - 1.313 ppm

TABLE 5-7: FEDERAL AND STATE PCB WATER QUALITY CRITERIA

| Total PCB Water Quality Criteria (µg/L) | Upper Hudson 1993 (µg/L) | |
|--|-----------------------------|---------|
| | Average | Maximum |
| USEPA/NYSDEC - Benthic Aquatic Life | | |
| Acute Toxicity - Freshwater | 2 | |
| Acute Toxicity - Saltwater | 10 | |
| Chronic Toxicity - Freshwater | 0.014 | 0.071 |
| Chronic Toxicity - Saltwater | 0.03 | 0.226 |
| NYSDEC - Wildlife Bioaccumulation | | |
| Freshwater | 0.001 | 0.071 |
| Saltwater | 0.001 | 0.226 |
| NYSDEC Surface Water Standards | | |
| Wildlife Criterion | 0.00012 | 0.071 |

Sources: NYSDEC June, 1998 and March 1998; USEPA, 1991

TABLE 5-8: RATIO OF OBSERVED SEDIMENT CONCENTRATIONS TO GUIDELINES

| Location | TEC | | MEC | | EEC | | NYSDEC Benthic Chronic | | NYSDEC Wildlife | |
|----------------------------|-----------------------|-------------|----------------------|------------|----------------------|-------------|------------------------|------------|-----------------|-------------|
| | 0.04 mg/kg dry weight | | 0.4 mg/kg dry weight | | 1.7 mg/kg dry weight | | 19.3 mg/kg OC | | 1.4 mg/kg OC | |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| | Sediment | Sediment | Sediment | Sediment | Sediment | Sediment | Sediment | Sediment | Sediment | Sediment |
| <i>Upper River</i> | | | | | | | | | | |
| Thompson Island Pool (189) | 297 | 435 | 30 | 43 | 7.0 | 10.2 | 12 | 16 | 169 | 215 |
| Stillwater (168) | 776 | 1354 | 78 | 135 | 18 | 32 | 43 | 74 | 587 | 1025 |
| Federal Dam (154) | 70 | 117 | 7.0 | 12 | 1.6 | 2.8 | 9.2 | 15 | 127 | 213 |
| <i>Lower River</i> | | | | | | | | | | |
| 143.5 | 22 | 24 | 2.2 | 2.4 | 0.5 | 0.6 | 1.4 | 1.4 | 20 | 20 |
| 137.2 | 38 | 77 | 3.8 | 7.7 | 0.9 | 1.8 | 2.8 | 5.7 | 39 | 79 |
| 122.4 | 24 | 27 | 2.4 | 2.7 | 0.6 | 0.6 | 2.4 | 2.6 | 33 | 36 |
| 113.8 | 25 | 42 | 2.5 | 4.2 | 0.6 | 1.0 | 1.6 | 2.6 | 22 | 37 |
| 100 | 10 | 215 | 1.0 | 22 | 0.2 | 5.1 | 0.9 | 17 | 12 | 241 |
| 88.9 | 20 | 57 | 2.0 | 5.7 | 0.5 | 1.3 | 1.1 | 3.3 | 16 | 45 |
| 58.7 | 6.3 | 70 | 0.6 | 7.0 | 0.1 | 1.6 | 1.1 | 9.6 | 16 | 133 |
| 47.3 | 38 | 150 | 3.8 | 15 | 0.9 | 3.5 | 2.3 | 8.8 | 32 | 121 |
| 25.8 | 14 | 39 | 1.4 | 3.9 | 0.3 | 0.9 | 1.5 | 3.9 | 20 | 54 |

TABLE 5-8: RATIO OF OBSERVED SEDIMENT CONCENTRATIONS TO GUIDELINES

| Location | Persaud LEL 0.007 mg/kg dry weight | | Persaud SEL 53 mg/kg OC | | Washington State PAET 1242 100 ppb | | Washington State PAET Microtox | | Washington State AET Microtox OC | |
|----------------------------|---------------------------------------|---------------------|----------------------------|---------------------|---------------------------------------|---------------------|-----------------------------------|---------------------|-------------------------------------|---------------------|
| | Average Sediment | 95% UCL Sediment | Average Sediment | 95% UCL Sediment | Average Sediment | 95% UCL Sediment | Average Sediment | 95% UCL Sediment | Average Sediment | 95% UCL Sediment |
| <i>Upper River</i> | | | | | | | | | | |
| Thompson Island Pool (189) | 1697 | 2483 | 4.5 | 5.7 | 119 | 174 | 566 | 828 | 91 | 116 |
| Stillwater (168) | 4433 | 7739 | 16 | 27 | 310 | 542 | 1478 | 2580 | 316 | 552 |
| Federal Dam (154) | 399 | 669 | 3.4 | 5.6 | 28 | 47 | 133 | 223 | 69 | 115 |
| <i>Lower River</i> | | | | | | | | | | |
| 143.5 | 123 | 135 | 0.5 | 0.5 | 8.6 | 9.4 | 41 | 45 | 11 | 11 |
| 137.2 | 217 | 438 | 1.0 | 2.1 | 15 | 31 | 72 | 146 | 21 | 43 |
| 122.4 | 138 | 153 | 0.9 | 0.9 | 10 | 11 | 46 | 51 | 18 | 19 |
| 113.8 | 144 | 238 | 0.6 | 1.0 | 10 | 17 | 48 | 79 | 12 | 20 |
| 100 | 57 | 1230 | 0.3 | 6.4 | 4.0 | 86 | 19 | 410 | 6.7 | 130 |
| 88.9 | 112 | 326 | 0.4 | 1.2 | 7.8 | 23 | 37 | 109 | 8.5 | 24 |
| 58.7 | 36 | 399 | 0.4 | 3.5 | 2.5 | 28 | 12 | 133 | 8.4 | 71 |
| 47.3 | 220 | 857 | 0.8 | 3.2 | 15 | 60 | 73 | 286 | 17 | 65 |
| 25.8 | 83 | 223 | 0.5 | 1.4 | 5.8 | 16 | 28 | 74 | 11 | 29 |

TABLE 5-9: RATIO OF HUDTOX PREDICTED SEDIMENT CONCENTRATIONS TO SEDIMENT GUIDELINES

| Year | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | |
|------|----------------------------|-----------|-----------|----------------------|-----------|-----------|---------------------------|-----------|-----------|----------------------|-----------|-----------|---------------------------|-----------|-----------|----------------------|-----------|-----------|
| | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total |
| | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc |
| | TEC: 0.04 mg/kg dry weight | | | | | | MEC: 0.4 mg/kg dry weight | | | | | | EEC: 1.7 mg/kg dry weight | | | | | |
| 1993 | 720 | 232 | 103 | 760 | 233 | 104 | 72 | 23 | 10.3 | 76 | 23 | 10.4 | 17 | 5.5 | 2.4 | 18 | 5.5 | 2.4 |
| 1994 | 667 | 218 | 95 | 704 | 219 | 95 | 67 | 22 | 9.5 | 70 | 22 | 9.5 | 16 | 5.1 | 2.2 | 17 | 5.1 | 2.2 |
| 1995 | 628 | 212 | 93 | 662 | 212 | 93 | 63 | 21 | 9.3 | 66 | 21 | 9.3 | 15 | 5.0 | 2.2 | 16 | 5.0 | 2.2 |
| 1996 | 570 | 190 | 78 | 601 | 191 | 78 | 57 | 19 | 7.8 | 60 | 19 | 7.8 | 13 | 4.5 | 1.8 | 14 | 4.5 | 1.8 |
| 1997 | 527 | 172 | 68 | 555 | 173 | 68 | 53 | 17 | 6.8 | 56 | 17 | 6.8 | 12 | 4.1 | 1.6 | 13 | 4.1 | 1.6 |
| 1998 | 475 | 155 | 58 | 501 | 156 | 59 | 48 | 15 | 5.8 | 50 | 16 | 5.9 | 11 | 3.6 | 1.4 | 12 | 3.7 | 1.4 |
| 1999 | 433 | 141 | 51 | 457 | 142 | 52 | 43 | 14 | 5.1 | 46 | 14 | 5.2 | 10 | 3.3 | 1.2 | 11 | 3.3 | 1.2 |
| 2000 | 396 | 129 | 45 | 417 | 130 | 45 | 40 | 13 | 4.5 | 42 | 13 | 4.5 | 9.3 | 3.0 | 1.0 | 9.8 | 3.1 | 1.1 |
| 2001 | 374 | 127 | 42 | 395 | 127 | 42 | 37 | 13 | 4.2 | 39 | 13 | 4.2 | 8.8 | 3.0 | 1.0 | 9.3 | 3.0 | 1.0 |
| 2002 | 354 | 122 | 40 | 373 | 122 | 40 | 35 | 12 | 4.0 | 37 | 12 | 4.0 | 8.3 | 2.9 | 0.9 | 8.8 | 2.9 | 0.9 |
| 2003 | 330 | 116 | 38 | 348 | 117 | 38 | 33 | 12 | 3.8 | 35 | 12 | 3.8 | 7.8 | 2.7 | 0.9 | 8.2 | 2.7 | 0.9 |
| 2004 | 302 | 108 | 35 | 318 | 109 | 35 | 30 | 11 | 3.5 | 32 | 11 | 3.5 | 7.1 | 2.5 | 0.8 | 7.5 | 2.6 | 0.8 |
| 2005 | 278 | 97 | 30 | 293 | 97 | 31 | 28 | 9.7 | 3.0 | 29 | 9.7 | 3.1 | 6.5 | 2.3 | 0.7 | 6.9 | 2.3 | 0.7 |
| 2006 | 268 | 94 | 29 | 282 | 94 | 29 | 27 | 9.4 | 2.9 | 28 | 9.4 | 2.9 | 6.3 | 2.2 | 0.7 | 6.6 | 2.2 | 0.7 |
| 2007 | 251 | 87 | 26 | 265 | 87 | 26 | 25 | 8.7 | 2.6 | 27 | 8.7 | 2.6 | 5.9 | 2.0 | 0.6 | 6.2 | 2.1 | 0.6 |
| 2008 | 234 | 81 | 24 | 247 | 82 | 24 | 23 | 8.1 | 2.4 | 25 | 8.2 | 2.4 | 5.5 | 1.9 | 0.6 | 5.8 | 1.9 | 0.6 |
| 2009 | 223 | 78 | 23 | 235 | 78 | 25 | 22 | 7.8 | 2.3 | 23 | 7.8 | 2.5 | 5.2 | 1.8 | 0.5 | 5.5 | 1.8 | 0.6 |
| 2010 | 209 | 74 | 22 | 220 | 75 | 22 | 21 | 7.4 | 2.2 | 22 | 7.5 | 2.2 | 4.9 | 1.7 | 0.5 | 5.2 | 1.8 | 0.5 |
| 2011 | 185 | 67 | 20 | 194 | 67 | 20 | 18 | 6.7 | 2.0 | 19 | 6.7 | 2.0 | 4.3 | 1.6 | 0.5 | 4.6 | 1.6 | 0.5 |
| 2012 | 170 | 63 | 19 | 179 | 63 | 19 | 17 | 6.3 | 1.9 | 18 | 6.3 | 1.9 | 4.0 | 1.5 | 0.4 | 4.2 | 1.5 | 0.4 |
| 2013 | 161 | 62 | 18 | 170 | 62 | 18 | 16 | 6.2 | 1.8 | 17 | 6.2 | 1.8 | 3.8 | 1.5 | 0.4 | 4.0 | 1.5 | 0.4 |
| 2014 | 148 | 57 | 16 | 157 | 57 | 16 | 15 | 5.7 | 1.6 | 16 | 5.7 | 1.6 | 3.5 | 1.3 | 0.4 | 3.7 | 1.3 | 0.4 |
| 2015 | 139 | 54 | 15 | 147 | 54 | 15 | 14 | 5.4 | 1.5 | 15 | 5.4 | 1.5 | 3.3 | 1.3 | 0.4 | 3.5 | 1.3 | 0.4 |
| 2016 | 133 | 53 | 15 | 141 | 53 | 15 | 13 | 5.3 | 1.5 | 14 | 5.3 | 1.5 | 3.1 | 1.2 | 0.4 | 3.3 | 1.2 | 0.4 |
| 2017 | 122 | 49 | 13 | 129 | 49 | 13 | 12 | 4.9 | 1.3 | 13 | 4.9 | 1.3 | 2.9 | 1.1 | 0.3 | 3.0 | 1.1 | 0.3 |
| 2018 | 115 | 46 | 12 | 121 | 46 | 12 | 11 | 4.6 | 1.2 | 12 | 4.6 | 1.2 | 2.7 | 1.1 | 0.3 | 2.9 | 1.1 | 0.3 |

TABLE 5-9: RATIO OF HUDTOX PREDICTED SEDIMENT CONCENTRATIONS TO SEDIMENT GUIDELINES

| Year | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | |
|------|--------------------------------------|-----------|-----------|----------------------|-----------|-----------|------------------------------|-----------|-----------|----------------------|-----------|-----------|------------------------------------|-----------|-----------|----------------------|-----------|-----------|
| | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total |
| | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc |
| | NYSDEC Benthic Chronic 19.3 mg/Kg OC | | | | | | NYSDEC Wildlife 1.4 mg/Kg OC | | | | | | Persaud LEL 0.007 mg/Kg dry weight | | | | | |
| 1993 | 74 | 47 | 26 | 84 | 48 | 26 | 1026 | 651 | 361 | 1157 | 655 | 365 | 4116 | 1326 | 589 | 4343 | 1333 | 594 |
| 1994 | 69 | 45 | 24 | 78 | 45 | 24 | 951 | 615 | 336 | 1074 | 616 | 337 | 3812 | 1248 | 542 | 4023 | 1251 | 543 |
| 1995 | 65 | 43 | 24 | 73 | 43 | 24 | 898 | 598 | 326 | 1012 | 598 | 327 | 3588 | 1211 | 529 | 3784 | 1212 | 530 |
| 1996 | 59 | 39 | 19 | 67 | 39 | 20 | 818 | 537 | 268 | 921 | 540 | 271 | 3258 | 1087 | 444 | 3435 | 1092 | 447 |
| 1997 | 55 | 35 | 17 | 62 | 35 | 17 | 757 | 487 | 240 | 852 | 489 | 241 | 3009 | 985 | 387 | 3172 | 988 | 388 |
| 1998 | 50 | 32 | 12 | 56 | 32 | 12 | 684 | 437 | 166 | 770 | 440 | 168 | 2716 | 884 | 333 | 2863 | 890 | 337 |
| 1999 | 45 | 29 | 11 | 51 | 29 | 11 | 624 | 400 | 147 | 703 | 400 | 147 | 2477 | 808 | 294 | 2611 | 810 | 294 |
| 2000 | 41 | 27 | 9.2 | 47 | 27 | 9.3 | 571 | 366 | 127 | 643 | 368 | 128 | 2262 | 739 | 255 | 2385 | 742 | 257 |
| 2001 | 39 | 26 | 8.6 | 44 | 26 | 8.6 | 540 | 360 | 119 | 608 | 360 | 119 | 2139 | 723 | 238 | 2256 | 724 | 238 |
| 2002 | 37 | 25 | 8.3 | 42 | 25 | 8.3 | 511 | 347 | 114 | 576 | 348 | 114 | 2024 | 697 | 228 | 2134 | 698 | 228 |
| 2003 | 35 | 24 | 7.8 | 39 | 24 | 7.8 | 477 | 332 | 108 | 537 | 333 | 108 | 1884 | 665 | 215 | 1986 | 666 | 216 |
| 2004 | 32 | 22 | 7.2 | 36 | 22 | 7.2 | 438 | 308 | 99 | 494 | 310 | 100 | 1726 | 618 | 198 | 1820 | 621 | 199 |
| 2005 | 29 | 20 | 6.3 | 33 | 20 | 6.3 | 404 | 275 | 87 | 456 | 276 | 87 | 1590 | 554 | 174 | 1676 | 556 | 175 |
| 2006 | 28 | 19 | 6.0 | 32 | 19 | 6.0 | 389 | 265 | 83 | 439 | 265 | 83 | 1530 | 535 | 165 | 1614 | 535 | 166 |
| 2007 | 27 | 18 | 5.4 | 30 | 18 | 5.4 | 367 | 245 | 74 | 413 | 246 | 75 | 1437 | 496 | 149 | 1515 | 498 | 150 |
| 2008 | 25 | 17 | 4.9 | 28 | 17 | 4.9 | 342 | 228 | 68 | 385 | 229 | 68 | 1338 | 465 | 136 | 1410 | 466 | 137 |
| 2009 | 24 | 16 | 4.7 | 27 | 16 | 4.7 | 326 | 217 | 65 | 367 | 217 | 65 | 1275 | 443 | 130 | 1343 | 443 | 141 |
| 2010 | 22 | 15 | 4.6 | 25 | 15 | 4.6 | 306 | 208 | 63 | 343 | 208 | 63 | 1193 | 425 | 126 | 1257 | 426 | 127 |
| 2011 | 20 | 14 | 4.1 | 22 | 14 | 4.1 | 270 | 186 | 57 | 304 | 187 | 57 | 1055 | 383 | 114 | 1111 | 385 | 114 |
| 2012 | 18 | 13 | 3.8 | 20 | 13 | 3.8 | 248 | 175 | 53 | 280 | 175 | 53 | 970 | 361 | 106 | 1024 | 361 | 106 |
| 2013 | 17 | 12 | 3.8 | 19 | 12 | 3.8 | 234 | 170 | 52 | 265 | 171 | 52 | 919 | 353 | 104 | 970 | 353 | 104 |
| 2014 | 16 | 11 | 3.3 | 18 | 11 | 3.4 | 215 | 155 | 46 | 244 | 156 | 46 | 847 | 324 | 92 | 895 | 325 | 93 |
| 2015 | 15 | 11 | 3.1 | 17 | 11 | 3.2 | 202 | 148 | 43 | 230 | 149 | 43 | 796 | 310 | 87 | 841 | 311 | 87 |
| 2016 | 14 | 10 | 3.1 | 16 | 10 | 3.1 | 192 | 144 | 43 | 219 | 145 | 43 | 759 | 303 | 85 | 803 | 303 | 85 |
| 2017 | 13 | 10 | 2.6 | 15 | 10 | 2.6 | 176 | 131 | 36 | 201 | 132 | 36 | 695 | 278 | 72 | 736 | 279 | 73 |
| 2018 | 12 | 8.9 | 2.4 | 14 | 8.9 | 2.4 | 166 | 122 | 33 | 189 | 123 | 33 | 655 | 261 | 67 | 693 | 261 | 67 |

TABLE 5-9: RATIO OF HUDTOX PREDICTED SEDIMENT CONCENTRATIONS TO SEDIMENT GUIDELINES

| Year | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | | Average PCB Results | | | Tri+ 95% UCL Results | | |
|------|-------------------------|-----------|-----------|----------------------|-----------|-----------|---|-----------|-----------|----------------------|-----------|-----------|---|-----------|-----------|----------------------|-----------|-----------|
| | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total | 189 Total | 168 Total | 154 Total |
| | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc | Sed Conc |
| | Persaud SEL 53 mg/Kg OC | | | | | | Washington State PAET 1242 0.1 mg/Kg dry weight | | | | | | Washington PAET Microtox 0.021 mg/Kg dry weight | | | | | |
| 1993 | 27 | 17 | 10 | 31 | 17 | 10 | 288 | 93 | 41 | 304 | 93 | 42 | 1372 | 442 | 196 | 1448 | 444 | 198 |
| 1994 | 25 | 16 | 8.9 | 28 | 16 | 8.9 | 267 | 87 | 38 | 282 | 88 | 38 | 1271 | 416 | 181 | 1341 | 417 | 181 |
| 1995 | 24 | 16 | 8.6 | 27 | 16 | 8.6 | 251 | 85 | 37 | 265 | 85 | 37 | 1196 | 404 | 176 | 1261 | 404 | 177 |
| 1996 | 22 | 14 | 7.1 | 24 | 14 | 7.2 | 228 | 76 | 31 | 240 | 76 | 31 | 1086 | 362 | 148 | 1145 | 364 | 149 |
| 1997 | 20 | 13 | 6.3 | 23 | 13 | 6.4 | 211 | 69 | 27 | 222 | 69 | 27 | 1003 | 328 | 129 | 1057 | 329 | 129 |
| 1998 | 18 | 12 | 4.4 | 20 | 12 | 4.4 | 190 | 62 | 23 | 200 | 62 | 24 | 905 | 295 | 111 | 954 | 297 | 112 |
| 1999 | 16 | 11 | 3.9 | 19 | 11 | 3.9 | 173 | 57 | 21 | 183 | 57 | 21 | 826 | 269 | 98 | 870 | 270 | 98 |
| 2000 | 15 | 10 | 3.4 | 17 | 10 | 3.4 | 158 | 52 | 18 | 167 | 52 | 18 | 754 | 246 | 85 | 795 | 247 | 86 |
| 2001 | 14 | 10 | 3.1 | 16 | 10 | 3.1 | 150 | 51 | 17 | 158 | 51 | 17 | 713 | 241 | 79 | 752 | 241 | 79 |
| 2002 | 14 | 9.2 | 3.0 | 15 | 9.2 | 3.0 | 142 | 49 | 16 | 149 | 49 | 16 | 675 | 232 | 76 | 711 | 233 | 76 |
| 2003 | 13 | 8.8 | 2.8 | 14 | 8.8 | 2.9 | 132 | 47 | 15 | 139 | 47 | 15 | 628 | 222 | 72 | 662 | 222 | 72 |
| 2004 | 12 | 8.1 | 2.6 | 13 | 8.2 | 2.6 | 121 | 43 | 14 | 127 | 43 | 14 | 575 | 206 | 66 | 607 | 207 | 66 |
| 2005 | 11 | 7.3 | 2.3 | 12 | 7.3 | 2.3 | 111 | 39 | 12 | 117 | 39 | 12 | 530 | 185 | 58 | 559 | 185 | 58 |
| 2006 | 10 | 7.0 | 2.2 | 12 | 7.0 | 2.2 | 107 | 37 | 12 | 113 | 37 | 12 | 510 | 178 | 55 | 538 | 178 | 55 |
| 2007 | 10 | 6.5 | 2.0 | 11 | 6.5 | 2.0 | 101 | 35 | 10 | 106 | 35 | 10 | 479 | 165 | 50 | 505 | 166 | 50 |
| 2008 | 9.0 | 6.0 | 1.8 | 10 | 6.1 | 1.8 | 94 | 33 | 10 | 99 | 33 | 10 | 446 | 155 | 45 | 470 | 155 | 46 |
| 2009 | 8.6 | 5.7 | 1.7 | 10 | 5.7 | 1.7 | 89 | 31 | 9.1 | 94 | 31 | 10 | 425 | 148 | 43 | 448 | 148 | 47 |
| 2010 | 8.1 | 5.5 | 1.7 | 9.1 | 5.5 | 1.7 | 84 | 30 | 8.8 | 88 | 30 | 8.9 | 398 | 142 | 42 | 419 | 142 | 42 |
| 2011 | 7.1 | 4.9 | 1.5 | 8.0 | 4.9 | 1.5 | 74 | 27 | 8.0 | 78 | 27 | 8.0 | 352 | 128 | 38 | 370 | 128 | 38 |
| 2012 | 6.5 | 4.6 | 1.4 | 7.4 | 4.6 | 1.4 | 68 | 25 | 7.4 | 72 | 25 | 7.4 | 323 | 120 | 35 | 341 | 120 | 35 |
| 2013 | 6.2 | 4.5 | 1.4 | 7.0 | 4.5 | 1.4 | 64 | 25 | 7.3 | 68 | 25 | 7.3 | 306 | 118 | 35 | 323 | 118 | 35 |
| 2014 | 5.7 | 4.1 | 1.2 | 6.4 | 4.1 | 1.2 | 59 | 23 | 6.4 | 63 | 23 | 6.5 | 282 | 108 | 31 | 298 | 108 | 31 |
| 2015 | 5.3 | 3.9 | 1.1 | 6.1 | 3.9 | 1.1 | 56 | 22 | 6.1 | 59 | 22 | 6.1 | 265 | 103 | 29 | 280 | 104 | 29 |
| 2016 | 5.1 | 3.8 | 1.1 | 5.8 | 3.8 | 1.1 | 53 | 21 | 6.0 | 56 | 21 | 6.0 | 253 | 101 | 28 | 268 | 101 | 28 |
| 2017 | 4.6 | 3.5 | 1.0 | 5.3 | 3.5 | 1.0 | 49 | 19 | 5.1 | 52 | 20 | 5.1 | 232 | 93 | 24 | 245 | 93 | 24 |
| 2018 | 4.4 | 3.2 | 0.9 | 5.0 | 3.2 | 0.9 | 46 | 18 | 4.7 | 49 | 18 | 4.7 | 218 | 87 | 22 | 231 | 87 | 22 |

TABLE 5-10: RATIO OF MEASURED WHOLE WATER CONCENTRATIONS TO BENCHMARKS

| Location | USEPA/NYSDEC - Benthic Aquatic Life 0.014 µg/L freshwater and 0.03 saltwater | | NYSDEC - Wildlife Bioaccumulation 0.001 µg/L | | USEPA Wildlife Criterion 1.2E-04 µg/L | |
|----------------------------|---|----------------|---|----------------|--|----------------|
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| | Conc. in Water | Conc. In Water | Conc. in Water | Conc. In Water | Conc. in Water | Conc. In Water |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 5.3 | 17 | 74 | 233 | 613 | 1942 |
| Stillwater (168) | 9.3 | 30 | 131 | 415 | 1090 | 3458 |
| Federal Dam (154) | 6.5 | 14 | 91 | 196 | 762 | 1634 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 5.1 | 55 | 71 | 770 | 589 | 6420 |
| 137.2 | 5.1 | 55 | 71 | 770 | 589 | 6420 |
| 122.4 | 2.3 | 30 | 32 | 415 | 270 | 3460 |
| 113.8 | 2.3 | 30 | 32 | 415 | 270 | 3460 |
| 100 | 2.3 | 30 | 32 | 415 | 270 | 3460 |
| 88.9 | 1.5 | 6.8 | 21 | 95 | 178 | 790 |
| 58.7 | 0.7 | 3.2 | 21 | 95 | 178 | 790 |
| 47.3 | 0.7 | 3.2 | 21 | 95 | 178 | 790 |
| 25.8 | 0.7 | 3.2 | 21 | 95 | 178 | 790 |

Notes:

Source: TAMS/Gradient Database Release 4.1b

TABLE 5-11: RATIO OF HUDTOX PREDICTED WHOLE WATER CONCENTRATIONS TO CRITERIA AND BENCHMARKS

| Year | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | | Tri+ Average PCB Results | | | Tri+ 95% UCL Results | | |
|------|--|------------------|------------------|----------------------|------------------|------------------|--|------------------|------------------|----------------------|------------------|------------------|---------------------------------------|------------------|------------------|----------------------|------------------|------------------|
| | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 | 189 | 168 | 154 |
| | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc | Whole Water Conc |
| | USEPA/NYSDEC - Benthic Aquatic Life 0.014 ug/L | | | | | | NYSDEC - Wildlife Bioaccumulation 0.001 ug/L | | | | | | USEPA Wildlife Criterion 1.2E-04 ug/l | | | | | |
| 1993 | 3.5 | 5.2 | 3.5 | 3.5 | 5.3 | 0.8 | 49 | 73 | 49 | 49 | 75 | 11 | 41 | 61 | 41 | 41 | 62 | 9 |
| 1994 | 3.0 | 4.0 | 2.7 | 3.0 | 4.2 | 0.6 | 41 | 57 | 38 | 42 | 59 | 8 | 35 | 47 | 32 | 35 | 49 | 7 |
| 1995 | 4.7 | 6.9 | 4.3 | 4.7 | 7.1 | 1.0 | 66 | 96 | 60 | 66 | 99 | 14 | 55 | 80 | 50 | 55 | 82 | 12 |
| 1996 | 1.9 | 2.8 | 1.8 | 1.9 | 3.2 | 0.5 | 26 | 40 | 26 | 26 | 44 | 7 | 22 | 33 | 21 | 22 | 37 | 6 |
| 1997 | 2.0 | 3.5 | 2.3 | 2.0 | 3.5 | 0.7 | 28 | 48 | 32 | 28 | 49 | 9 | 23 | 40 | 27 | 23 | 41 | 8 |
| 1998 | 2.1 | 3.5 | 1.9 | 2.1 | 3.6 | 2.0 | 29 | 50 | 27 | 29 | 50 | 27 | 24 | 41 | 22 | 24 | 42 | 23 |
| 1999 | 1.7 | 3.1 | 1.8 | 1.7 | 3.2 | 1.9 | 24 | 43 | 25 | 24 | 44 | 27 | 20 | 36 | 21 | 20 | 37 | 22 |
| 2000 | 1.4 | 2.6 | 1.6 | 1.5 | 2.7 | 1.7 | 20 | 36 | 22 | 21 | 38 | 24 | 17 | 30 | 18 | 17 | 31 | 20 |
| 2001 | 1.8 | 3.0 | 1.8 | 1.9 | 3.1 | 1.9 | 26 | 42 | 26 | 27 | 43 | 26 | 22 | 35 | 21 | 22 | 36 | 22 |
| 2002 | 1.5 | 2.8 | 1.6 | 1.6 | 2.9 | 1.7 | 22 | 39 | 22 | 22 | 41 | 24 | 18 | 33 | 19 | 18 | 34 | 20 |
| 2003 | 1.2 | 2.2 | 1.3 | 1.2 | 2.3 | 1.5 | 17 | 31 | 19 | 17 | 32 | 21 | 14 | 26 | 16 | 14 | 27 | 17 |
| 2004 | 1.1 | 2.1 | 1.4 | 1.1 | 2.2 | 1.6 | 16 | 29 | 19 | 16 | 31 | 22 | 13 | 24 | 16 | 13 | 26 | 18 |
| 2005 | 1.4 | 2.5 | 1.5 | 1.4 | 2.5 | 1.6 | 19 | 34 | 22 | 19 | 35 | 23 | 16 | 29 | 18 | 16 | 29 | 19 |
| 2006 | 1.4 | 2.5 | 1.5 | 1.4 | 2.5 | 1.6 | 19 | 34 | 22 | 19 | 35 | 23 | 16 | 29 | 18 | 16 | 29 | 19 |
| 2007 | 0.7 | 1.3 | 0.7 | 0.7 | 1.4 | 0.8 | 10 | 19 | 10 | 10 | 19 | 11 | 8.5 | 15 | 8.5 | 8.6 | 16 | 9.2 |
| 2008 | 1.0 | 1.7 | 1.1 | 1.0 | 1.8 | 1.2 | 14 | 24 | 15 | 14 | 26 | 17 | 11 | 20 | 13 | 12 | 21 | 14 |
| 2009 | 1.2 | 2.0 | 1.1 | 1.2 | 2.1 | 1.3 | 16 | 28 | 16 | 17 | 29 | 18 | 14 | 23 | 13 | 14 | 24 | 15 |
| 2010 | 0.7 | 1.3 | 0.7 | 0.7 | 1.3 | 0.8 | 10 | 18 | 10 | 10 | 19 | 11 | 8.0 | 15 | 8.5 | 8.1 | 16 | 9.4 |
| 2011 | 0.7 | 1.3 | 0.8 | 0.7 | 1.4 | 0.9 | 9.4 | 18 | 11 | 10 | 20 | 13 | 7.8 | 15 | 9.2 | 8.0 | 16 | 11 |
| 2012 | 0.9 | 1.7 | 1.0 | 0.9 | 1.7 | 1.1 | 12 | 23 | 14 | 12 | 24 | 15 | 10 | 19 | 12 | 10 | 20 | 12 |
| 2013 | 0.6 | 1.2 | 0.6 | 0.6 | 1.2 | 0.7 | 8.8 | 16 | 8.7 | 8.9 | 17 | 9.5 | 7.3 | 14 | 7.3 | 7.5 | 14 | 7.9 |
| 2014 | 0.5 | 1.3 | 0.8 | 0.5 | 1.4 | 0.8 | 7.4 | 19 | 11 | 8 | 19 | 11 | 6.2 | 16 | 8.9 | 6.3 | 16 | 9.3 |
| 2015 | 0.5 | 1.0 | 0.6 | 0.5 | 1.0 | 0.6 | 7.2 | 14 | 7.9 | 7.3 | 14 | 8.5 | 6.0 | 11 | 6.6 | 6.1 | 12 | 7.1 |
| 2016 | 0.9 | 1.7 | 1.0 | 0.9 | 1.7 | 1.0 | 12 | 24 | 14 | 12 | 24 | 14 | 10 | 20 | 11 | 10 | 20 | 12 |
| 2017 | 0.4 | 0.8 | 0.4 | 0.4 | 0.8 | 0.5 | 5.5 | 11 | 6.2 | 5.6 | 12 | 7.0 | 4.6 | 9.0 | 5.2 | 4.7 | 10 | 5.8 |
| 2018 | 0.5 | 1.0 | 0.6 | 0.5 | 1.0 | 0.7 | 6.6 | 14 | 8.8 | 6.8 | 14 | 9.5 | 5.5 | 12 | 7.3 | 5.6 | 12 | 7.9 |

**TABLE 5-12: RATIO OF MEASURED FORAGE FISH CONCENTRATIONS
TO TOXICITY BENCHMARKS**

| Location | Pumpkinseed field-based NOAEL | | Spottail shiner lab-based NOAEL | | Spottail shiner lab-based LOAEL | |
|----------------------------|-------------------------------|---------------|---------------------------------|---------------|---------------------------------|---------------|
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| | Forage Fish | Forage Fish | Forage Fish | Forage Fish | Forage Fish | Forage Fish |
| | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg | Conc mg/Kg |
| <i>Upper River</i> | | | | | | |
| Thompson Island Pool (189) | 42 | 85 | 1.4 | 2.8 | 0.1 | 0.3 |
| Stillwater (168) | 14 | 20 | 0.5 | 0.7 | 0.04 | 0.06 |
| Federal Dam (154) | 3.3 | 4.8 | 0.1 | 0.2 | 0.01 | 0.01 |
| <i>Lower River</i> | | | | | | |
| 143.5 | 3.9 | 4.6 | 0.1 | 0.2 | 0.01 | 0.01 |
| 137.2 | 7.8 | 17 | 0.3 | 0.6 | 0.02 | 0.05 |
| 122.4 | 3.0 | 4.8 | 0.1 | 0.2 | 0.01 | 0.01 |
| 113.8 | 3.1 | 3.2 | 0.1 | 0.1 | 0.009 | 0.01 |
| 100 | 1.4 | 2.3 | 0.05 | 0.08 | 0.004 | 0.007 |
| 88.9 | 2.7 | 3.7 | 0.09 | 0.1 | 0.008 | 0.01 |
| 58.7 | 2.9 | 3.3 | 0.1 | 0.1 | 0.009 | 0.01 |
| 47.3 | 2.6 | 3.5 | 0.09 | 0.1 | 0.008 | 0.01 |
| 25.8 | 2.0 | 2.4 | 0.07 | 0.08 | 0.006 | 0.007 |

Source: TAMS/Gradient Database Release 4.1b

**TABLE 5-13: RATIO OF PREDICTED PUMPKINSEED CONCENTRATIONS TO
FIELD-BASED NOAEL FOR TRI+ PCBS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 7.8 | 20 | 43 | 7.4 | 21 | 49 | 3.7 | 11 | 26 |
| 1994 | 6.4 | 16 | 31 | 6.0 | 16 | 35 | 3.0 | 7.9 | 18 |
| 1995 | 7.0 | 17 | 37 | 6.9 | 19 | 45 | 3.3 | 9.8 | 24 |
| 1996 | 4.9 | 12 | 30 | 4.6 | 12 | 26 | 2.2 | 6.0 | 16 |
| 1997 | 4.2 | 10 | 22 | 4.0 | 10 | 23 | 2.1 | 5.8 | 14 |
| 1998 | 3.6 | 8.5 | 18 | 3.7 | 9.5 | 21 | 1.8 | 4.9 | 12 |
| 1999 | 2.6 | 6.2 | 15 | 2.8 | 6.9 | 15 | 1.2 | 3.2 | 7.7 |
| 2000 | 2.6 | 6.4 | 14 | 2.7 | 6.6 | 14 | 1.1 | 3.0 | 7.0 |
| 2001 | 2.5 | 5.9 | 14 | 2.9 | 7.5 | 16 | 1.3 | 3.6 | 8.5 |
| 2002 | 2.4 | 5.9 | 13 | 2.7 | 6.7 | 13 | 1.2 | 3.1 | 7.0 |
| 2003 | 2.0 | 4.9 | 11 | 2.2 | 5.6 | 12 | 1.0 | 2.7 | 6.8 |
| 2004 | 2.0 | 4.8 | 11 | 2.2 | 5.3 | 11 | 1.1 | 2.9 | 7.1 |
| 2005 | 1.7 | 4.0 | 9.9 | 1.9 | 4.5 | 9.3 | 0.9 | 2.3 | 5.4 |
| 2006 | 1.7 | 4.1 | 9.6 | 1.9 | 4.9 | 11 | 0.8 | 2.5 | 6.0 |
| 2007 | 1.6 | 3.8 | 8.6 | 1.7 | 4.2 | 8.1 | 0.7 | 1.9 | 4.4 |
| 2008 | 1.4 | 3.4 | 8.5 | 1.5 | 3.5 | 6.6 | 0.6 | 1.6 | 4.1 |
| 2009 | 1.5 | 3.7 | 8.2 | 1.7 | 4.3 | 9.7 | 0.7 | 2.1 | 5.3 |
| 2010 | 1.5 | 3.5 | 7.5 | 1.6 | 3.9 | 7.9 | 0.6 | 1.8 | 4.4 |
| 2011 | 1.2 | 2.9 | 6.4 | 1.3 | 3.1 | 6.3 | 0.5 | 1.5 | 3.8 |
| 2012 | 1.2 | 2.9 | 6.3 | 1.5 | 3.7 | 8.1 | 0.6 | 1.7 | 4.3 |
| 2013 | 1.2 | 2.9 | 6.3 | 1.4 | 3.5 | 7.2 | 0.5 | 1.5 | 3.6 |
| 2014 | 1.0 | 2.4 | 5.6 | 1.3 | 3.2 | 6.7 | 0.5 | 1.3 | 3.4 |
| 2015 | 1.0 | 2.4 | 5.2 | 1.2 | 2.9 | 6.1 | 0.5 | 1.3 | 3.1 |
| 2016 | 0.9 | 2.4 | 5.3 | 1.2 | 3.3 | 7.8 | 0.5 | 1.6 | 4.3 |
| 2017 | 0.8 | 2.1 | 4.5 | 1.0 | 2.4 | 5.0 | 0.4 | 1.1 | 3.2 |
| 2018 | 0.8 | 1.9 | 4.4 | 0.9 | 2.1 | 4.3 | 0.3 | 0.9 | 2.3 |

TABLE 5-14: RATIO OF PREDICTED SPOTTAIL SHINER CONCENTRATIONS TO LABORATORY-DERIVED NOAEL FOR TRI+ PCBS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.3 | 0.7 | 1.4 | 0.3 | 0.7 | 1.4 | 0.1 | 0.4 | 0.7 |
| 1994 | 0.3 | 0.6 | 1.3 | 0.2 | 0.6 | 1.0 | 0.1 | 0.3 | 0.5 |
| 1995 | 0.3 | 0.6 | 1.4 | 0.2 | 0.6 | 1.2 | 0.1 | 0.3 | 0.7 |
| 1996 | 0.2 | 0.5 | 1.2 | 0.2 | 0.4 | 0.7 | 0.08 | 0.2 | 0.5 |
| 1997 | 0.2 | 0.4 | 0.9 | 0.2 | 0.4 | 0.7 | 0.08 | 0.2 | 0.4 |
| 1998 | 0.1 | 0.3 | 0.8 | 0.1 | 0.3 | 0.6 | 0.06 | 0.2 | 0.3 |
| 1999 | 0.1 | 0.3 | 0.7 | 0.1 | 0.2 | 0.5 | 0.05 | 0.1 | 0.2 |
| 2000 | 0.1 | 0.3 | 0.7 | 0.1 | 0.2 | 0.4 | 0.04 | 0.1 | 0.2 |
| 2001 | 0.1 | 0.2 | 0.7 | 0.1 | 0.2 | 0.4 | 0.05 | 0.1 | 0.2 |
| 2002 | 0.1 | 0.2 | 0.6 | 0.1 | 0.2 | 0.4 | 0.04 | 0.1 | 0.2 |
| 2003 | 0.08 | 0.2 | 0.6 | 0.09 | 0.2 | 0.4 | 0.04 | 0.09 | 0.2 |
| 2004 | 0.08 | 0.2 | 0.5 | 0.08 | 0.2 | 0.3 | 0.04 | 0.09 | 0.2 |
| 2005 | 0.07 | 0.2 | 0.5 | 0.07 | 0.2 | 0.3 | 0.03 | 0.07 | 0.1 |
| 2006 | 0.07 | 0.2 | 0.5 | 0.07 | 0.2 | 0.3 | 0.03 | 0.07 | 0.2 |
| 2007 | 0.06 | 0.2 | 0.4 | 0.06 | 0.1 | 0.3 | 0.02 | 0.06 | 0.1 |
| 2008 | 0.06 | 0.1 | 0.4 | 0.06 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2009 | 0.06 | 0.1 | 0.4 | 0.06 | 0.1 | 0.3 | 0.03 | 0.06 | 0.1 |
| 2010 | 0.06 | 0.1 | 0.3 | 0.06 | 0.1 | 0.3 | 0.02 | 0.06 | 0.1 |
| 2011 | 0.05 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2012 | 0.05 | 0.1 | 0.3 | 0.06 | 0.1 | 0.2 | 0.02 | 0.06 | 0.1 |
| 2013 | 0.05 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2014 | 0.04 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.04 | 0.1 |
| 2015 | 0.04 | 0.1 | 0.2 | 0.05 | 0.1 | 0.2 | 0.02 | 0.04 | 0.08 |
| 2016 | 0.04 | 0.09 | 0.2 | 0.04 | 0.1 | 0.2 | 0.02 | 0.05 | 0.12 |
| 2017 | 0.03 | 0.08 | 0.2 | 0.04 | 0.08 | 0.1 | 0.01 | 0.03 | 0.08 |
| 2018 | 0.03 | 0.08 | 0.2 | 0.04 | 0.08 | 0.1 | 0.01 | 0.03 | 0.07 |

TABLE 5-15: RATIO OF PREDICTED SPOTTAIL SHINER CONCENTRATIONS TO LABORATORY-DERIVED LOEL FOR TRI+ PCBS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.026 | 0.058 | 0.123 | 0.024 | 0.061 | 0.120 | 0.012 | 0.031 | 0.064 |
| 1994 | 0.024 | 0.054 | 0.111 | 0.021 | 0.049 | 0.089 | 0.010 | 0.023 | 0.044 |
| 1995 | 0.025 | 0.056 | 0.121 | 0.022 | 0.055 | 0.108 | 0.011 | 0.028 | 0.058 |
| 1996 | 0.018 | 0.041 | 0.108 | 0.015 | 0.036 | 0.064 | 0.007 | 0.018 | 0.041 |
| 1997 | 0.015 | 0.035 | 0.082 | 0.015 | 0.031 | 0.058 | 0.007 | 0.017 | 0.034 |
| 1998 | 0.012 | 0.029 | 0.073 | 0.012 | 0.028 | 0.053 | 0.006 | 0.013 | 0.029 |
| 1999 | 0.009 | 0.023 | 0.064 | 0.010 | 0.022 | 0.040 | 0.004 | 0.010 | 0.020 |
| 2000 | 0.009 | 0.023 | 0.059 | 0.008 | 0.020 | 0.036 | 0.003 | 0.008 | 0.017 |
| 2001 | 0.009 | 0.021 | 0.057 | 0.010 | 0.022 | 0.040 | 0.004 | 0.010 | 0.021 |
| 2002 | 0.009 | 0.021 | 0.054 | 0.009 | 0.021 | 0.037 | 0.004 | 0.009 | 0.018 |
| 2003 | 0.007 | 0.018 | 0.049 | 0.008 | 0.017 | 0.032 | 0.003 | 0.008 | 0.018 |
| 2004 | 0.007 | 0.018 | 0.047 | 0.007 | 0.017 | 0.030 | 0.003 | 0.008 | 0.018 |
| 2005 | 0.006 | 0.015 | 0.042 | 0.006 | 0.014 | 0.025 | 0.003 | 0.006 | 0.013 |
| 2006 | 0.006 | 0.015 | 0.042 | 0.006 | 0.014 | 0.026 | 0.003 | 0.006 | 0.014 |
| 2007 | 0.006 | 0.014 | 0.037 | 0.006 | 0.013 | 0.023 | 0.002 | 0.005 | 0.011 |
| 2008 | 0.005 | 0.013 | 0.036 | 0.005 | 0.011 | 0.020 | 0.002 | 0.005 | 0.010 |
| 2009 | 0.005 | 0.013 | 0.034 | 0.006 | 0.013 | 0.023 | 0.002 | 0.006 | 0.013 |
| 2010 | 0.005 | 0.012 | 0.030 | 0.005 | 0.013 | 0.023 | 0.002 | 0.005 | 0.011 |
| 2011 | 0.004 | 0.011 | 0.027 | 0.004 | 0.009 | 0.017 | 0.002 | 0.004 | 0.009 |
| 2012 | 0.004 | 0.010 | 0.027 | 0.005 | 0.012 | 0.022 | 0.002 | 0.005 | 0.011 |
| 2013 | 0.004 | 0.010 | 0.026 | 0.004 | 0.011 | 0.019 | 0.002 | 0.004 | 0.008 |
| 2014 | 0.004 | 0.009 | 0.022 | 0.004 | 0.010 | 0.018 | 0.002 | 0.004 | 0.009 |
| 2015 | 0.004 | 0.008 | 0.021 | 0.004 | 0.009 | 0.016 | 0.002 | 0.004 | 0.007 |
| 2016 | 0.003 | 0.008 | 0.022 | 0.004 | 0.009 | 0.018 | 0.002 | 0.004 | 0.011 |
| 2017 | 0.003 | 0.007 | 0.019 | 0.003 | 0.007 | 0.013 | 0.001 | 0.003 | 0.007 |
| 2018 | 0.003 | 0.007 | 0.020 | 0.003 | 0.007 | 0.013 | 0.001 | 0.003 | 0.006 |

**TABLE 5-16: RATIO OF PREDICTED PUMPKINSEED CONCENTRATIONS TO
LABORATORY-DERIVED NOAEL ON A TEQ BASIS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.8 | 2.2 | 5.8 | 0.8 | 2.3 | 6.1 | 0.004 | 0.007 | 0.018 |
| 1994 | 0.6 | 1.7 | 4.4 | 0.6 | 1.8 | 4.5 | 0.003 | 0.007 | 0.016 |
| 1995 | 0.7 | 1.9 | 5.0 | 0.7 | 2.1 | 5.6 | 0.003 | 0.007 | 0.017 |
| 1996 | 0.5 | 1.3 | 4.0 | 0.5 | 1.3 | 3.2 | 0.002 | 0.005 | 0.012 |
| 1997 | 0.4 | 1.1 | 2.9 | 0.4 | 1.1 | 2.9 | 0.003 | 0.005 | 0.012 |
| 1998 | 0.4 | 1.0 | 2.5 | 0.4 | 1.1 | 2.7 | 0.002 | 0.004 | 0.010 |
| 1999 | 0.3 | 0.7 | 2.1 | 0.3 | 0.8 | 1.9 | 0.002 | 0.004 | 0.009 |
| 2000 | 0.3 | 0.7 | 2.1 | 0.3 | 0.7 | 1.8 | 0.002 | 0.004 | 0.008 |
| 2001 | 0.3 | 0.7 | 1.9 | 0.3 | 0.8 | 2.1 | 0.002 | 0.004 | 0.008 |
| 2002 | 0.3 | 0.7 | 1.8 | 0.3 | 0.7 | 1.8 | 0.002 | 0.003 | 0.007 |
| 2003 | 0.2 | 0.5 | 1.6 | 0.2 | 0.6 | 1.5 | 0.002 | 0.003 | 0.007 |
| 2004 | 0.2 | 0.5 | 1.5 | 0.2 | 0.6 | 1.5 | 0.001 | 0.003 | 0.006 |
| 2005 | 0.2 | 0.5 | 1.4 | 0.2 | 0.5 | 1.2 | 0.001 | 0.003 | 0.006 |
| 2006 | 0.2 | 0.5 | 1.4 | 0.2 | 0.5 | 1.4 | 0.001 | 0.002 | 0.005 |
| 2007 | 0.2 | 0.4 | 1.2 | 0.2 | 0.5 | 1.1 | 0.002 | 0.003 | 0.005 |
| 2008 | 0.1 | 0.4 | 1.1 | 0.2 | 0.4 | 0.9 | 0.001 | 0.002 | 0.004 |
| 2009 | 0.2 | 0.4 | 1.1 | 0.2 | 0.5 | 1.2 | 0.001 | 0.002 | 0.004 |
| 2010 | 0.2 | 0.4 | 1.0 | 0.2 | 0.4 | 1.0 | 0.001 | 0.002 | 0.005 |
| 2011 | 0.1 | 0.3 | 0.9 | 0.1 | 0.3 | 0.8 | 0.001 | 0.002 | 0.004 |
| 2012 | 0.1 | 0.3 | 0.9 | 0.1 | 0.4 | 1.0 | 0.001 | 0.002 | 0.004 |
| 2013 | 0.1 | 0.3 | 0.9 | 0.2 | 0.4 | 0.9 | 0.001 | 0.002 | 0.004 |
| 2014 | 0.1 | 0.3 | 0.8 | 0.1 | 0.4 | 0.9 | 0.001 | 0.002 | 0.004 |
| 2015 | 0.1 | 0.3 | 0.7 | 0.1 | 0.3 | 0.8 | 0.001 | 0.002 | 0.003 |
| 2016 | 0.1 | 0.3 | 0.7 | 0.1 | 0.4 | 1.0 | 0.001 | 0.002 | 0.004 |
| 2017 | 0.1 | 0.2 | 0.6 | 0.1 | 0.3 | 0.6 | 0.001 | 0.002 | 0.003 |
| 2018 | 0.1 | 0.2 | 0.6 | 0.1 | 0.2 | 0.6 | 0.001 | 0.001 | 0.003 |

**TABLE 5-17: RATIO OF PREDICTED PUMPKINSEED CONCENTRATIONS TO
LABORATORY-DERIVED LOEL ON A TEQ BASIS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.4 | 1.05 | 2.8 | 0.4 | 1.1 | 3.0 | 0.002 | 0.004 | 0.009 |
| 1994 | 0.3 | 0.8 | 2.1 | 0.3 | 0.9 | 2.2 | 0.002 | 0.003 | 0.008 |
| 1995 | 0.3 | 0.9 | 2.4 | 0.3 | 1.0 | 2.7 | 0.002 | 0.003 | 0.008 |
| 1996 | 0.2 | 0.6 | 1.9 | 0.2 | 0.6 | 1.6 | 0.001 | 0.002 | 0.006 |
| 1997 | 0.2 | 0.5 | 1.4 | 0.2 | 0.6 | 1.4 | 0.001 | 0.002 | 0.006 |
| 1998 | 0.2 | 0.5 | 1.2 | 0.2 | 0.5 | 1.3 | 0.001 | 0.002 | 0.005 |
| 1999 | 0.1 | 0.3 | 1.0 | 0.1 | 0.4 | 0.9 | 0.001 | 0.002 | 0.004 |
| 2000 | 0.1 | 0.4 | 1.0 | 0.1 | 0.4 | 0.9 | 0.001 | 0.002 | 0.004 |
| 2001 | 0.1 | 0.3 | 0.9 | 0.1 | 0.4 | 1.0 | 0.001 | 0.002 | 0.004 |
| 2002 | 0.1 | 0.3 | 0.9 | 0.1 | 0.4 | 0.9 | 0.001 | 0.001 | 0.003 |
| 2003 | 0.1 | 0.3 | 0.8 | 0.11 | 0.3 | 0.7 | 0.001 | 0.002 | 0.003 |
| 2004 | 0.1 | 0.3 | 0.7 | 0.11 | 0.3 | 0.7 | 0.001 | 0.001 | 0.003 |
| 2005 | 0.09 | 0.2 | 0.7 | 0.10 | 0.2 | 0.6 | 0.001 | 0.001 | 0.003 |
| 2006 | 0.09 | 0.2 | 0.7 | 0.10 | 0.3 | 0.7 | 0.001 | 0.001 | 0.003 |
| 2007 | 0.08 | 0.2 | 0.6 | 0.09 | 0.2 | 0.5 | 0.001 | 0.001 | 0.003 |
| 2008 | 0.07 | 0.2 | 0.5 | 0.08 | 0.2 | 0.4 | 0.001 | 0.001 | 0.002 |
| 2009 | 0.08 | 0.2 | 0.5 | 0.08 | 0.2 | 0.6 | 0.001 | 0.001 | 0.002 |
| 2010 | 0.07 | 0.2 | 0.5 | 0.08 | 0.2 | 0.5 | 0.001 | 0.001 | 0.002 |
| 2011 | 0.06 | 0.2 | 0.4 | 0.07 | 0.2 | 0.4 | 0.001 | 0.001 | 0.002 |
| 2012 | 0.06 | 0.2 | 0.4 | 0.07 | 0.2 | 0.5 | 0.001 | 0.001 | 0.002 |
| 2013 | 0.06 | 0.2 | 0.4 | 0.07 | 0.2 | 0.5 | 0.001 | 0.001 | 0.002 |
| 2014 | 0.05 | 0.1 | 0.4 | 0.06 | 0.2 | 0.4 | 0.001 | 0.001 | 0.002 |
| 2015 | 0.05 | 0.1 | 0.3 | 0.06 | 0.2 | 0.4 | 0.001 | 0.001 | 0.002 |
| 2016 | 0.05 | 0.1 | 0.3 | 0.06 | 0.2 | 0.5 | 0.001 | 0.001 | 0.002 |
| 2017 | 0.04 | 0.1 | 0.3 | 0.05 | 0.1 | 0.3 | 0.000 | 0.001 | 0.001 |
| 2018 | 0.04 | 0.1 | 0.3 | 0.05 | 0.1 | 0.3 | 0.000 | 0.001 | 0.001 |

TABLE 5-18: RATIO OF PREDICTED SPOTTAIL SHINER CONCENTRATIONS TO LABORATORY-DERIVED NOAEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.092 | 0.233 | 0.690 | 0.088 | 0.248 | 0.655 | 0.005 | 0.012 | 0.029 |
| 1994 | 0.081 | 0.215 | 0.615 | 0.076 | 0.198 | 0.497 | 0.005 | 0.011 | 0.026 |
| 1995 | 0.089 | 0.228 | 0.680 | 0.079 | 0.221 | 0.583 | 0.005 | 0.011 | 0.027 |
| 1996 | 0.060 | 0.168 | 0.599 | 0.055 | 0.142 | 0.363 | 0.004 | 0.008 | 0.021 |
| 1997 | 0.053 | 0.144 | 0.455 | 0.050 | 0.128 | 0.338 | 0.004 | 0.008 | 0.019 |
| 1998 | 0.045 | 0.118 | 0.382 | 0.044 | 0.113 | 0.294 | 0.003 | 0.006 | 0.016 |
| 1999 | 0.032 | 0.096 | 0.337 | 0.035 | 0.088 | 0.231 | 0.003 | 0.006 | 0.014 |
| 2000 | 0.033 | 0.094 | 0.320 | 0.032 | 0.081 | 0.201 | 0.003 | 0.006 | 0.014 |
| 2001 | 0.031 | 0.088 | 0.304 | 0.035 | 0.088 | 0.226 | 0.003 | 0.005 | 0.012 |
| 2002 | 0.031 | 0.086 | 0.281 | 0.034 | 0.083 | 0.206 | 0.002 | 0.005 | 0.011 |
| 2003 | 0.026 | 0.073 | 0.255 | 0.028 | 0.070 | 0.184 | 0.002 | 0.005 | 0.011 |
| 2004 | 0.026 | 0.071 | 0.240 | 0.028 | 0.067 | 0.168 | 0.002 | 0.004 | 0.010 |
| 2005 | 0.022 | 0.062 | 0.218 | 0.022 | 0.054 | 0.142 | 0.002 | 0.004 | 0.009 |
| 2006 | 0.022 | 0.061 | 0.216 | 0.024 | 0.058 | 0.148 | 0.002 | 0.004 | 0.009 |
| 2007 | 0.020 | 0.056 | 0.186 | 0.021 | 0.052 | 0.128 | 0.002 | 0.004 | 0.009 |
| 2008 | 0.019 | 0.053 | 0.189 | 0.019 | 0.045 | 0.112 | 0.002 | 0.003 | 0.007 |
| 2009 | 0.019 | 0.052 | 0.181 | 0.020 | 0.050 | 0.136 | 0.002 | 0.003 | 0.008 |
| 2010 | 0.018 | 0.049 | 0.164 | 0.020 | 0.050 | 0.129 | 0.002 | 0.004 | 0.008 |
| 2011 | 0.015 | 0.043 | 0.150 | 0.015 | 0.038 | 0.096 | 0.002 | 0.003 | 0.007 |
| 2012 | 0.015 | 0.042 | 0.143 | 0.018 | 0.046 | 0.124 | 0.002 | 0.004 | 0.008 |
| 2013 | 0.015 | 0.040 | 0.134 | 0.017 | 0.042 | 0.106 | 0.002 | 0.003 | 0.007 |
| 2014 | 0.013 | 0.036 | 0.124 | 0.015 | 0.039 | 0.104 | 0.002 | 0.003 | 0.006 |
| 2015 | 0.012 | 0.034 | 0.113 | 0.015 | 0.037 | 0.093 | 0.001 | 0.003 | 0.006 |
| 2016 | 0.012 | 0.032 | 0.110 | 0.014 | 0.037 | 0.096 | 0.002 | 0.003 | 0.007 |
| 2017 | 0.011 | 0.030 | 0.098 | 0.012 | 0.029 | 0.073 | 0.001 | 0.002 | 0.005 |
| 2018 | 0.010 | 0.028 | 0.099 | 0.011 | 0.029 | 0.076 | 0.001 | 0.002 | 0.004 |

**TABLE 5-19: RATIO OF PREDICTED SPOTTAIL SHINER CONCENTRATIONS TO
LABORATORY-DERIVED LOEL ON A TEQ BASIS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.005 | 0.012 | 0.036 | 0.005 | 0.013 | 0.034 | 0.0003 | 0.0006 | 0.0015 |
| 1994 | 0.004 | 0.011 | 0.032 | 0.004 | 0.010 | 0.026 | 0.0002 | 0.0006 | 0.0014 |
| 1995 | 0.005 | 0.012 | 0.036 | 0.004 | 0.012 | 0.031 | 0.0003 | 0.0006 | 0.0014 |
| 1996 | 0.003 | 0.009 | 0.031 | 0.003 | 0.007 | 0.019 | 0.0002 | 0.0004 | 0.0011 |
| 1997 | 0.003 | 0.008 | 0.024 | 0.003 | 0.007 | 0.018 | 0.0002 | 0.0004 | 0.0010 |
| 1998 | 0.002 | 0.006 | 0.020 | 0.002 | 0.006 | 0.015 | 0.0002 | 0.0003 | 0.0008 |
| 1999 | 0.002 | 0.005 | 0.018 | 0.002 | 0.005 | 0.012 | 0.0001 | 0.0003 | 0.0007 |
| 2000 | 0.002 | 0.005 | 0.017 | 0.002 | 0.004 | 0.011 | 0.0002 | 0.0003 | 0.0007 |
| 2001 | 0.002 | 0.005 | 0.016 | 0.002 | 0.005 | 0.012 | 0.0001 | 0.0003 | 0.0007 |
| 2002 | 0.002 | 0.004 | 0.015 | 0.002 | 0.004 | 0.011 | 0.0001 | 0.0003 | 0.0006 |
| 2003 | 0.001 | 0.004 | 0.013 | 0.001 | 0.004 | 0.010 | 0.0001 | 0.0003 | 0.0006 |
| 2004 | 0.001 | 0.004 | 0.013 | 0.001 | 0.004 | 0.009 | 0.0001 | 0.0002 | 0.0005 |
| 2005 | 0.001 | 0.003 | 0.011 | 0.001 | 0.003 | 0.007 | 0.0001 | 0.0002 | 0.0005 |
| 2006 | 0.001 | 0.003 | 0.011 | 0.001 | 0.003 | 0.008 | 0.0001 | 0.0002 | 0.0004 |
| 2007 | 0.001 | 0.003 | 0.010 | 0.001 | 0.003 | 0.007 | 0.0001 | 0.0002 | 0.0005 |
| 2008 | 0.001 | 0.003 | 0.010 | 0.001 | 0.002 | 0.006 | 0.0001 | 0.0002 | 0.0004 |
| 2009 | 0.001 | 0.003 | 0.010 | 0.001 | 0.003 | 0.007 | 0.0001 | 0.0002 | 0.0004 |
| 2010 | 0.001 | 0.003 | 0.009 | 0.001 | 0.003 | 0.007 | 0.0001 | 0.0002 | 0.0004 |
| 2011 | 0.001 | 0.002 | 0.008 | 0.001 | 0.002 | 0.005 | 0.0001 | 0.0002 | 0.0004 |
| 2012 | 0.001 | 0.002 | 0.007 | 0.001 | 0.002 | 0.006 | 0.0001 | 0.0002 | 0.0004 |
| 2013 | 0.001 | 0.002 | 0.007 | 0.001 | 0.002 | 0.006 | 0.0001 | 0.0002 | 0.0003 |
| 2014 | 0.001 | 0.002 | 0.007 | 0.001 | 0.002 | 0.005 | 0.0001 | 0.0002 | 0.0003 |
| 2015 | 0.001 | 0.002 | 0.006 | 0.001 | 0.002 | 0.005 | 0.0001 | 0.0002 | 0.0003 |
| 2016 | 0.001 | 0.002 | 0.006 | 0.001 | 0.002 | 0.005 | 0.0001 | 0.0002 | 0.0003 |
| 2017 | 0.001 | 0.002 | 0.005 | 0.001 | 0.002 | 0.004 | 0.0001 | 0.0001 | 0.0003 |
| 2018 | 0.001 | 0.001 | 0.005 | 0.001 | 0.002 | 0.004 | 0.0001 | 0.0001 | 0.0002 |

**TABLE 5-20: RATIO OF PREDICTED BROWN BULLHEAD CONCENTRATIONS TO
LABORATORY-DERIVED NOAEL FOR TRI+ PCBS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) |
| 1993 | 6.5 | 17 | 55 | 4.6 | 10 | 22 | 2.1 | 4.6 | 10 |
| 1994 | 5.2 | 15 | 51 | 3.8 | 8.8 | 21 | 1.8 | 4.0 | 9.0 |
| 1995 | 5.4 | 15 | 49 | 4.1 | 9.3 | 21 | 1.9 | 4.3 | 9.0 |
| 1996 | 4.7 | 14 | 48 | 3.3 | 7.7 | 18 | 1.5 | 3.4 | 7.8 |
| 1997 | 4.0 | 12 | 41 | 3.1 | 7.1 | 17 | 1.4 | 3.0 | 6.6 |
| 1998 | 3.6 | 11 | 37 | 2.6 | 6.2 | 15 | 1.1 | 2.5 | 5.8 |
| 1999 | 3.1 | 9.7 | 34 | 2.3 | 5.5 | 14 | 0.9 | 2.1 | 5.0 |
| 2000 | 3.0 | 9.0 | 31 | 2.3 | 5.2 | 13 | 0.9 | 1.9 | 4.5 |
| 2001 | 2.9 | 8.5 | 30 | 2.1 | 5.1 | 13 | 0.8 | 1.9 | 4.3 |
| 2002 | 2.7 | 7.9 | 29 | 2.2 | 5.0 | 12 | 0.8 | 1.8 | 4.0 |
| 2003 | 2.5 | 7.5 | 26 | 1.9 | 4.5 | 11 | 0.7 | 1.6 | 3.8 |
| 2004 | 2.3 | 7.0 | 23 | 1.8 | 4.3 | 10 | 0.7 | 1.6 | 3.5 |
| 2005 | 2.1 | 6.4 | 21 | 1.5 | 3.8 | 9.3 | 0.6 | 1.3 | 3.0 |
| 2006 | 2.0 | 6.1 | 22 | 1.6 | 3.8 | 9.2 | 0.6 | 1.3 | 2.9 |
| 2007 | 1.9 | 5.7 | 20 | 1.4 | 3.4 | 8.3 | 0.5 | 1.1 | 2.5 |
| 2008 | 1.8 | 5.3 | 18 | 1.3 | 3.2 | 7.9 | 0.5 | 1.1 | 2.5 |
| 2009 | 1.7 | 5.1 | 17 | 1.3 | 3.1 | 7.6 | 0.5 | 1.0 | 2.4 |
| 2010 | 1.5 | 4.6 | 16 | 1.2 | 2.9 | 7.0 | 0.4 | 0.9 | 2.1 |
| 2011 | 1.4 | 4.2 | 14 | 1.1 | 2.6 | 6.4 | 0.4 | 0.9 | 1.9 |
| 2012 | 1.3 | 4.0 | 13 | 1.1 | 2.6 | 6.3 | 0.4 | 0.9 | 1.9 |
| 2013 | 1.2 | 3.7 | 13 | 1.0 | 2.4 | 5.9 | 0.3 | 0.8 | 1.8 |
| 2014 | 1.1 | 3.4 | 12 | 1.0 | 2.3 | 5.6 | 0.3 | 0.7 | 1.7 |
| 2015 | 1.1 | 3.2 | 11 | 0.9 | 2.2 | 5.4 | 0.3 | 0.7 | 1.5 |
| 2016 | 1.1 | 3.1 | 10 | 0.9 | 2.2 | 5.2 | 0.3 | 0.7 | 1.5 |
| 2017 | 0.9 | 2.8 | 10 | 0.8 | 1.9 | 4.7 | 0.3 | 0.6 | 1.3 |
| 2018 | 0.9 | 2.8 | 10 | 0.8 | 1.9 | 4.7 | 0.2 | 0.6 | 1.3 |

TABLE 5-21: RATIO OF PREDICTED BROWN BULLHEAD CONCENTRATIONS TO LABORATORY-DERIVED LOEL FOR TRI+ PCBS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 0.6 | 1.5 | 4.8 | 0.4 | 0.9 | 2.0 | 0.2 | 0.4 | 0.9 |
| 1994 | 0.5 | 1.3 | 4.5 | 0.3 | 0.8 | 1.8 | 0.2 | 0.4 | 0.8 |
| 1995 | 0.5 | 1.3 | 4.3 | 0.4 | 0.8 | 1.8 | 0.2 | 0.4 | 0.8 |
| 1996 | 0.4 | 1.2 | 4.3 | 0.3 | 0.7 | 1.6 | 0.1 | 0.3 | 0.7 |
| 1997 | 0.4 | 1.1 | 3.6 | 0.3 | 0.6 | 1.5 | 0.1 | 0.3 | 0.6 |
| 1998 | 0.3 | 1.0 | 3.3 | 0.2 | 0.5 | 1.3 | 0.1 | 0.2 | 0.5 |
| 1999 | 0.3 | 0.9 | 3.0 | 0.2 | 0.5 | 1.2 | 0.08 | 0.2 | 0.4 |
| 2000 | 0.3 | 0.8 | 2.8 | 0.2 | 0.5 | 1.1 | 0.08 | 0.2 | 0.4 |
| 2001 | 0.3 | 0.7 | 2.6 | 0.2 | 0.5 | 1.1 | 0.07 | 0.2 | 0.4 |
| 2002 | 0.2 | 0.7 | 2.5 | 0.2 | 0.4 | 1.0 | 0.07 | 0.2 | 0.4 |
| 2003 | 0.2 | 0.7 | 2.3 | 0.2 | 0.4 | 1.0 | 0.06 | 0.1 | 0.3 |
| 2004 | 0.2 | 0.6 | 2.1 | 0.2 | 0.4 | 0.9 | 0.06 | 0.1 | 0.3 |
| 2005 | 0.2 | 0.6 | 1.9 | 0.1 | 0.3 | 0.8 | 0.05 | 0.1 | 0.3 |
| 2006 | 0.2 | 0.5 | 1.9 | 0.1 | 0.3 | 0.8 | 0.05 | 0.1 | 0.3 |
| 2007 | 0.2 | 0.5 | 1.7 | 0.1 | 0.3 | 0.7 | 0.04 | 0.1 | 0.2 |
| 2008 | 0.2 | 0.5 | 1.6 | 0.1 | 0.3 | 0.7 | 0.04 | 0.09 | 0.2 |
| 2009 | 0.2 | 0.5 | 1.5 | 0.1 | 0.3 | 0.7 | 0.04 | 0.09 | 0.2 |
| 2010 | 0.1 | 0.4 | 1.4 | 0.1 | 0.3 | 0.6 | 0.04 | 0.08 | 0.2 |
| 2011 | 0.1 | 0.4 | 1.3 | 0.09 | 0.2 | 0.6 | 0.03 | 0.08 | 0.2 |
| 2012 | 0.1 | 0.4 | 1.2 | 0.10 | 0.2 | 0.6 | 0.03 | 0.08 | 0.2 |
| 2013 | 0.1 | 0.3 | 1.2 | 0.09 | 0.2 | 0.5 | 0.03 | 0.07 | 0.2 |
| 2014 | 0.1 | 0.3 | 1.0 | 0.09 | 0.2 | 0.5 | 0.03 | 0.07 | 0.1 |
| 2015 | 0.09 | 0.3 | 1.0 | 0.08 | 0.2 | 0.5 | 0.03 | 0.06 | 0.1 |
| 2016 | 0.09 | 0.3 | 0.9 | 0.08 | 0.2 | 0.5 | 0.03 | 0.06 | 0.1 |
| 2017 | 0.08 | 0.2 | 0.9 | 0.07 | 0.2 | 0.4 | 0.02 | 0.05 | 0.1 |
| 2018 | 0.08 | 0.2 | 0.8 | 0.07 | 0.2 | 0.4 | 0.02 | 0.05 | 0.1 |

TABLE 5-22: RATIO OF PREDICTED BROWN BULLHEAD CONCENTRATIONS TO LABORATORY-DERIVED NOAEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.05 | 0.19 | 0.9 | 0.04 | 0.11 | 0.39 | 0.01 | 0.03 | 0.13 |
| 1994 | 0.04 | 0.17 | 0.7 | 0.03 | 0.10 | 0.35 | 0.01 | 0.03 | 0.12 |
| 1995 | 0.05 | 0.17 | 0.8 | 0.03 | 0.10 | 0.36 | 0.01 | 0.03 | 0.12 |
| 1996 | 0.04 | 0.15 | 0.7 | 0.03 | 0.09 | 0.31 | 0.01 | 0.02 | 0.10 |
| 1997 | 0.03 | 0.14 | 0.6 | 0.03 | 0.08 | 0.28 | 0.01 | 0.02 | 0.09 |
| 1998 | 0.03 | 0.12 | 0.6 | 0.02 | 0.07 | 0.25 | 0.004 | 0.02 | 0.08 |
| 1999 | 0.03 | 0.11 | 0.5 | 0.02 | 0.06 | 0.23 | 0.004 | 0.02 | 0.07 |
| 2000 | 0.03 | 0.10 | 0.5 | 0.02 | 0.06 | 0.22 | 0.004 | 0.01 | 0.06 |
| 2001 | 0.02 | 0.10 | 0.4 | 0.02 | 0.06 | 0.21 | 0.003 | 0.01 | 0.05 |
| 2002 | 0.02 | 0.09 | 0.4 | 0.02 | 0.06 | 0.20 | 0.003 | 0.01 | 0.05 |
| 2003 | 0.02 | 0.09 | 0.4 | 0.02 | 0.05 | 0.19 | 0.003 | 0.01 | 0.05 |
| 2004 | 0.02 | 0.08 | 0.4 | 0.01 | 0.05 | 0.17 | 0.003 | 0.01 | 0.04 |
| 2005 | 0.02 | 0.07 | 0.3 | 0.01 | 0.04 | 0.16 | 0.003 | 0.01 | 0.04 |
| 2006 | 0.02 | 0.07 | 0.3 | 0.01 | 0.04 | 0.15 | 0.002 | 0.01 | 0.04 |
| 2007 | 0.02 | 0.07 | 0.3 | 0.01 | 0.04 | 0.14 | 0.002 | 0.01 | 0.03 |
| 2008 | 0.01 | 0.06 | 0.3 | 0.01 | 0.04 | 0.13 | 0.002 | 0.01 | 0.03 |
| 2009 | 0.01 | 0.06 | 0.3 | 0.01 | 0.03 | 0.13 | 0.002 | 0.01 | 0.03 |
| 2010 | 0.01 | 0.05 | 0.2 | 0.01 | 0.03 | 0.12 | 0.002 | 0.01 | 0.03 |
| 2011 | 0.01 | 0.05 | 0.2 | 0.01 | 0.03 | 0.11 | 0.002 | 0.01 | 0.03 |
| 2012 | 0.01 | 0.05 | 0.2 | 0.01 | 0.03 | 0.10 | 0.002 | 0.01 | 0.02 |
| 2013 | 0.01 | 0.04 | 0.2 | 0.01 | 0.03 | 0.10 | 0.002 | 0.01 | 0.02 |
| 2014 | 0.01 | 0.04 | 0.2 | 0.01 | 0.03 | 0.09 | 0.001 | 0.01 | 0.02 |
| 2015 | 0.01 | 0.04 | 0.2 | 0.01 | 0.02 | 0.09 | 0.001 | 0.00 | 0.02 |
| 2016 | 0.01 | 0.04 | 0.2 | 0.01 | 0.02 | 0.09 | 0.001 | 0.00 | 0.02 |
| 2017 | 0.01 | 0.03 | 0.1 | 0.01 | 0.02 | 0.08 | 0.001 | 0.00 | 0.02 |
| 2018 | 0.01 | 0.03 | 0.1 | 0.01 | 0.02 | 0.08 | 0.001 | 0.00 | 0.02 |

TABLE 5-23: RATIO OF PREDICTED BROWN BULLHEAD CONCENTRATIONS TO LABORATORY-DERIVED LOEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.024 | 0.086 | 0.381 | 0.017 | 0.049 | 0.172 | 0.004 | 0.014 | 0.060 |
| 1994 | 0.019 | 0.076 | 0.333 | 0.014 | 0.043 | 0.155 | 0.003 | 0.013 | 0.054 |
| 1995 | 0.020 | 0.077 | 0.338 | 0.016 | 0.045 | 0.161 | 0.003 | 0.012 | 0.053 |
| 1996 | 0.018 | 0.069 | 0.319 | 0.012 | 0.038 | 0.138 | 0.003 | 0.010 | 0.045 |
| 1997 | 0.015 | 0.062 | 0.263 | 0.011 | 0.035 | 0.125 | 0.002 | 0.009 | 0.039 |
| 1998 | 0.014 | 0.055 | 0.246 | 0.010 | 0.031 | 0.113 | 0.002 | 0.008 | 0.034 |
| 1999 | 0.012 | 0.050 | 0.221 | 0.008 | 0.028 | 0.101 | 0.002 | 0.007 | 0.029 |
| 2000 | 0.011 | 0.046 | 0.203 | 0.008 | 0.026 | 0.096 | 0.002 | 0.006 | 0.026 |
| 2001 | 0.010 | 0.045 | 0.194 | 0.008 | 0.025 | 0.092 | 0.001 | 0.006 | 0.024 |
| 2002 | 0.010 | 0.041 | 0.189 | 0.008 | 0.025 | 0.087 | 0.001 | 0.005 | 0.023 |
| 2003 | 0.009 | 0.038 | 0.173 | 0.007 | 0.023 | 0.083 | 0.001 | 0.005 | 0.022 |
| 2004 | 0.008 | 0.035 | 0.161 | 0.007 | 0.022 | 0.077 | 0.001 | 0.005 | 0.020 |
| 2005 | 0.008 | 0.033 | 0.146 | 0.005 | 0.019 | 0.069 | 0.001 | 0.004 | 0.018 |
| 2006 | 0.008 | 0.031 | 0.145 | 0.006 | 0.019 | 0.068 | 0.001 | 0.004 | 0.017 |
| 2007 | 0.007 | 0.029 | 0.127 | 0.005 | 0.017 | 0.062 | 0.001 | 0.004 | 0.015 |
| 2008 | 0.007 | 0.028 | 0.126 | 0.005 | 0.016 | 0.059 | 0.001 | 0.003 | 0.014 |
| 2009 | 0.006 | 0.026 | 0.117 | 0.005 | 0.015 | 0.057 | 0.001 | 0.003 | 0.013 |
| 2010 | 0.006 | 0.024 | 0.105 | 0.004 | 0.014 | 0.052 | 0.001 | 0.003 | 0.013 |
| 2011 | 0.005 | 0.021 | 0.098 | 0.004 | 0.013 | 0.048 | 0.001 | 0.003 | 0.012 |
| 2012 | 0.005 | 0.020 | 0.089 | 0.004 | 0.013 | 0.046 | 0.001 | 0.003 | 0.011 |
| 2013 | 0.005 | 0.019 | 0.085 | 0.004 | 0.012 | 0.043 | 0.001 | 0.003 | 0.010 |
| 2014 | 0.004 | 0.017 | 0.076 | 0.004 | 0.012 | 0.041 | 0.001 | 0.002 | 0.009 |
| 2015 | 0.004 | 0.017 | 0.075 | 0.003 | 0.011 | 0.039 | 0.001 | 0.002 | 0.009 |
| 2016 | 0.004 | 0.016 | 0.070 | 0.003 | 0.011 | 0.038 | 0.001 | 0.002 | 0.009 |
| 2017 | 0.003 | 0.014 | 0.063 | 0.003 | 0.009 | 0.035 | 0.001 | 0.002 | 0.008 |
| 2018 | 0.003 | 0.014 | 0.064 | 0.003 | 0.010 | 0.034 | 0.000 | 0.002 | 0.007 |

TABLE 5-24: RATIO OF OBSERVED LARGEMOUTH BASS AND BROWN BULLHEAD CONCENTRATIONS TO TOXICITY BENCHMARKS USING NYSDEC DATASET

| RATIO OF WET WEIGHT CONCENTRATION TO NOAEL | | | | | | | | | | | | | | | |
|--|------------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| | Largemouth Bass 113 | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 22 | * | 69 | 34 | 43 | 76 | 8.5 | 13 | 17 | 189 | 366 | 692 | 17 | 33 | 28 |
| 1994 | 31 | 72 | 104 | 27 | 55 | 64 | 5.7 | 13 | 18 | 90 | 136 | 193 | 18 | 23 | 69 |
| 1995 | 16 | 21 | 60 | 26 | 36 | 57 | 6.0 | 7.7 | 13 | 112 | 187 | 256 | 13 | 17 | 18 |
| 1996 | 19 | 32 | 53 | | | | | | | 56 | 75 | 114 | 11 | * | 13 |

| RATIO OF WET WEIGHT CONCENTRATION TO LOAEL | | | | | | | | | | | | | | | |
|--|------------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| | Largemouth Bass 113 | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | NA | * | NA | NA | NA | NA | 0.7 | 1.2 | 1.5 | NA | NA | NA | 1.5 | 2.9 | 2.5 |
| 1994 | NA | NA | NA | NA | NA | NA | 0.5 | 1.2 | 1.6 | NA | NA | NA | 1.5 | 2.0 | 6.1 |
| 1995 | NA | NA | NA | NA | NA | NA | 0.5 | 0.7 | 1.1 | NA | NA | NA | 1.2 | 1.5 | 1.6 |
| 1996 | NA | NA | NA | | | | | | | NA | NA | NA | 0.9 | * | 1.1 |

| RATIO OF LIPID NORMALIZED CONCENTRATIONS: TEQ BASIS TO NOAEL | | | | | | | | | | | | | | | |
|--|------------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| | Largemouth Bass 113 | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 2.4 | 5.1 | 4.2 | 9.5 | 11 | 15 | 0.2 | * | 0.3 | 42 | 64 | 94 | 0.6 | 1.1 | 1.2 |
| 1994 | 2.8 | 4.2 | 8.2 | 9.0 | 17 | 21 | 0.1 | 0.2 | 0.5 | 23 | 29 | 48 | 0.5 | 0.7 | 2.1 |
| 1995 | 2.7 | 3.3 | 4.3 | 10.5 | 12 | 18 | 0.1 | 0.1 | 0.3 | 20 | 31 | 34 | 0.2 | 0.3 | 0.4 |
| 1996 | 2.2 | 3.0 | 4.5 | | | | | | | 15 | 19 | 27 | 0.2 | * | 0.3 |

| RATIO OF LIPID NORMALIZED CONCENTRATIONS: TEQ BASIS TO LOAEL | | | | | | | | | | | | | | | |
|--|------------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|------------------------|---------|---------|-----------------------|---------|---------|
| | Largemouth Bass 113 | | | Largemouth Bass 168 | | | Brown Bullhead 168 | | | Largemouth Bass 189 | | | Brown Bullhead 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 1.2 | 2.5 | 2.0 | 4.6 | 5.4 | 7.2 | 0.1 | * | 0.1 | 20 | 31 | 45 | 0.3 | 0.5 | 0.5 |
| 1994 | 1.4 | 2.1 | 4.0 | 4.4 | 8.1 | 10.1 | 0.1 | 0.1 | 0.2 | 11 | 14 | 23 | 0.2 | 0.3 | 0.9 |
| 1995 | 1.3 | 1.6 | 2.1 | 5.1 | 6.0 | 8.9 | 0.0 | 0.1 | 0.1 | 9.8 | 15 | 17 | 0.1 | 0.1 | 0.2 |
| 1996 | 1.1 | 1.5 | 2.2 | | | | | | | 7.1 | 9.0 | 13 | 0.1 | * | 0.1 |

**TABLE 5-25: RATIO OF OBSERVED WHITE PERCH AND YELLOW PERCH CONCENTRATIONS
TO TOXICITY BENCHMARKS USING NYSDEC DATASET**

| RATIO OF WET WEIGHT CONCENTRATION TO NOAEL | | | | | | | | | | | | | | | |
|--|--------------------|------------|------------|--------------------|------------|------------|---------------------|------------|------------|---------------------|---------|------------|---------------------|------------|------------|
| | White Perch 113 | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 0.8 | * | 1.8 | 0.8 | * | 1.2 | 0.7 | * | 2.2 | 7.0 | * | 14 | 24 | 44 | 88 |
| 1994 | 0.3 | * | 0.7 | 1.6 | * | 2.8 | 0.3 | * | 0.5 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | 1.6 | * | 2.9 | 0.9 | * | 2.6 | | | | | | | | | |
| RATIO OF WET WEIGHT CONCENTRATION TO LOAEL | | | | | | | | | | | | | | | |
| | White Perch 113 | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | NA | NA | NA | NA | NA | NA | 0.1 | * | 0.2 | 0.6 | * | 1.2 | 2.1 | 3.9 | 7.7 |
| 1994 | NA | NA | NA | NA | NA | NA | 0.03 | * | 0.05 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | NA | NA | NA | NA | NA | NA | | | | | | | | | |
| RATIO OF LIPID NORMALIZED CONCENTRATIONS: TEQ BASIS TO NOAEL | | | | | | | | | | | | | | | |
| | White Perch 113 | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 2.3 | 3.3 | 4.1 | 2.0 | 2.3 | 3.0 | 2.0 | 2.5 | 8.0 | 9.0 | * | 13 | 51 | 75 | 255 |
| 1994 | 2.1 | 3.0 | 3.6 | 5.8 | 7.1 | 12 | 1.0 | 1.3 | 1.4 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | 2.8 | 7.3 | 6.5 | | | | | | | | | | | | |
| RATIO OF LIPID NORMALIZED CONCENTRATIONS: TEQ BASIS TO LOAEL | | | | | | | | | | | | | | | |
| | White Perch 113 | | | White Perch 152 | | | Yellow Perch 113 | | | Yellow Perch 168 | | | Yellow Perch 189 | | |
| | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum | Average | 95% UCL | Maximum |
| 1993 | 1.1 | 1.6 | 2.0 | 1.0 | 1.1 | 1.5 | 1.0 | 1.2 | 3.9 | 4.4 | * | 6.5 | 25 | 36 | 123 |
| 1994 | 1.0 | 1.5 | 1.8 | 2.8 | 3.4 | 5.9 | 0.5 | 0.6 | 0.7 | | | | | | |
| 1995 | | | | | | | | | | | | | | | |
| 1996 | 1.4 | 3.5 | 3.2 | | | | | | | | | | | | |

**TABLE 5-26: RATIO OF PREDICTED WHITE PERCH CONCENTRATIONS TO
FIELD-BASED NOAEL FOR TRI+ PCBS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 2.2 | 5.1 | 14.2 | 1.7 | 3.5 | 6.5 | 0.8 | 1.6 | 3.1 |
| 1994 | 1.6 | 4.1 | 13.1 | 1.4 | 2.9 | 5.6 | 0.6 | 1.4 | 2.6 |
| 1995 | 1.8 | 4.4 | 12.7 | 1.5 | 3.2 | 6.2 | 0.7 | 1.5 | 2.9 |
| 1996 | 1.4 | 3.7 | 12.8 | 1.1 | 2.4 | 4.9 | 0.5 | 1.1 | 2.2 |
| 1997 | 1.2 | 3.3 | 10.4 | 1.0 | 2.3 | 4.6 | 0.5 | 1.0 | 2.1 |
| 1998 | 1.1 | 2.8 | 9.3 | 0.9 | 2.0 | 4.0 | 0.4 | 0.9 | 1.6 |
| 1999 | 0.9 | 2.5 | 8.7 | 0.8 | 1.7 | 3.7 | 0.3 | 0.7 | 1.4 |
| 2000 | 0.9 | 2.4 | 7.8 | 0.8 | 1.7 | 3.4 | 0.3 | 0.7 | 1.3 |
| 2001 | 0.8 | 2.3 | 7.4 | 0.7 | 1.6 | 3.3 | 0.3 | 0.6 | 1.2 |
| 2002 | 0.8 | 2.1 | 7.1 | 0.7 | 1.6 | 3.2 | 0.3 | 0.6 | 1.2 |
| 2003 | 0.7 | 1.9 | 6.7 | 0.6 | 1.4 | 2.9 | 0.2 | 0.5 | 1.1 |
| 2004 | 0.7 | 1.8 | 6.0 | 0.6 | 1.4 | 2.8 | 0.3 | 0.5 | 1.1 |
| 2005 | 0.6 | 1.7 | 5.4 | 0.5 | 1.1 | 2.4 | 0.2 | 0.4 | 0.9 |
| 2006 | 0.6 | 1.6 | 5.6 | 0.5 | 1.1 | 2.5 | 0.2 | 0.4 | 0.8 |
| 2007 | 0.5 | 1.5 | 4.9 | 0.5 | 1.0 | 2.2 | 0.2 | 0.4 | 0.7 |
| 2008 | 0.5 | 1.4 | 4.6 | 0.4 | 1.0 | 2.1 | 0.2 | 0.3 | 0.7 |
| 2009 | 0.5 | 1.4 | 4.3 | 0.4 | 0.9 | 2.0 | 0.2 | 0.4 | 0.7 |
| 2010 | 0.4 | 1.2 | 4.0 | 0.4 | 0.9 | 1.9 | 0.2 | 0.3 | 0.6 |
| 2011 | 0.4 | 1.1 | 3.6 | 0.4 | 0.8 | 1.6 | 0.1 | 0.3 | 0.6 |
| 2012 | 0.4 | 1.1 | 3.4 | 0.4 | 0.8 | 1.7 | 0.1 | 0.3 | 0.6 |
| 2013 | 0.4 | 1.0 | 3.2 | 0.3 | 0.8 | 1.6 | 0.1 | 0.3 | 0.5 |
| 2014 | 0.3 | 0.9 | 2.9 | 0.3 | 0.7 | 1.5 | 0.1 | 0.3 | 0.5 |
| 2015 | 0.3 | 0.8 | 2.7 | 0.3 | 0.7 | 1.4 | 0.1 | 0.2 | 0.5 |
| 2016 | 0.3 | 0.8 | 2.7 | 0.3 | 0.7 | 1.5 | 0.1 | 0.2 | 0.5 |
| 2017 | 0.3 | 0.7 | 2.5 | 0.3 | 0.6 | 1.2 | 0.1 | 0.2 | 0.4 |
| 2018 | 0.3 | 0.7 | 2.4 | 0.3 | 0.6 | 1.3 | 0.1 | 0.2 | 0.4 |

**TABLE 5-27: RATIO OF PREDICTED YELLOW PERCH CONCENTRATIONS TO
LABORATORY-DERIVED NOAEL FOR TRI+ PCBS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 3.4 | 8.7 | 18 | 2.7 | 7.4 | 17 | 1.3 | 3.7 | 9.0 |
| 1994 | 2.5 | 6.0 | 13 | 2.3 | 5.7 | 12 | 1.1 | 2.8 | 6.3 |
| 1995 | 2.7 | 6.8 | 14 | 2.5 | 6.7 | 15 | 1.2 | 3.4 | 8.0 |
| 1996 | 2.0 | 5.0 | 13 | 1.8 | 4.4 | 9.2 | 0.8 | 2.2 | 5.5 |
| 1997 | 1.8 | 4.3 | 10 | 1.6 | 4.0 | 8.4 | 0.8 | 2.1 | 4.8 |
| 1998 | 1.6 | 3.6 | 8.8 | 1.5 | 3.8 | 7.9 | 0.7 | 1.8 | 4.3 |
| 1999 | 1.2 | 2.9 | 7.5 | 1.1 | 2.8 | 5.8 | 0.5 | 1.3 | 2.9 |
| 2000 | 1.2 | 2.8 | 6.9 | 1.1 | 2.7 | 5.2 | 0.5 | 1.2 | 2.7 |
| 2001 | 1.0 | 2.6 | 6.6 | 1.1 | 2.8 | 5.7 | 0.5 | 1.3 | 2.9 |
| 2002 | 1.1 | 2.5 | 6.3 | 1.1 | 2.7 | 5.3 | 0.5 | 1.2 | 2.7 |
| 2003 | 0.9 | 2.2 | 5.6 | 0.9 | 2.3 | 4.6 | 0.4 | 1.0 | 2.4 |
| 2004 | 0.8 | 2.1 | 5.2 | 0.9 | 2.2 | 4.4 | 0.4 | 1.1 | 2.6 |
| 2005 | 0.7 | 1.9 | 4.7 | 0.8 | 1.8 | 3.6 | 0.3 | 0.9 | 2.0 |
| 2006 | 0.8 | 1.8 | 4.8 | 0.8 | 2.0 | 4.2 | 0.3 | 1.0 | 2.3 |
| 2007 | 0.7 | 1.6 | 4.2 | 0.7 | 1.7 | 3.3 | 0.3 | 0.7 | 1.7 |
| 2008 | 0.6 | 1.5 | 4.1 | 0.6 | 1.4 | 2.7 | 0.2 | 0.6 | 1.5 |
| 2009 | 0.7 | 1.6 | 4.0 | 0.7 | 1.7 | 3.4 | 0.3 | 0.7 | 1.8 |
| 2010 | 0.6 | 1.4 | 3.6 | 0.6 | 1.5 | 3.0 | 0.3 | 0.7 | 1.6 |
| 2011 | 0.5 | 1.3 | 3.2 | 0.5 | 1.3 | 2.4 | 0.2 | 0.6 | 1.4 |
| 2012 | 0.5 | 1.2 | 3.0 | 0.6 | 1.4 | 2.8 | 0.2 | 0.6 | 1.5 |
| 2013 | 0.5 | 1.2 | 2.9 | 0.5 | 1.4 | 2.7 | 0.2 | 0.6 | 1.4 |
| 2014 | 0.4 | 1.1 | 2.8 | 0.5 | 1.3 | 2.5 | 0.2 | 0.5 | 1.2 |
| 2015 | 0.4 | 1.0 | 2.5 | 0.5 | 1.2 | 2.3 | 0.2 | 0.5 | 1.1 |
| 2016 | 0.4 | 1.0 | 2.5 | 0.5 | 1.3 | 2.8 | 0.2 | 0.6 | 1.4 |
| 2017 | 0.3 | 0.8 | 2.2 | 0.4 | 1.0 | 1.9 | 0.1 | 0.4 | 1.1 |
| 2018 | 0.3 | 0.8 | 2.2 | 0.4 | 0.8 | 1.6 | 0.1 | 0.3 | 0.8 |

TABLE 5-28: RATIO OF PREDICTED YELLOW PERCH CONCENTRATIONS TO LABORATORY-DERIVED LOEL FOR TRI+ PCBS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.3 | 0.8 | 1.6 | 0.2 | 0.6 | 1.5 | 0.1 | 0.3 | 0.8 |
| 1994 | 0.2 | 0.5 | 1.1 | 0.2 | 0.5 | 1.1 | 0.1 | 0.3 | 0.6 |
| 1995 | 0.2 | 0.6 | 1.3 | 0.2 | 0.6 | 1.3 | 0.1 | 0.3 | 0.7 |
| 1996 | 0.2 | 0.4 | 1.2 | 0.2 | 0.4 | 0.8 | 0.07 | 0.2 | 0.5 |
| 1997 | 0.2 | 0.4 | 0.9 | 0.1 | 0.4 | 0.7 | 0.07 | 0.2 | 0.4 |
| 1998 | 0.1 | 0.3 | 0.8 | 0.1 | 0.3 | 0.7 | 0.06 | 0.2 | 0.4 |
| 1999 | 0.1 | 0.3 | 0.7 | 0.1 | 0.2 | 0.5 | 0.04 | 0.1 | 0.3 |
| 2000 | 0.1 | 0.2 | 0.6 | 0.1 | 0.2 | 0.5 | 0.04 | 0.1 | 0.2 |
| 2001 | 0.09 | 0.2 | 0.6 | 0.1 | 0.2 | 0.5 | 0.04 | 0.1 | 0.3 |
| 2002 | 0.09 | 0.2 | 0.6 | 0.1 | 0.2 | 0.5 | 0.04 | 0.1 | 0.2 |
| 2003 | 0.08 | 0.2 | 0.5 | 0.08 | 0.2 | 0.4 | 0.03 | 0.1 | 0.2 |
| 2004 | 0.07 | 0.2 | 0.5 | 0.08 | 0.2 | 0.4 | 0.04 | 0.1 | 0.2 |
| 2005 | 0.07 | 0.2 | 0.4 | 0.07 | 0.2 | 0.3 | 0.03 | 0.08 | 0.2 |
| 2006 | 0.07 | 0.2 | 0.4 | 0.07 | 0.2 | 0.4 | 0.03 | 0.09 | 0.2 |
| 2007 | 0.06 | 0.1 | 0.4 | 0.06 | 0.1 | 0.3 | 0.02 | 0.06 | 0.1 |
| 2008 | 0.05 | 0.1 | 0.4 | 0.05 | 0.1 | 0.2 | 0.02 | 0.06 | 0.1 |
| 2009 | 0.06 | 0.1 | 0.4 | 0.06 | 0.1 | 0.3 | 0.02 | 0.07 | 0.2 |
| 2010 | 0.05 | 0.1 | 0.3 | 0.06 | 0.1 | 0.3 | 0.02 | 0.06 | 0.1 |
| 2011 | 0.05 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2012 | 0.04 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.06 | 0.1 |
| 2013 | 0.04 | 0.1 | 0.3 | 0.05 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2014 | 0.04 | 0.1 | 0.2 | 0.04 | 0.1 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2015 | 0.04 | 0.09 | 0.2 | 0.04 | 0.1 | 0.2 | 0.02 | 0.04 | 0.1 |
| 2016 | 0.04 | 0.09 | 0.2 | 0.05 | 0.12 | 0.2 | 0.02 | 0.05 | 0.1 |
| 2017 | 0.03 | 0.07 | 0.2 | 0.03 | 0.08 | 0.2 | 0.01 | 0.04 | 0.1 |
| 2018 | 0.03 | 0.07 | 0.2 | 0.03 | 0.07 | 0.1 | 0.01 | 0.03 | 0.07 |

**TABLE 5-29: RATIO OF PREDICTED WHITE PERCH CONCENTRATIONS TO
LABORATORY-DERIVED NOEL ON A TEQ BASIS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-----------------------|-----------------------|-------------------------------------|-----------------------|-----------------------|-------------------------------------|-----------------------|-----------------------|-------------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 0.7 | 2.5 | 9.6 | 0.5 | 1.6 | 5.2 | 0.1 | 0.46 | 1.9 |
| 1994 | 0.5 | 2.1 | 8.5 | 0.4 | 1.4 | 4.5 | 0.1 | 0.43 | 1.7 |
| 1995 | 0.6 | 2.1 | 8.4 | 0.5 | 1.5 | 4.8 | 0.11 | 0.41 | 1.7 |
| 1996 | 0.5 | 1.9 | 8.2 | 0.3 | 1.1 | 3.7 | 0.09 | 0.34 | 1.4 |
| 1997 | 0.4 | 1.6 | 6.8 | 0.3 | 1.1 | 3.6 | 0.08 | 0.31 | 1.2 |
| 1998 | 0.4 | 1.4 | 6.2 | 0.3 | 0.9 | 3.2 | 0.07 | 0.26 | 1.0 |
| 1999 | 0.3 | 1.3 | 5.6 | 0.2 | 0.8 | 2.8 | 0.06 | 0.23 | 0.9 |
| 2000 | 0.3 | 1.2 | 5.1 | 0.2 | 0.8 | 2.6 | 0.06 | 0.21 | 0.8 |
| 2001 | 0.3 | 1.1 | 4.9 | 0.2 | 0.8 | 2.5 | 0.05 | 0.19 | 0.8 |
| 2002 | 0.3 | 1.1 | 4.7 | 0.2 | 0.8 | 2.5 | 0.05 | 0.18 | 0.7 |
| 2003 | 0.2 | 1.0 | 4.2 | 0.2 | 0.7 | 2.3 | 0.05 | 0.18 | 0.7 |
| 2004 | 0.2 | 0.9 | 4.0 | 0.2 | 0.6 | 2.1 | 0.04 | 0.16 | 0.6 |
| 2005 | 0.2 | 0.8 | 3.6 | 0.2 | 0.5 | 1.8 | 0.04 | 0.14 | 0.6 |
| 2006 | 0.2 | 0.8 | 3.6 | 0.2 | 0.6 | 1.9 | 0.04 | 0.13 | 0.5 |
| 2007 | 0.2 | 0.7 | 3.4 | 0.1 | 0.5 | 1.6 | 0.04 | 0.13 | 0.5 |
| 2008 | 0.2 | 0.7 | 3.1 | 0.1 | 0.5 | 1.6 | 0.03 | 0.11 | 0.4 |
| 2009 | 0.2 | 0.7 | 2.9 | 0.1 | 0.4 | 1.6 | 0.03 | 0.11 | 0.4 |
| 2010 | 0.1 | 0.6 | 2.7 | 0.1 | 0.4 | 1.4 | 0.03 | 0.11 | 0.4 |
| 2011 | 0.1 | 0.5 | 2.4 | 0.11 | 0.4 | 1.3 | 0.03 | 0.10 | 0.4 |
| 2012 | 0.1 | 0.5 | 2.3 | 0.11 | 0.4 | 1.3 | 0.03 | 0.10 | 0.4 |
| 2013 | 0.1 | 0.5 | 2.1 | 0.11 | 0.4 | 1.2 | 0.03 | 0.09 | 0.3 |
| 2014 | 0.1 | 0.4 | 1.9 | 0.10 | 0.3 | 1.1 | 0.02 | 0.08 | 0.3 |
| 2015 | 0.10 | 0.4 | 1.8 | 0.10 | 0.3 | 1.1 | 0.02 | 0.08 | 0.3 |
| 2016 | 0.10 | 0.4 | 1.8 | 0.10 | 0.3 | 1.1 | 0.02 | 0.08 | 0.3 |
| 2017 | 0.09 | 0.4 | 1.7 | 0.08 | 0.3 | 0.9 | 0.02 | 0.06 | 0.2 |
| 2018 | 0.09 | 0.4 | 1.6 | 0.08 | 0.3 | 0.9 | 0.02 | 0.06 | 0.2 |

TABLE 5-30: RATIO OF PREDICTED WHITE PERCH CONCENTRATIONS TO LABORATORY-DERIVED LOEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.3 | 1.2 | 4.6 | 0.2 | 0.8 | 2.5 | 0.06 | 0.22 | 0.90 |
| 1994 | 0.3 | 1.0 | 4.1 | 0.2 | 0.7 | 2.2 | 0.06 | 0.21 | 0.82 |
| 1995 | 0.3 | 1.0 | 4.1 | 0.2 | 0.7 | 2.3 | 0.05 | 0.20 | 0.80 |
| 1996 | 0.2 | 0.9 | 3.9 | 0.2 | 0.6 | 1.8 | 0.04 | 0.16 | 0.67 |
| 1997 | 0.2 | 0.8 | 3.3 | 0.2 | 0.5 | 1.7 | 0.04 | 0.15 | 0.59 |
| 1998 | 0.2 | 0.7 | 3.0 | 0.1 | 0.5 | 1.5 | 0.03 | 0.12 | 0.50 |
| 1999 | 0.2 | 0.6 | 2.7 | 0.1 | 0.4 | 1.4 | 0.03 | 0.11 | 0.45 |
| 2000 | 0.1 | 0.6 | 2.5 | 0.1 | 0.4 | 1.3 | 0.03 | 0.10 | 0.39 |
| 2001 | 0.1 | 0.6 | 2.4 | 0.1 | 0.4 | 1.2 | 0.03 | 0.09 | 0.37 |
| 2002 | 0.1 | 0.5 | 2.3 | 0.1 | 0.4 | 1.2 | 0.02 | 0.09 | 0.35 |
| 2003 | 0.1 | 0.5 | 2.0 | 0.1 | 0.3 | 1.1 | 0.02 | 0.08 | 0.33 |
| 2004 | 0.1 | 0.4 | 1.9 | 0.1 | 0.3 | 1.0 | 0.02 | 0.08 | 0.30 |
| 2005 | 0.09 | 0.4 | 1.7 | 0.1 | 0.3 | 0.9 | 0.02 | 0.07 | 0.27 |
| 2006 | 0.09 | 0.4 | 1.7 | 0.1 | 0.3 | 0.9 | 0.02 | 0.06 | 0.25 |
| 2007 | 0.09 | 0.4 | 1.6 | 0.1 | 0.2 | 0.8 | 0.02 | 0.06 | 0.23 |
| 2008 | 0.08 | 0.3 | 1.5 | 0.1 | 0.2 | 0.8 | 0.01 | 0.05 | 0.21 |
| 2009 | 0.08 | 0.3 | 1.4 | 0.1 | 0.2 | 0.8 | 0.01 | 0.05 | 0.20 |
| 2010 | 0.07 | 0.3 | 1.3 | 0.1 | 0.2 | 0.7 | 0.02 | 0.05 | 0.20 |
| 2011 | 0.06 | 0.3 | 1.1 | 0.1 | 0.2 | 0.6 | 0.01 | 0.05 | 0.18 |
| 2012 | 0.06 | 0.3 | 1.1 | 0.1 | 0.2 | 0.6 | 0.01 | 0.05 | 0.17 |
| 2013 | 0.06 | 0.2 | 1.0 | 0.1 | 0.2 | 0.6 | 0.01 | 0.04 | 0.16 |
| 2014 | 0.05 | 0.2 | 0.9 | 0.1 | 0.2 | 0.6 | 0.01 | 0.04 | 0.15 |
| 2015 | 0.05 | 0.2 | 0.9 | 0.0 | 0.2 | 0.5 | 0.01 | 0.04 | 0.14 |
| 2016 | 0.05 | 0.2 | 0.9 | 0.0 | 0.2 | 0.5 | 0.01 | 0.04 | 0.13 |
| 2017 | 0.04 | 0.2 | 0.8 | 0.0 | 0.1 | 0.4 | 0.01 | 0.03 | 0.12 |
| 2018 | 0.04 | 0.2 | 0.8 | 0.0 | 0.1 | 0.5 | 0.01 | 0.03 | 0.11 |

TABLE 5-31: RATIO OF PREDICTED YELLOW PERCH CONCENTRATIONS TO LABORATORY-DERIVED NOEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 1.0 | 2.7 | 7.5 | 0.8 | 2.4 | 6.5 | 0.4 | 1.4 | 4.8 |
| 1994 | 0.7 | 1.9 | 5.2 | 0.6 | 1.8 | 4.8 | 0.4 | 1.3 | 4.5 |
| 1995 | 0.8 | 2.1 | 5.6 | 0.7 | 2.1 | 5.9 | 0.4 | 1.3 | 4.3 |
| 1996 | 0.6 | 1.6 | 5.1 | 0.5 | 1.4 | 3.6 | 0.3 | 1.1 | 3.5 |
| 1997 | 0.5 | 1.4 | 3.9 | 0.5 | 1.3 | 3.3 | 0.3 | 0.9 | 3.1 |
| 1998 | 0.4 | 1.2 | 3.3 | 0.4 | 1.2 | 3.1 | 0.3 | 0.8 | 2.7 |
| 1999 | 0.3 | 0.9 | 2.9 | 0.3 | 0.9 | 2.3 | 0.2 | 0.7 | 2.3 |
| 2000 | 0.3 | 0.9 | 2.7 | 0.3 | 0.8 | 2.1 | 0.2 | 0.7 | 2.1 |
| 2001 | 0.3 | 0.8 | 2.4 | 0.3 | 0.9 | 2.3 | 0.2 | 0.6 | 2.0 |
| 2002 | 0.3 | 0.8 | 2.4 | 0.3 | 0.8 | 2.1 | 0.2 | 0.6 | 1.9 |
| 2003 | 0.3 | 0.7 | 2.1 | 0.3 | 0.7 | 1.8 | 0.2 | 0.5 | 1.8 |
| 2004 | 0.2 | 0.7 | 2.1 | 0.3 | 0.7 | 1.8 | 0.2 | 0.5 | 1.6 |
| 2005 | 0.2 | 0.6 | 1.8 | 0.2 | 0.6 | 1.5 | 0.1 | 0.4 | 1.5 |
| 2006 | 0.2 | 0.6 | 1.8 | 0.2 | 0.6 | 1.7 | 0.1 | 0.4 | 1.4 |
| 2007 | 0.2 | 0.5 | 1.6 | 0.2 | 0.5 | 1.3 | 0.1 | 0.4 | 1.3 |
| 2008 | 0.2 | 0.5 | 1.5 | 0.2 | 0.4 | 1.1 | 0.1 | 0.3 | 1.1 |
| 2009 | 0.2 | 0.5 | 1.5 | 0.2 | 0.5 | 1.3 | 0.1 | 0.3 | 1.1 |
| 2010 | 0.2 | 0.5 | 1.3 | 0.2 | 0.5 | 1.2 | 0.1 | 0.3 | 1.1 |
| 2011 | 0.1 | 0.4 | 1.2 | 0.2 | 0.4 | 1.0 | 0.1 | 0.3 | 1.0 |
| 2012 | 0.1 | 0.4 | 1.2 | 0.2 | 0.4 | 1.1 | 0.1 | 0.3 | 1.0 |
| 2013 | 0.1 | 0.4 | 1.1 | 0.2 | 0.4 | 1.1 | 0.1 | 0.3 | 0.9 |
| 2014 | 0.1 | 0.3 | 1.0 | 0.1 | 0.4 | 1.0 | 0.1 | 0.3 | 0.8 |
| 2015 | 0.1 | 0.3 | 0.9 | 0.1 | 0.4 | 0.9 | 0.1 | 0.3 | 0.8 |
| 2016 | 0.1 | 0.3 | 0.9 | 0.1 | 0.4 | 1.1 | 0.1 | 0.3 | 0.7 |
| 2017 | 0.1 | 0.3 | 0.8 | 0.1 | 0.3 | 0.8 | 0.1 | 0.2 | 0.6 |
| 2018 | 0.09 | 0.3 | 0.8 | 0.1 | 0.3 | 0.6 | 0.1 | 0.2 | 0.6 |

**TABLE 5-32: RATIO OF PREDICTED YELLOW PERCH CONCENTRATIONS TO
LABORATORY-DERIVED LOEL ON A TEQ BASIS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.5 | 1.3 | 3.6 | 0.4 | 1.1 | 3.2 | 0.2 | 0.7 | 2.3 |
| 1994 | 0.3 | 0.9 | 2.5 | 0.3 | 0.9 | 2.3 | 0.2 | 0.6 | 2.2 |
| 1995 | 0.4 | 1.0 | 2.7 | 0.4 | 1.0 | 2.9 | 0.2 | 0.6 | 2.1 |
| 1996 | 0.3 | 0.8 | 2.5 | 0.2 | 0.7 | 1.8 | 0.2 | 0.5 | 1.7 |
| 1997 | 0.2 | 0.7 | 1.9 | 0.2 | 0.6 | 1.6 | 0.1 | 0.5 | 1.5 |
| 1998 | 0.2 | 0.6 | 1.6 | 0.2 | 0.6 | 1.5 | 0.1 | 0.4 | 1.3 |
| 1999 | 0.2 | 0.4 | 1.4 | 0.2 | 0.4 | 1.1 | 0.1 | 0.3 | 1.1 |
| 2000 | 0.1 | 0.4 | 1.3 | 0.1 | 0.4 | 1.0 | 0.1 | 0.3 | 1.0 |
| 2001 | 0.1 | 0.4 | 1.2 | 0.2 | 0.4 | 1.1 | 0.1 | 0.3 | 1.0 |
| 2002 | 0.1 | 0.4 | 1.1 | 0.2 | 0.4 | 1.0 | 0.09 | 0.3 | 0.9 |
| 2003 | 0.1 | 0.3 | 1.0 | 0.1 | 0.4 | 0.9 | 0.09 | 0.3 | 0.9 |
| 2004 | 0.1 | 0.3 | 1.0 | 0.1 | 0.3 | 0.9 | 0.08 | 0.2 | 0.8 |
| 2005 | 0.1 | 0.3 | 0.9 | 0.1 | 0.3 | 0.7 | 0.07 | 0.2 | 0.7 |
| 2006 | 0.1 | 0.3 | 0.9 | 0.1 | 0.3 | 0.8 | 0.07 | 0.2 | 0.7 |
| 2007 | 0.09 | 0.3 | 0.8 | 0.09 | 0.3 | 0.6 | 0.07 | 0.2 | 0.6 |
| 2008 | 0.08 | 0.2 | 0.7 | 0.08 | 0.2 | 0.5 | 0.06 | 0.2 | 0.6 |
| 2009 | 0.09 | 0.2 | 0.7 | 0.09 | 0.3 | 0.6 | 0.06 | 0.2 | 0.6 |
| 2010 | 0.08 | 0.2 | 0.7 | 0.09 | 0.2 | 0.6 | 0.06 | 0.2 | 0.5 |
| 2011 | 0.07 | 0.2 | 0.6 | 0.07 | 0.2 | 0.5 | 0.06 | 0.2 | 0.5 |
| 2012 | 0.07 | 0.2 | 0.6 | 0.08 | 0.2 | 0.5 | 0.06 | 0.2 | 0.5 |
| 2013 | 0.07 | 0.2 | 0.5 | 0.08 | 0.2 | 0.5 | 0.05 | 0.1 | 0.4 |
| 2014 | 0.06 | 0.2 | 0.5 | 0.07 | 0.2 | 0.5 | 0.05 | 0.1 | 0.4 |
| 2015 | 0.06 | 0.2 | 0.5 | 0.07 | 0.2 | 0.4 | 0.05 | 0.1 | 0.4 |
| 2016 | 0.06 | 0.2 | 0.5 | 0.07 | 0.2 | 0.5 | 0.05 | 0.1 | 0.4 |
| 2017 | 0.05 | 0.1 | 0.4 | 0.05 | 0.15 | 0.4 | 0.04 | 0.1 | 0.3 |
| 2018 | 0.04 | 0.1 | 0.4 | 0.05 | 0.13 | 0.3 | 0.03 | 0.09 | 0.3 |

**TABLE 5-33: RATIO OF PREDICTED LARGEMOUTH BASS CONCENTRATIONS TO
FIELD-BASED NOAEL FOR TRI+ PCBS**

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|--------------------|--------------------|-------------------------------|
| | 25th | Median | 95th | 25th | Median | 95th | 25th | Median | 95th |
| | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) | (mg/kg wet weight) | (mg/kg wet weight) | Percentile (mg/kg wet weight) |
| 1993 | 42 | 88 | 163 | 34 | 68 | 118 | 16 | 33 | 60 |
| 1994 | 24 | 50 | 117 | 24 | 47 | 80 | 11 | 23 | 41 |
| 1995 | 26 | 56 | 129 | 27 | 52 | 89 | 13 | 26 | 44 |
| 1996 | 22 | 48 | 132 | 20 | 42 | 71 | 10 | 22 | 41 |
| 1997 | 16 | 37 | 98 | 19 | 37 | 62 | 9 | 19 | 35 |
| 1998 | 14 | 31 | 84 | 16 | 33 | 56 | 8.0 | 16 | 29 |
| 1999 | 11 | 26 | 75 | 13 | 26 | 47 | 5.7 | 12 | 23 |
| 2000 | 11 | 25 | 70 | 12 | 25 | 44 | 5.7 | 12 | 21 |
| 2001 | 10 | 23 | 64 | 12 | 24 | 43 | 5.4 | 11 | 20 |
| 2002 | 10 | 23 | 64 | 13 | 26 | 44 | 5.5 | 12 | 21 |
| 2003 | 8.6 | 20 | 57 | 10 | 22 | 38 | 4.7 | 10 | 18 |
| 2004 | 8.1 | 19 | 52 | 10 | 21 | 36 | 4.9 | 10 | 19 |
| 2005 | 7.0 | 16 | 48 | 8.0 | 16 | 28 | 3.7 | 7.8 | 14 |
| 2006 | 7.3 | 16 | 48 | 8.8 | 18 | 32 | 4.2 | 8.4 | 15 |
| 2007 | 6.4 | 15 | 44 | 7.5 | 15 | 26 | 3.2 | 6.5 | 12 |
| 2008 | 5.8 | 14 | 41 | 6.6 | 13 | 24 | 2.9 | 6.1 | 12 |
| 2009 | 6.2 | 14 | 39 | 7.2 | 14 | 26 | 3.2 | 6.6 | 13 |
| 2010 | 5.7 | 13 | 36 | 6.4 | 13 | 23 | 2.8 | 5.6 | 10 |
| 2011 | 4.5 | 11 | 32 | 5.5 | 11 | 20 | 2.5 | 5.3 | 10 |
| 2012 | 4.6 | 11 | 30 | 5.5 | 11 | 20 | 2.5 | 5.2 | 10 |
| 2013 | 4.5 | 10 | 29 | 5.6 | 12 | 20 | 2.4 | 5.0 | 9.6 |
| 2014 | 4.2 | 9.7 | 26 | 5.5 | 11 | 20 | 2.3 | 4.7 | 8.5 |
| 2015 | 3.8 | 8.8 | 24 | 4.9 | 10 | 17 | 2.0 | 4.3 | 7.8 |
| 2016 | 4.0 | 9.1 | 25 | 5.5 | 11 | 19 | 2.3 | 4.7 | 8.4 |
| 2017 | 3.2 | 7.5 | 22 | 4.1 | 8.3 | 15 | 1.6 | 3.6 | 7.5 |
| 2018 | 3.3 | 7.4 | 22 | 3.9 | 8.0 | 14 | 1.6 | 3.4 | 6.7 |

TABLE 5-34: RATIO OF PREDICTED LARGEMOUTH BASS CONCENTRATIONS TO LABORATORY-DERIVED NOAEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|----------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 1.8 | 6.0 | 25.2 | 1.5 | 4.6 | 19.5 | 0.4 | 1.1 | 3.3 |
| 1994 | 1.1 | 3.6 | 16.5 | 1.0 | 3.2 | 13.6 | 0.4 | 1.0 | 3.1 |
| 1995 | 1.2 | 4.0 | 17.9 | 1.2 | 3.6 | 14.8 | 0.4 | 1.0 | 3.1 |
| 1996 | 1.0 | 3.4 | 18.7 | 0.9 | 2.8 | 11.9 | 0.3 | 0.8 | 2.5 |
| 1997 | 0.8 | 2.6 | 13.3 | 0.8 | 2.5 | 10.3 | 0.3 | 0.7 | 2.2 |
| 1998 | 0.6 | 2.3 | 11.7 | 0.7 | 2.2 | 9.3 | 0.2 | 0.6 | 1.9 |
| 1999 | 0.5 | 1.9 | 9.8 | 0.6 | 1.8 | 7.5 | 0.2 | 0.5 | 1.7 |
| 2000 | 0.5 | 1.8 | 8.9 | 0.5 | 1.7 | 7.2 | 0.2 | 0.5 | 1.6 |
| 2001 | 0.5 | 1.7 | 9.0 | 0.5 | 1.7 | 6.9 | 0.2 | 0.5 | 1.5 |
| 2002 | 0.5 | 1.6 | 8.3 | 0.6 | 1.8 | 7.2 | 0.2 | 0.4 | 1.3 |
| 2003 | 0.4 | 1.4 | 7.5 | 0.5 | 1.5 | 6.0 | 0.2 | 0.4 | 1.4 |
| 2004 | 0.4 | 1.3 | 7.0 | 0.5 | 1.4 | 5.9 | 0.2 | 0.4 | 1.2 |
| 2005 | 0.3 | 1.2 | 6.3 | 0.4 | 1.1 | 4.6 | 0.1 | 0.4 | 1.1 |
| 2006 | 0.3 | 1.2 | 6.1 | 0.4 | 1.2 | 5.1 | 0.1 | 0.3 | 1.0 |
| 2007 | 0.3 | 1.1 | 5.5 | 0.3 | 1.0 | 4.3 | 0.1 | 0.3 | 1.0 |
| 2008 | 0.3 | 1.0 | 5.4 | 0.3 | 0.9 | 3.8 | 0.1 | 0.3 | 0.9 |
| 2009 | 0.3 | 1.0 | 5.3 | 0.3 | 1.0 | 4.0 | 0.1 | 0.3 | 0.9 |
| 2010 | 0.3 | 0.9 | 4.7 | 0.3 | 0.9 | 3.8 | 0.1 | 0.3 | 0.9 |
| 2011 | 0.2 | 0.8 | 4.1 | 0.2 | 0.8 | 3.3 | 0.1 | 0.3 | 0.9 |
| 2012 | 0.2 | 0.8 | 4.0 | 0.2 | 0.8 | 3.3 | 0.1 | 0.3 | 0.9 |
| 2013 | 0.2 | 0.7 | 3.8 | 0.3 | 0.8 | 3.3 | 0.1 | 0.3 | 0.8 |
| 2014 | 0.2 | 0.7 | 3.5 | 0.2 | 0.8 | 3.1 | 0.1 | 0.3 | 0.7 |
| 2015 | 0.2 | 0.6 | 3.4 | 0.2 | 0.7 | 2.8 | 0.1 | 0.2 | 0.7 |
| 2016 | 0.2 | 0.7 | 3.2 | 0.2 | 0.7 | 3.1 | 0.1 | 0.3 | 0.7 |
| 2017 | 0.2 | 0.5 | 2.8 | 0.2 | 0.6 | 2.5 | 0.1 | 0.2 | 0.6 |
| 2018 | 0.2 | 0.5 | 2.7 | 0.2 | 0.5 | 2.3 | 0.1 | 0.2 | 0.5 |

TABLE 5-35: RATIO OF PREDICTED LARGEMOUTH BASS CONCENTRATIONS TO LABORATORY-DERIVED LOEL ON A TEQ BASIS

| Year | Thompson Island Pool | | | River Mile 168 | | | River Mile 154 | | |
|------|-------------------------------|---------------------------------|---|----------------------------|---------------------------------|---|-------------------------------|---------------------------------|---|
| | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) | 25th (mg/kg wet weight) | Median (mg/kg wet weight) | 95th Percentile (mg/kg wet weight) |
| 1993 | 0.9 | 2.9 | 12.2 | 0.7 | 2.2 | 9.4 | 0.2 | 0.5 | 1.6 |
| 1994 | 0.5 | 1.7 | 8.0 | 0.5 | 1.6 | 6.6 | 0.2 | 0.5 | 1.5 |
| 1995 | 0.6 | 1.9 | 8.7 | 0.6 | 1.7 | 7.1 | 0.2 | 0.5 | 1.5 |
| 1996 | 0.5 | 1.6 | 9.0 | 0.4 | 1.4 | 5.7 | 0.14 | 0.4 | 1.2 |
| 1997 | 0.4 | 1.3 | 6.4 | 0.4 | 1.2 | 5.0 | 0.14 | 0.3 | 1.1 |
| 1998 | 0.3 | 1.1 | 5.7 | 0.3 | 1.1 | 4.5 | 0.11 | 0.3 | 0.9 |
| 1999 | 0.3 | 0.9 | 4.8 | 0.3 | 0.9 | 3.6 | 0.10 | 0.2 | 0.8 |
| 2000 | 0.2 | 0.9 | 4.3 | 0.3 | 0.8 | 3.5 | 0.10 | 0.3 | 0.8 |
| 2001 | 0.2 | 0.8 | 4.3 | 0.3 | 0.8 | 3.4 | 0.09 | 0.2 | 0.7 |
| 2002 | 0.2 | 0.8 | 4.0 | 0.3 | 0.9 | 3.5 | 0.08 | 0.2 | 0.7 |
| 2003 | 0.2 | 0.7 | 3.6 | 0.2 | 0.7 | 2.9 | 0.09 | 0.2 | 0.7 |
| 2004 | 0.2 | 0.6 | 3.4 | 0.2 | 0.7 | 2.8 | 0.08 | 0.2 | 0.6 |
| 2005 | 0.2 | 0.6 | 3.0 | 0.2 | 0.5 | 2.2 | 0.07 | 0.2 | 0.5 |
| 2006 | 0.2 | 0.6 | 2.9 | 0.2 | 0.6 | 2.5 | 0.07 | 0.2 | 0.5 |
| 2007 | 0.15 | 0.5 | 2.6 | 0.2 | 0.5 | 2.1 | 0.07 | 0.2 | 0.5 |
| 2008 | 0.14 | 0.5 | 2.6 | 0.1 | 0.4 | 1.9 | 0.06 | 0.14 | 0.4 |
| 2009 | 0.14 | 0.5 | 2.6 | 0.2 | 0.5 | 1.9 | 0.05 | 0.13 | 0.4 |
| 2010 | 0.13 | 0.4 | 2.3 | 0.14 | 0.4 | 1.8 | 0.06 | 0.15 | 0.4 |
| 2011 | 0.10 | 0.4 | 2.0 | 0.12 | 0.4 | 1.6 | 0.06 | 0.14 | 0.4 |
| 2012 | 0.10 | 0.4 | 1.9 | 0.12 | 0.4 | 1.6 | 0.06 | 0.15 | 0.4 |
| 2013 | 0.10 | 0.4 | 1.8 | 0.12 | 0.4 | 1.6 | 0.06 | 0.14 | 0.4 |
| 2014 | 0.09 | 0.3 | 1.7 | 0.12 | 0.4 | 1.5 | 0.05 | 0.12 | 0.4 |
| 2015 | 0.09 | 0.3 | 1.6 | 0.10 | 0.3 | 1.3 | 0.05 | 0.12 | 0.3 |
| 2016 | 0.09 | 0.3 | 1.5 | 0.12 | 0.4 | 1.5 | 0.05 | 0.13 | 0.4 |
| 2017 | 0.08 | 0.3 | 1.4 | 0.09 | 0.3 | 1.2 | 0.04 | 0.10 | 0.3 |
| 2018 | 0.07 | 0.3 | 1.3 | 0.09 | 0.3 | 1.1 | 0.04 | 0.08 | 0.2 |

**TABLE 5-36
COMPARISON OF MEASURED STRIPED BASS CONCENTRATIONS
TO TOXICITY REFERENCE VALUES**

| River Mile | Year | <<--- Tri+ in Tissue --->> | | << -- TEQ in Eggs: Lipid Normalized -->> | | | |
|--------------|------|----------------------------|------------|--|------------|------------|-------------|
| | | Field-Based NOAEL | | LOAEL | | NOAEL | |
| | | Average | 95%UCL | Average | 95% UCL | Average | 95% UCL |
| GW Bridge | 1993 | 0.4 | 0.7 | 0.2 | 0.4 | 0.5 | 0.7 |
| Nyack | 1993 | 0.8 | 1.3 | 0.4 | 0.6 | 0.9 | 1.3 |
| Croton | 1993 | 1.4 | 2.7 | 0.5 | 0.9 | 1.0 | 1.8 |
| Stoney Point | 1993 | 0.5 | 0.7 | 0.3 | 0.4 | 0.6 | 0.8 |
| Cornwall | | | | | | | |
| Poughkeepsie | 1993 | 1.0 | 1.6 | 0.7 | 1.2 | 1.4 | 2.5 |
| Catskill | 1993 | 1.2 | 1.8 | 1.1 | 1.9 | 2.3 | 3.9 |
| Troy | 1993 | 4.0 | 5.7 | 2.8 | 4.9 | 5.8 | 10.0 |
| GW Bridge | 1994 | 0.0 | | | | | |
| Nyack | 1994 | 0.5 | 0.8 | 0.3 | 0.5 | 0.7 | 1.0 |
| Croton | 1994 | 0.6 | 1.0 | 0.4 | 0.8 | 0.9 | 1.6 |
| Stoney Point | 1994 | 0.6 | 0.9 | 0.4 | 0.5 | 0.8 | 1.1 |
| Cornwall | | | | | | | |
| Poughkeepsie | 1994 | 0.8 | 1.4 | 0.5 | 0.7 | 0.9 | 1.4 |
| Catskill | 1994 | 1.0 | 2.7 | 1.2 | 4.3 | 2.5 | 9.0 |
| Troy | 1994 | 2.1 | 3.2 | 2.3 | 3.1 | 4.8 | 6.4 |
| GW Bridge | 1995 | 0.0 | | | | | |
| Nyack | 1995 | 0.6 | 0.9 | 0.5 | 0.8 | 1.0 | 1.7 |
| Croton | 1995 | 0.4 | 0.5 | 0.2 | 0.3 | 0.5 | 0.7 |
| Stoney Point | 1995 | | | | | | |
| Cornwall | 1995 | 0.7 | 0.9 | 1.5 | 2.3 | 3.1 | 4.7 |
| Poughkeepsie | 1995 | 0.6 | 0.7 | 0.3 | 0.3 | 0.6 | 0.7 |
| Catskill | 1995 | 0.5 | 0.9 | 0.3 | 0.5 | 0.7 | 1.1 |
| Troy | 1995 | 2.1 | 2.7 | 1.6 | 2.1 | 3.3 | 4.4 |
| GW Bridge | 1996 | 0.0 | | | | | |
| Nyack | 1996 | 0.4 | 0.6 | 0.3 | 0.5 | 0.7 | 1.0 |
| Croton | 1996 | 0.6 | 0.9 | 0.4 | 0.5 | 0.8 | 1.1 |
| Stoney Point | 1996 | 0.5 | 0.7 | 0.4 | 0.7 | 0.9 | 1.3 |
| Cornwall | | | | | | | |
| Poughkeepsie | 1996 | 0.6 | 0.8 | 0.4 | 0.6 | 0.9 | 1.3 |
| Catskill | 1996 | 0.6 | 1.0 | 0.8 | 1.7 | 1.6 | 3.5 |
| Troy | 1996 | 1.6 | 3.6 | 2.0 | 4.8 | 4.1 | 10.0 |

Bold values indicate exceedances

**TABLE 5-37: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE TREE SWALLOW FOR TRI+ CONGENERS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|-------------|-------------|------------|---|----------|-------------|------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% UCL | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% |
| | ADD | ADD | ADD | UCL ADD | Conc. | UCL Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | NA | NA | 0.7 | 1.2 | NA | NA | 1.0 | 1.7 |
| Stillwater (168) | NA | NA | 1.2 | 5.5 | NA | NA | 1.7 | 7.8 |
| Federal Dam (154) | NA | NA | 0.3 | 0.5 | NA | NA | 0.4 | 0.7 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | NA | NA | 0.04 | 0.1 | NA | NA | 0.1 | 0.1 |
| 137.2 | NA | NA | 0.08 | 0.3 | NA | NA | 0.1 | 0.4 |
| 122.4 | NA | NA | 0.04 | 0.1 | NA | NA | 0.1 | 0.2 |
| 113.8 | NA | NA | 0.04 | 0.2 | NA | NA | 0.1 | 0.2 |
| 100 | NA | NA | 0.02 | 0.1 | NA | NA | 0.03 | 0.2 |
| 88.9 | NA | NA | 0.01 | 0.02 | NA | NA | 0.01 | 0.03 |
| 58.7 | NA | NA | 0.03 | 0.3 | NA | NA | 0.04 | 0.4 |
| 47.3 | NA | NA | 0.04 | 0.3 | NA | NA | 0.05 | 0.4 |
| 25.8 | NA | NA | 0.01 | 0.0 | NA | NA | 0.01 | 0.03 |

Bold value indicates exceedances

TABLE 5-38: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE TREE SWALLOWS BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|---------|---------|------------|---------|---------|---------|---------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | NA | NA | 0.8 | 1.2 | NA | NA | 0.4 | 0.7 | NA | NA | 0.2 | 0.2 |
| 1994 | NA | NA | 0.7 | 1.1 | NA | NA | 0.4 | 0.7 | NA | NA | 0.2 | 0.2 |
| 1995 | NA | NA | 0.7 | 1.0 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.2 |
| 1996 | NA | NA | 0.6 | 0.9 | NA | NA | 0.3 | 0.6 | NA | NA | 0.1 | 0.2 |
| 1997 | NA | NA | 0.6 | 0.8 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 1998 | NA | NA | 0.5 | 0.8 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 1999 | NA | NA | 0.5 | 0.7 | NA | NA | 0.3 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2000 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2001 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2002 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2003 | NA | NA | 0.4 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2004 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2005 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2006 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2007 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.3 | NA | NA | 0.05 | 0.1 |
| 2008 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.2 | NA | NA | 0.04 | 0.05 |
| 2009 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.05 |
| 2010 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2011 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2012 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2013 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.04 |
| 2014 | NA | NA | 0.2 | 0.2 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.03 |
| 2015 | NA | NA | 0.2 | 0.2 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.03 |
| 2016 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.03 |
| 2017 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.03 |
| 2018 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.03 |

Bold value indicates exceedances

TABLE 5-39 : RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS FOR FEMALE TREE SWALLOWS BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|---------|------------|------------|---------|---------|---------|------------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | NA | NA | 1.1 | 1.6 | NA | NA | 0.6 | 1.0 | NA | NA | 0.3 | 0.3 |
| 1994 | NA | NA | 1.0 | 1.5 | NA | NA | 0.6 | 0.9 | NA | NA | 0.3 | 0.3 |
| 1995 | NA | NA | 1.0 | 1.4 | NA | NA | 0.6 | 0.9 | NA | NA | 0.2 | 0.3 |
| 1996 | NA | NA | 0.9 | 1.3 | NA | NA | 0.5 | 0.8 | NA | NA | 0.2 | 0.2 |
| 1997 | NA | NA | 0.8 | 1.2 | NA | NA | 0.4 | 0.7 | NA | NA | 0.2 | 0.2 |
| 1998 | NA | NA | 0.7 | 1.1 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.2 |
| 1999 | NA | NA | 0.7 | 1.0 | NA | NA | 0.4 | 0.6 | NA | NA | 0.1 | 0.1 |
| 2000 | NA | NA | 0.6 | 0.9 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 2001 | NA | NA | 0.6 | 0.8 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 2002 | NA | NA | 0.5 | 0.8 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 2003 | NA | NA | 0.5 | 0.7 | NA | NA | 0.3 | 0.5 | NA | NA | 0.1 | 0.1 |
| 2004 | NA | NA | 0.5 | 0.7 | NA | NA | 0.3 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2005 | NA | NA | 0.4 | 0.6 | NA | NA | 0.3 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2006 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2007 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 2008 | NA | NA | 0.4 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2009 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2010 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2011 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.1 |
| 2012 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.3 | NA | NA | 0.05 | 0.1 |
| 2013 | NA | NA | 0.2 | 0.4 | NA | NA | 0.2 | 0.3 | NA | NA | 0.05 | 0.1 |
| 2014 | NA | NA | 0.2 | 0.3 | NA | NA | 0.2 | 0.2 | NA | NA | 0.04 | 0.05 |
| 2015 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2016 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2017 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.04 |
| 2018 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.04 |

Bold value indicates exceedances

**TABLE 5-40: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE TREE SWALLOW ON TEQ BASIS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|----------|-------------|------------|---|----------|-------------|------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% |
| | ADD | UCL ADD | ADD | UCL ADD | Conc. | UCL Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | NA | NA | 0.1 | 0.2 | NA | NA | 0.4 | 0.6 |
| Stillwater (168) | NA | NA | 0.2 | 1.0 | NA | NA | 0.6 | 3.0 |
| Federal Dam (154) | NA | NA | 0.05 | 0.08 | NA | NA | 0.2 | 0.3 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | NA | NA | 0.01 | 0.02 | NA | NA | 0.02 | 0.05 |
| 137.2 | NA | NA | 0.01 | 0.06 | NA | NA | 0.04 | 0.17 |
| 122.4 | NA | NA | 0.01 | 0.02 | NA | NA | 0.02 | 0.06 |
| 113.8 | NA | NA | 0.01 | 0.03 | NA | NA | 0.02 | 0.09 |
| 100 | NA | NA | 0.00 | 0.02 | NA | NA | 0.01 | 0.07 |
| 88.9 | NA | NA | 0.00 | 0.00 | NA | NA | 0.01 | 0.01 |
| 58.7 | NA | NA | 0.01 | 0.05 | NA | NA | 0.02 | 0.15 |
| 47.3 | NA | NA | 0.01 | 0.05 | NA | NA | 0.02 | 0.14 |
| 25.8 | NA | NA | 0.00 | 0.00 | NA | NA | 0.01 | 0.01 |

**TABLE 5-41: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR
FEMALE TREE SWALLOW USING TEQ FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL |
| 1993 | NA | NA | 0.138 | 0.202 | NA | NA | 0.077 | 0.121 | NA | NA | 0.034 | 0.036 |
| 1994 | NA | NA | 0.130 | 0.192 | NA | NA | 0.073 | 0.116 | NA | NA | 0.032 | 0.034 |
| 1995 | NA | NA | 0.121 | 0.175 | NA | NA | 0.069 | 0.108 | NA | NA | 0.029 | 0.031 |
| 1996 | NA | NA | 0.109 | 0.160 | NA | NA | 0.061 | 0.097 | NA | NA | 0.025 | 0.027 |
| 1997 | NA | NA | 0.100 | 0.146 | NA | NA | 0.056 | 0.089 | NA | NA | 0.021 | 0.023 |
| 1998 | NA | NA | 0.092 | 0.132 | NA | NA | 0.050 | 0.080 | NA | NA | 0.019 | 0.020 |
| 1999 | NA | NA | 0.083 | 0.123 | NA | NA | 0.046 | 0.073 | NA | NA | 0.016 | 0.018 |
| 2000 | NA | NA | 0.077 | 0.110 | NA | NA | 0.043 | 0.068 | NA | NA | 0.015 | 0.016 |
| 2001 | NA | NA | 0.072 | 0.105 | NA | NA | 0.042 | 0.067 | NA | NA | 0.014 | 0.015 |
| 2002 | NA | NA | 0.068 | 0.100 | NA | NA | 0.040 | 0.064 | NA | NA | 0.013 | 0.014 |
| 2003 | NA | NA | 0.063 | 0.092 | NA | NA | 0.038 | 0.060 | NA | NA | 0.012 | 0.013 |
| 2004 | NA | NA | 0.059 | 0.085 | NA | NA | 0.035 | 0.055 | NA | NA | 0.011 | 0.012 |
| 2005 | NA | NA | 0.054 | 0.079 | NA | NA | 0.032 | 0.051 | NA | NA | 0.010 | 0.011 |
| 2006 | NA | NA | 0.052 | 0.076 | NA | NA | 0.031 | 0.049 | NA | NA | 0.009 | 0.010 |
| 2007 | NA | NA | 0.048 | 0.070 | NA | NA | 0.029 | 0.045 | NA | NA | 0.009 | 0.009 |
| 2008 | NA | NA | 0.046 | 0.066 | NA | NA | 0.027 | 0.043 | NA | NA | 0.008 | 0.009 |
| 2009 | NA | NA | 0.043 | 0.062 | NA | NA | 0.026 | 0.041 | NA | NA | 0.008 | 0.008 |
| 2010 | NA | NA | 0.039 | 0.058 | NA | NA | 0.024 | 0.038 | NA | NA | 0.007 | 0.008 |
| 2011 | NA | NA | 0.036 | 0.052 | NA | NA | 0.022 | 0.035 | NA | NA | 0.007 | 0.007 |
| 2012 | NA | NA | 0.033 | 0.048 | NA | NA | 0.021 | 0.033 | NA | NA | 0.006 | 0.007 |
| 2013 | NA | NA | 0.031 | 0.046 | NA | NA | 0.020 | 0.032 | NA | NA | 0.006 | 0.006 |
| 2014 | NA | NA | 0.029 | 0.042 | NA | NA | 0.019 | 0.030 | NA | NA | 0.005 | 0.006 |
| 2015 | NA | NA | 0.027 | 0.040 | NA | NA | 0.018 | 0.029 | NA | NA | 0.005 | 0.006 |
| 2016 | NA | NA | 0.025 | 0.037 | NA | NA | 0.017 | 0.027 | NA | NA | 0.005 | 0.005 |
| 2017 | NA | NA | 0.024 | 0.035 | NA | NA | 0.016 | 0.026 | NA | NA | 0.004 | 0.005 |
| 2018 | NA | NA | 0.023 | 0.035 | NA | NA | 0.016 | 0.025 | NA | NA | 0.004 | 0.004 |

**TABLE 5-42: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS
BASED ON FISHRAND FOR FEMALE TREE SWALLOW USING TEQ FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL |
| 1993 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 1994 | NA | NA | 0.4 | 0.6 | NA | NA | 0.2 | 0.4 | NA | NA | 0.1 | 0.1 |
| 1995 | NA | NA | 0.4 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.09 | 0.1 |
| 1996 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.08 | 0.08 |
| 1997 | NA | NA | 0.3 | 0.5 | NA | NA | 0.2 | 0.3 | NA | NA | 0.07 | 0.07 |
| 1998 | NA | NA | 0.3 | 0.4 | NA | NA | 0.2 | 0.2 | NA | NA | 0.06 | 0.06 |
| 1999 | NA | NA | 0.3 | 0.4 | NA | NA | 0.1 | 0.2 | NA | NA | 0.05 | 0.05 |
| 2000 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.05 | 0.05 |
| 2001 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.05 |
| 2002 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2003 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.04 | 0.04 |
| 2004 | NA | NA | 0.2 | 0.3 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.04 |
| 2005 | NA | NA | 0.2 | 0.2 | NA | NA | 0.1 | 0.2 | NA | NA | 0.03 | 0.03 |
| 2006 | NA | NA | 0.2 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.03 | 0.03 |
| 2007 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.03 | 0.03 |
| 2008 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.03 |
| 2009 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.03 |
| 2010 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2011 | NA | NA | 0.1 | 0.2 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2012 | NA | NA | 0.1 | 0.1 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2013 | NA | NA | 0.095 | 0.1 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2014 | NA | NA | 0.088 | 0.1 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2015 | NA | NA | 0.083 | 0.1 | NA | NA | 0.1 | 0.1 | NA | NA | 0.02 | 0.02 |
| 2016 | NA | NA | 0.078 | 0.1 | NA | NA | 0.1 | 0.1 | NA | NA | 0.01 | 0.02 |
| 2017 | NA | NA | 0.072 | 0.1 | NA | NA | 0.05 | 0.1 | NA | NA | 0.01 | 0.01 |
| 2018 | NA | NA | 0.072 | 0.1 | NA | NA | 0.05 | 0.1 | NA | NA | 0.01 | 0.01 |

**TABLE 5-43: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE MALLARD FOR TRI+ CONGENERS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|---|---|---|---|---|---|---|---|
| | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient | LOAEL vs. Average Conc. Hazard Quotient | LOAEL vs. 95% UCL Conc. Hazard Quotient | NOAEL vs. Average Conc. Hazard Quotient | NOAEL vs. 95% UCL Conc. Hazard Quotient |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 0.8 | 1.4 | 8.1 | 14 | 19 | 30 | 129 | 202 |
| Stillwater (168) | 1.5 | 2.7 | 15 | 27 | 36 | 62 | 240 | 417 |
| Federal Dam (154) | 0.4 | 0.8 | 4.2 | 7.6 | 8.5 | 15 | 57 | 99 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 0.1 | 0.2 | 1.0 | 2.3 | 1.2 | 2.1 | 8 | 14 |
| 137.2 | 0.1 | 0.3 | 1.4 | 3.1 | 2.3 | 4.1 | 16 | 27 |
| 122.4 | 0.1 | 0.6 | 0.7 | 5.6 | 1.1 | 2.7 | 7.3 | 18 |
| 113.8 | 0.1 | 0.5 | 0.7 | 5.1 | 0.9 | 1.6 | 6.3 | 11 |
| 100 | 0.0 | 0.6 | 0.5 | 5.9 | 0.5 | 3.5 | 3.5 | 24 |
| 88.9 | 0.0 | 0.4 | 0.3 | 4.3 | 0.3 | 0.5 | 1.7 | 3.1 |
| 58.7 | 0.0 | 0.5 | 0.5 | 4.6 | 0.7 | 1.2 | 4.5 | 7.8 |
| 47.3 | 0.1 | 0.7 | 0.6 | 6.8 | 0.9 | 6.6 | 6.1 | 44 |
| 25.8 | 0.0 | 0.4 | 0.3 | 4.3 | 0.3 | 0.5 | 1.8 | 3.0 |

Bold values indicate exceedances

TABLE 5-44: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS FOR FEMALE MALLARD BASED ON FISHRAND RESULTS FOR THE TRI+ CONGENERS

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|------------|------------|------------|---------|---------|------------|------------|---------|---------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 0.9 | 1.3 | 9.1 | 13 | 0.6 | 0.8 | 5.7 | 8.2 | 0.2 | 0.2 | 2.1 | 2.3 |
| 1994 | 0.9 | 1.2 | 8.7 | 12 | 0.6 | 0.8 | 5.6 | 8.1 | 0.2 | 0.2 | 2.0 | 2.2 |
| 1995 | 0.8 | 1.1 | 8.4 | 11 | 0.5 | 0.8 | 5.3 | 7.6 | 0.2 | 0.2 | 2.0 | 2.2 |
| 1996 | 0.7 | 1.0 | 6.8 | 9.7 | 0.4 | 0.6 | 4.1 | 6.2 | 0.2 | 0.2 | 1.5 | 1.7 |
| 1997 | 0.6 | 0.9 | 6.4 | 9.0 | 0.4 | 0.6 | 4.2 | 6.1 | 0.1 | 0.2 | 1.4 | 1.5 |
| 1998 | 0.5 | 0.8 | 5.5 | 7.7 | 0.3 | 0.5 | 3.2 | 4.9 | 0.1 | 0.1 | 1.2 | 1.3 |
| 1999 | 0.5 | 0.7 | 5.0 | 7.3 | 0.3 | 0.5 | 3.1 | 4.6 | 0.1 | 0.1 | 1.1 | 1.2 |
| 2000 | 0.5 | 0.7 | 4.6 | 6.5 | 0.3 | 0.4 | 2.8 | 4.2 | 0.1 | 0.1 | 1.0 | 1.0 |
| 2001 | 0.4 | 0.6 | 4.4 | 6.3 | 0.3 | 0.4 | 3.0 | 4.4 | 0.1 | 0.1 | 0.9 | 1.0 |
| 2002 | 0.4 | 0.6 | 4.1 | 6.0 | 0.3 | 0.4 | 2.7 | 4.1 | 0.1 | 0.1 | 0.8 | 0.9 |
| 2003 | 0.4 | 0.5 | 3.8 | 5.5 | 0.3 | 0.4 | 2.6 | 3.8 | 0.1 | 0.1 | 0.8 | 0.9 |
| 2004 | 0.4 | 0.5 | 3.5 | 5.0 | 0.2 | 0.4 | 2.5 | 3.6 | 0.1 | 0.1 | 0.7 | 0.8 |
| 2005 | 0.3 | 0.5 | 3.3 | 4.7 | 0.2 | 0.3 | 2.1 | 3.2 | 0.1 | 0.1 | 0.7 | 0.7 |
| 2006 | 0.3 | 0.5 | 3.2 | 4.5 | 0.2 | 0.3 | 2.1 | 3.2 | 0.1 | 0.1 | 0.6 | 0.7 |
| 2007 | 0.3 | 0.4 | 2.9 | 4.2 | 0.2 | 0.3 | 1.9 | 2.9 | 0.1 | 0.1 | 0.6 | 0.6 |
| 2008 | 0.3 | 0.4 | 2.8 | 3.9 | 0.2 | 0.3 | 1.7 | 2.6 | 0.05 | 0.05 | 0.5 | 0.5 |
| 2009 | 0.3 | 0.4 | 2.7 | 3.8 | 0.2 | 0.3 | 1.8 | 2.6 | 0.05 | 0.06 | 0.5 | 0.6 |
| 2010 | 0.2 | 0.3 | 2.4 | 3.5 | 0.2 | 0.2 | 1.6 | 2.4 | 0.05 | 0.06 | 0.5 | 0.6 |
| 2011 | 0.2 | 0.3 | 2.1 | 3.1 | 0.1 | 0.2 | 1.4 | 2.2 | 0.05 | 0.05 | 0.5 | 0.5 |
| 2012 | 0.2 | 0.3 | 2.1 | 2.9 | 0.1 | 0.2 | 1.5 | 2.2 | 0.05 | 0.05 | 0.5 | 0.5 |
| 2013 | 0.2 | 0.3 | 1.9 | 2.7 | 0.1 | 0.2 | 1.4 | 2.0 | 0.04 | 0.05 | 0.4 | 0.5 |
| 2014 | 0.2 | 0.3 | 1.8 | 2.5 | 0.1 | 0.2 | 1.3 | 1.9 | 0.04 | 0.04 | 0.4 | 0.4 |
| 2015 | 0.2 | 0.2 | 1.6 | 2.4 | 0.1 | 0.2 | 1.2 | 1.8 | 0.04 | 0.04 | 0.4 | 0.4 |
| 2016 | 0.2 | 0.2 | 1.6 | 2.2 | 0.1 | 0.2 | 1.2 | 1.8 | 0.04 | 0.04 | 0.4 | 0.4 |
| 2017 | 0.1 | 0.2 | 1.4 | 2.1 | 0.1 | 0.2 | 1.0 | 1.6 | 0.03 | 0.03 | 0.3 | 0.3 |
| 2018 | 0.1 | 0.2 | 1.5 | 2.1 | 0.1 | 0.2 | 1.1 | 1.7 | 0.04 | 0.04 | 0.4 | 0.4 |

Bold values indicate exceedances

TABLE 5-45: RATIO OF EGG CONCENTRATIONS TO BENCHMARKS FOR FEMALE MALLARD BASED ON FISHRAND RESULTS FOR THE TRI+ CONGENERs

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|------------|-----------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 20 | 29 | 135 | 197 | 11 | 18 | 75 | 118 | 4.9 | 5.3 | 33 | 35 |
| 1994 | 19 | 28 | 126 | 187 | 11 | 17 | 71 | 113 | 4.6 | 5.0 | 31 | 33 |
| 1995 | 18 | 26 | 118 | 171 | 10 | 16 | 67 | 106 | 4.2 | 4.6 | 28 | 31 |
| 1996 | 16 | 23 | 107 | 156 | 8.9 | 14 | 60 | 94 | 3.6 | 3.9 | 24 | 26 |
| 1997 | 15 | 21 | 98 | 143 | 8.1 | 13 | 54 | 86 | 3.1 | 3.4 | 21 | 22 |
| 1998 | 13 | 19 | 90 | 128 | 7.3 | 12 | 49 | 78 | 2.7 | 2.9 | 18 | 20 |
| 1999 | 12 | 18 | 81 | 120 | 6.7 | 11 | 45 | 71 | 2.4 | 2.6 | 16 | 17 |
| 2000 | 11 | 16 | 75 | 107 | 6.3 | 10 | 42 | 67 | 2.1 | 2.3 | 14 | 15 |
| 2001 | 11 | 15 | 71 | 103 | 6.1 | 10 | 41 | 65 | 2.0 | 2.2 | 13 | 15 |
| 2002 | 10 | 15 | 66 | 98 | 5.9 | 9.3 | 39 | 62 | 1.9 | 2.1 | 13 | 14 |
| 2003 | 9.2 | 13 | 62 | 90 | 5.5 | 8.8 | 37 | 59 | 1.8 | 1.9 | 12 | 13 |
| 2004 | 8.5 | 12 | 57 | 82 | 5.1 | 8.0 | 34 | 54 | 1.6 | 1.7 | 11 | 12 |
| 2005 | 7.9 | 12 | 53 | 77 | 4.7 | 7.5 | 32 | 50 | 1.5 | 1.6 | 10 | 11 |
| 2006 | 7.6 | 11 | 51 | 74 | 4.5 | 7.1 | 30 | 47 | 1.4 | 1.5 | 9.1 | 10 |
| 2007 | 7.1 | 10 | 47 | 68 | 4.2 | 6.6 | 28 | 44 | 1.2 | 1.3 | 8.3 | 8.9 |
| 2008 | 6.7 | 10 | 45 | 64 | 3.9 | 6.2 | 26 | 42 | 1.1 | 1.2 | 7.7 | 8.3 |
| 2009 | 6.3 | 9.1 | 42 | 61 | 3.7 | 5.9 | 25 | 40 | 1.1 | 1.2 | 7.4 | 8.0 |
| 2010 | 5.7 | 8.4 | 38 | 56 | 3.5 | 5.6 | 23 | 37 | 1.0 | 1.1 | 7.0 | 7.6 |
| 2011 | 5.2 | 7.5 | 35 | 50 | 3.2 | 5.1 | 22 | 34 | 1.0 | 1.0 | 6.4 | 6.9 |
| 2012 | 4.8 | 7.0 | 32 | 47 | 3.1 | 4.9 | 21 | 33 | 0.9 | 1.0 | 6.1 | 6.5 |
| 2013 | 4.5 | 6.7 | 30 | 45 | 2.9 | 4.6 | 20 | 31 | 0.8 | 0.9 | 5.7 | 6.1 |
| 2014 | 4.2 | 6.2 | 28 | 41 | 2.7 | 4.4 | 18 | 29 | 0.8 | 0.8 | 5.2 | 5.6 |
| 2015 | 3.9 | 5.8 | 26 | 39 | 2.6 | 4.2 | 18 | 28 | 0.7 | 0.8 | 5.0 | 5.4 |
| 2016 | 3.7 | 5.4 | 25 | 36 | 2.5 | 4.0 | 17 | 27 | 0.7 | 0.7 | 4.6 | 4.9 |
| 2017 | 3.4 | 5.1 | 23 | 34 | 2.4 | 3.7 | 16 | 25 | 0.6 | 0.7 | 4.1 | 4.4 |
| 2018 | 3.4 | 5.1 | 23 | 34 | 2.3 | 3.6 | 15 | 24 | 0.6 | 0.6 | 4.0 | 4.3 |

Bold values indicate exceedances

**TABLE 5-47: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS
FOR FEMALE MALLARD FOR PERIOD 1993 - 2018 ON A TEQ BASIS**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|-------------|-------------|-----------|-----------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 182 | 189 | 1818 | 1893 | 196 | 203 | 1964 | 2026 | 36 | 37 | 360 | 369 |
| 1994 | 204 | 212 | 2039 | 2116 | 223 | 230 | 2231 | 2298 | 37 | 38 | 373 | 382 |
| 1995 | 217 | 225 | 2174 | 2251 | 210 | 216 | 2102 | 2164 | 58 | 60 | 584 | 597 |
| 1996 | 92 | 97 | 925 | 972 | 99 | 103 | 994 | 1032 | 21 | 21 | 209 | 214 |
| 1997 | 98 | 102 | 977 | 1023 | 158 | 163 | 1584 | 1633 | 33 | 34 | 329 | 337 |
| 1998 | 32 | 35 | 322 | 351 | 57 | 59 | 567 | 594 | 23 | 23 | 229 | 235 |
| 1999 | 42 | 45 | 419 | 451 | 73 | 75 | 725 | 754 | 23 | 24 | 234 | 240 |
| 2000 | 32 | 35 | 322 | 348 | 53 | 55 | 525 | 549 | 23 | 23 | 225 | 231 |
| 2001 | 44 | 46 | 436 | 464 | 91 | 94 | 909 | 941 | 17 | 17 | 170 | 174 |
| 2002 | 37 | 40 | 371 | 397 | 66 | 69 | 664 | 690 | 13 | 13 | 125 | 129 |
| 2003 | 31 | 34 | 313 | 336 | 62 | 64 | 617 | 641 | 21 | 21 | 209 | 214 |
| 2004 | 28 | 30 | 282 | 303 | 71 | 73 | 710 | 735 | 17 | 18 | 173 | 177 |
| 2005 | 27 | 29 | 267 | 287 | 40 | 42 | 398 | 416 | 19 | 19 | 187 | 191 |
| 2006 | 32 | 34 | 324 | 344 | 57 | 59 | 570 | 591 | 14 | 14 | 135 | 138 |
| 2007 | 24 | 26 | 242 | 259 | 40 | 42 | 404 | 421 | 19 | 19 | 186 | 191 |
| 2008 | 20 | 22 | 204 | 220 | 29 | 30 | 287 | 301 | 10 | 10 | 97 | 99 |
| 2009 | 32 | 34 | 318 | 336 | 49 | 50 | 487 | 505 | 12 | 13 | 122 | 126 |
| 2010 | 23 | 25 | 235 | 250 | 38 | 40 | 381 | 396 | 16 | 17 | 165 | 168 |
| 2011 | 16 | 17 | 157 | 170 | 22 | 24 | 225 | 236 | 12 | 12 | 120 | 123 |
| 2012 | 25 | 26 | 246 | 260 | 40 | 41 | 397 | 412 | 20 | 21 | 204 | 209 |
| 2013 | 16 | 18 | 165 | 177 | 31 | 32 | 309 | 322 | 17 | 17 | 168 | 172 |
| 2014 | 18 | 19 | 176 | 187 | 29 | 30 | 287 | 298 | 18 | 19 | 183 | 187 |
| 2015 | 15 | 16 | 146 | 156 | 25 | 26 | 248 | 259 | 13 | 13 | 131 | 135 |
| 2016 | 19 | 20 | 193 | 203 | 31 | 33 | 314 | 326 | 18 | 19 | 183 | 187 |
| 2017 | 11 | 12 | 107 | 116 | 15 | 16 | 147 | 155 | 9 | 10 | 93 | 96 |
| 2018 | 19 | 20 | 187 | 197 | 37 | 38 | 366 | 378 | 22 | 23 | 221 | 226 |

Bold values indicate exceedances

TABLE 5-48: RATIO OF MODELED EGG CONCENTRATION TO BENCHMARKS FOR FEMALE MALLARD FOR PERIOD 1993 - 2018 ON A TEQ BASIS

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|-------------|-------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 1106 | 1618 | 4422 | 6474 | 612 | 966 | 2449 | 3865 | 269 | 289 | 1077 | 1157 |
| 1994 | 1038 | 1535 | 4150 | 6140 | 583 | 926 | 2331 | 3702 | 254 | 275 | 1016 | 1098 |
| 1995 | 970 | 1403 | 3880 | 5611 | 548 | 868 | 2194 | 3472 | 233 | 251 | 932 | 1003 |
| 1996 | 874 | 1280 | 3497 | 5120 | 489 | 773 | 1958 | 3091 | 197 | 214 | 789 | 856 |
| 1997 | 804 | 1171 | 3215 | 4684 | 447 | 709 | 1786 | 2836 | 172 | 184 | 687 | 737 |
| 1998 | 737 | 1054 | 2948 | 4215 | 403 | 638 | 1611 | 2553 | 150 | 161 | 599 | 643 |
| 1999 | 666 | 987 | 2665 | 3949 | 366 | 583 | 1466 | 2332 | 131 | 141 | 522 | 564 |
| 2000 | 613 | 882 | 2451 | 3527 | 346 | 546 | 1382 | 2186 | 117 | 126 | 469 | 504 |
| 2001 | 579 | 842 | 2314 | 3368 | 337 | 536 | 1349 | 2142 | 111 | 119 | 442 | 477 |
| 2002 | 541 | 802 | 2163 | 3208 | 322 | 512 | 1287 | 2048 | 105 | 113 | 419 | 454 |
| 2003 | 505 | 738 | 2019 | 2950 | 304 | 483 | 1216 | 1930 | 98 | 106 | 392 | 423 |
| 2004 | 469 | 677 | 1876 | 2707 | 279 | 441 | 1116 | 1764 | 89 | 95 | 355 | 381 |
| 2005 | 433 | 635 | 1732 | 2539 | 258 | 410 | 1034 | 1639 | 81 | 87 | 323 | 348 |
| 2006 | 416 | 606 | 1665 | 2424 | 245 | 389 | 979 | 1556 | 75 | 81 | 299 | 323 |
| 2007 | 388 | 560 | 1551 | 2242 | 230 | 364 | 918 | 1456 | 68 | 73 | 272 | 293 |
| 2008 | 369 | 529 | 1476 | 2114 | 215 | 341 | 859 | 1363 | 63 | 68 | 252 | 273 |
| 2009 | 348 | 498 | 1392 | 1994 | 206 | 326 | 824 | 1306 | 61 | 66 | 244 | 263 |
| 2010 | 315 | 463 | 1260 | 1850 | 192 | 306 | 768 | 1223 | 57 | 62 | 229 | 248 |
| 2011 | 286 | 414 | 1142 | 1655 | 177 | 283 | 709 | 1131 | 52 | 57 | 210 | 226 |
| 2012 | 266 | 386 | 1063 | 1546 | 169 | 268 | 676 | 1071 | 50 | 54 | 199 | 215 |
| 2013 | 246 | 367 | 984 | 1470 | 161 | 255 | 643 | 1019 | 47 | 50 | 186 | 201 |
| 2014 | 230 | 338 | 918 | 1353 | 151 | 240 | 603 | 960 | 43 | 46 | 171 | 184 |
| 2015 | 215 | 320 | 858 | 1279 | 146 | 231 | 582 | 926 | 41 | 44 | 163 | 176 |
| 2016 | 204 | 294 | 814 | 1176 | 138 | 220 | 553 | 878 | 38 | 41 | 150 | 162 |
| 2017 | 188 | 279 | 753 | 1117 | 129 | 205 | 518 | 821 | 33 | 36 | 134 | 145 |
| 2018 | 187 | 278 | 746 | 1112 | 126 | 200 | 505 | 801 | 33 | 35 | 130 | 140 |

Bold values indicate exceedances

**TABLE 5-49: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE BELTED KINGFISHER FOR TRI+ CONGENERS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|------------|-------------|-------------|---|------------|-------------|-------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% |
| | ADD | UCL ADD | ADD | UCL ADD | Conc. | UCL Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 107 | 213 | 748 | 1494 | 257 | 513 | 1720 | 3439 |
| Stillwater (168) | 59 | 173 | 410 | 1209 | 140 | 414 | 936 | 2771 |
| Federal Dam (154) | 14 | 22 | 96 | 151 | 33 | 52 | 221 | 347 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 9.3 | 12 | 65 | 86 | 22 | 30 | 150 | 198 |
| 137.2 | 19 | 44 | 131 | 309 | 45 | 106 | 302 | 710 |
| 122.4 | 7.5 | 13 | 52 | 91 | 18 | 31 | 120 | 209 |
| 113.8 | 7.8 | 11 | 55 | 77 | 19 | 26 | 126 | 177 |
| 100 | 3.4 | 8.4 | 24 | 59 | 8.2 | 20 | 55 | 133 |
| 88.9 | 6.1 | 8.5 | 43 | 59 | 15 | 20 | 98 | 136 |
| 58.7 | 7.1 | 14 | 50 | 97 | 17 | 33 | 115 | 223 |
| 47.3 | 6.5 | 14 | 46 | 95 | 16 | 33 | 105 | 218 |
| 25.8 | 4.5 | 5.6 | 32 | 39 | 11 | 13 | 73 | 89 |

Bold values indicate exceedances

**TABLE 5-50: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE GREAT BLUE HERON FOR TRI+ CONGENERS**

| Location | <<<< ---- Dietary Dose ---- >>>> | | | | <<<< ---- Egg Concentration ---- >>>> | | | |
|----------------------------|----------------------------------|-------------|-------------|-------------|---------------------------------------|------------|-------------|-------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% | vs. Average | vs. 95% |
| | ADD | ADD | ADD | ADD | Conc. | UCL Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 47 | 95 | 327 | 667 | 284 | 580 | 1902 | 3883 |
| Stillwater (168) | 17 | 25 | 116 | 178 | 96 | 137 | 642 | 918 |
| Federal Dam (154) | 3.8 | 5.6 | 27 | 39 | 22 | 33 | 151 | 219 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 4.3 | 5.2 | 30 | 36 | 26 | 31 | 175 | 210 |
| 137.2 | 8.7 | 19 | 61 | 132 | 53 | 115 | 354 | 768 |
| 122.4 | 3.3 | 5.4 | 23 | 38 | 20 | 33 | 135 | 219 |
| 113.8 | 3.5 | 3.7 | 24 | 26 | 21 | 22 | 142 | 147 |
| 100 | 1.5 | 2.8 | 11 | 19 | 9 | 16 | 61 | 106 |
| 88.9 | 3.0 | 4.1 | 21 | 29 | 18 | 25 | 122 | 168 |
| 58.7 | 3.3 | 3.8 | 23 | 27 | 20 | 23 | 133 | 151 |
| 47.3 | 2.9 | 4.0 | 20 | 28 | 18 | 23 | 119 | 157 |
| 25.8 | 2.2 | 2.6 | 15 | 18 | 13 | 16 | 89 | 107 |

Bold values indicate exceedances

TABLE 5-51: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS BASED ON 1993 DATA FOR FEMALE EAGLE FOR TRI+ CONGENERS

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|-------------|-------------|-------------|---|----------|-------------|-------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% | vs. Average | vs. 95% |
| | ADD | ADD | ADD | ADD | Conc. | UCL Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 172 | 333 | 1204 | 2331 | NA | NA | 882 | 1707 |
| Stillwater (168) | 31 | 39 | 214 | 276 | NA | NA | 157 | 202 |
| Federal Dam (154) | 22 | 40 | 155 | 279 | NA | NA | 114 | 204 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 22 | 40 | 155 | 279 | NA | NA | 114 | 204 |
| 137.2 | 84 | 199 | 586 | 1394 | NA | NA | 429 | 1021 |
| 122.4 | 19 | 27 | 136 | 188 | NA | NA | 100 | 137 |
| 113.8 | 18 | 25 | 123 | 173 | NA | NA | 90 | 126 |
| 100 | 20 | 63 | 142 | 438 | NA | NA | 104 | 321 |
| 88.9 | 13 | 25 | 91 | 174 | NA | NA | 67 | 127 |
| 58.7 | 15 | 22 | 106 | 157 | NA | NA | 78 | 115 |
| 47.3 | 17 | 47 | 121 | 327 | NA | NA | 89 | 239 |
| 25.8 | 12 | 24 | 86 | 170 | NA | NA | 63 | 125 |

Bold values indicate exceedances

TABLE 5-52: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BELTED KINGFISHER BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018

| Year | LOAEL Average | LOAEL 95% UCL | NOAEL Average | NOAEL 95% UCL | LOAEL Average | LOAEL 95% UCL | NOAEL Average | NOAEL 95% UCL | LOAEL Average | LOAEL 95% UCL | NOAEL Average | NOAEL 95% UCL |
|------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|
| 1993 | 113 | 192 | 792 | 1346 | 56 | 75 | 392 | 524 | 28 | 48 | 193 | 339 |
| 1994 | 62 | 100 | 433 | 699 | 47 | 59 | 326 | 416 | 22 | 33 | 151 | 231 |
| 1995 | 57 | 90 | 401 | 631 | 50 | 67 | 352 | 471 | 25 | 42 | 173 | 297 |
| 1996 | 57 | 94 | 402 | 655 | 35 | 45 | 245 | 315 | 17 | 29 | 116 | 203 |
| 1997 | 45 | 79 | 312 | 555 | 31 | 40 | 217 | 277 | 15 | 24 | 107 | 167 |
| 1998 | 39 | 64 | 272 | 450 | 28 | 36 | 194 | 254 | 12 | 20 | 87 | 141 |
| 1999 | 34 | 59 | 236 | 414 | 23 | 29 | 158 | 204 | 9.4 | 14 | 66 | 99 |
| 2000 | 28 | 52 | 198 | 367 | 21 | 27 | 147 | 190 | 8.3 | 13 | 58 | 93 |
| 2001 | 27 | 49 | 192 | 342 | 22 | 28 | 154 | 199 | 9.4 | 15 | 66 | 107 |
| 2002 | 26 | 47 | 179 | 326 | 21 | 27 | 148 | 190 | 8.6 | 14 | 60 | 95 |
| 2003 | 24 | 44 | 171 | 307 | 18 | 23 | 127 | 164 | 7.7 | 13 | 54 | 89 |
| 2004 | 22 | 40 | 152 | 283 | 17 | 22 | 121 | 154 | 7.5 | 13 | 53 | 89 |
| 2005 | 21 | 38 | 145 | 265 | 15 | 19 | 102 | 132 | 6.2 | 10 | 43 | 70 |
| 2006 | 18 | 34 | 129 | 237 | 15 | 19 | 105 | 135 | 6.1 | 10 | 42 | 71 |
| 2007 | 18 | 33 | 127 | 232 | 14 | 18 | 96 | 124 | 5.1 | 8.6 | 36 | 60 |
| 2008 | 17 | 30 | 116 | 208 | 12 | 16 | 85 | 110 | 4.6 | 7.4 | 32 | 52 |
| 2009 | 16 | 29 | 111 | 205 | 13 | 17 | 90 | 117 | 5.3 | 9.3 | 37 | 65 |
| 2010 | 15 | 28 | 107 | 197 | 13 | 16 | 89 | 114 | 4.8 | 8.1 | 34 | 56 |
| 2011 | 14 | 25 | 100 | 174 | 10 | 13 | 70 | 92 | 4.0 | 6.9 | 28 | 48 |
| 2012 | 13 | 23 | 88 | 159 | 11 | 15 | 80 | 105 | 4.6 | 7.6 | 32 | 53 |
| 2013 | 12 | 22 | 84 | 151 | 11 | 14 | 75 | 96 | 3.9 | 6.5 | 28 | 46 |
| 2014 | 11 | 20 | 80 | 143 | 10 | 13 | 68 | 88 | 3.6 | 6.1 | 26 | 43 |
| 2015 | 10 | 18 | 73 | 129 | 9.4 | 12 | 66 | 84 | 3.4 | 5.3 | 24 | 37 |
| 2016 | 10 | 17 | 69 | 121 | 9.2 | 12 | 65 | 86 | 3.9 | 8.0 | 27 | 56 |
| 2017 | 9.3 | 17 | 65 | 122 | 7.6 | 10 | 53 | 69 | 2.8 | 5.2 | 20 | 37 |
| 2018 | 8.8 | 16 | 61 | 113 | 7.5 | 10 | 52 | 68 | 2.6 | 4.3 | 18 | 30 |

Bold values indicate exceedances

**TABLE 5-53: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BLUE HERON
BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL |
| 1993 | 48 | 84 | 337 | 589 | 23 | 30 | 164 | 211 | 12 | 16 | 83 | 110 |
| 1994 | 23 | 38 | 158 | 266 | 19 | 23 | 132 | 158 | 8.8 | 11 | 62 | 77 |
| 1995 | 21 | 34 | 146 | 239 | 21 | 27 | 147 | 190 | 11 | 14 | 74 | 97 |
| 1996 | 22 | 37 | 152 | 258 | 14 | 17 | 96 | 116 | 6.8 | 9.1 | 48 | 64 |
| 1997 | 16 | 31 | 111 | 214 | 12 | 14 | 84 | 100 | 6.3 | 7.9 | 44 | 56 |
| 1998 | 13 | 24 | 94 | 167 | 11 | 13 | 75 | 92 | 5.1 | 6.5 | 36 | 46 |
| 1999 | 11 | 22 | 80 | 153 | 8.4 | 10 | 59 | 70 | 3.7 | 4.6 | 26 | 32 |
| 2000 | 9.2 | 19 | 64 | 135 | 7.8 | 9.3 | 54 | 65 | 3.3 | 4.2 | 23 | 29 |
| 2001 | 9.1 | 18 | 63 | 125 | 8.4 | 10 | 59 | 70 | 3.8 | 5.0 | 27 | 35 |
| 2002 | 8.4 | 17 | 59 | 119 | 8.1 | 10 | 57 | 67 | 3.5 | 4.4 | 25 | 31 |
| 2003 | 8.1 | 16 | 57 | 113 | 6.7 | 7.9 | 47 | 55 | 3.1 | 4.0 | 22 | 28 |
| 2004 | 7.0 | 15 | 49 | 104 | 6.5 | 7.6 | 45 | 53 | 3.1 | 4.0 | 22 | 28 |
| 2005 | 6.9 | 14 | 48 | 98 | 5.2 | 6.2 | 37 | 43 | 2.5 | 3.2 | 17 | 22 |
| 2006 | 5.8 | 12 | 41 | 85 | 5.6 | 6.6 | 39 | 46 | 2.5 | 3.2 | 17 | 23 |
| 2007 | 5.9 | 12 | 41 | 85 | 5.0 | 6.0 | 35 | 42 | 2.0 | 2.7 | 14 | 19 |
| 2008 | 5.3 | 11 | 37 | 75 | 4.4 | 5.2 | 31 | 36 | 1.8 | 2.3 | 13 | 16 |
| 2009 | 5.1 | 11 | 36 | 75 | 4.9 | 5.8 | 34 | 41 | 2.2 | 2.9 | 15 | 20 |
| 2010 | 5.1 | 10 | 36 | 73 | 4.9 | 5.8 | 34 | 41 | 2.0 | 2.6 | 14 | 18 |
| 2011 | 4.9 | 9.2 | 34 | 64 | 3.6 | 4.3 | 25 | 30 | 1.6 | 2.1 | 11 | 15 |
| 2012 | 4.2 | 8.3 | 29 | 58 | 4.4 | 5.4 | 31 | 38 | 1.9 | 2.5 | 14 | 18 |
| 2013 | 4.0 | 7.9 | 28 | 55 | 4.1 | 4.9 | 29 | 34 | 1.6 | 2.1 | 11 | 15 |
| 2014 | 3.9 | 7.5 | 27 | 53 | 3.7 | 4.4 | 26 | 31 | 1.5 | 2.0 | 11 | 14 |
| 2015 | 3.5 | 6.7 | 24 | 47 | 3.6 | 4.2 | 25 | 29 | 1.4 | 1.7 | 10 | 12 |
| 2016 | 3.3 | 6.3 | 23 | 44 | 3.6 | 4.5 | 25 | 31 | 1.7 | 2.4 | 12 | 17 |
| 2017 | 3.1 | 6.5 | 22 | 46 | 2.8 | 3.3 | 19 | 23 | 1.2 | 1.6 | 8.1 | 11 |
| 2018 | 2.9 | 5.9 | 20 | 41 | 2.8 | 3.3 | 19 | 23 | 1.1 | 1.4 | 7.5 | 10 |

Bold values indicate exceedances

**TABLE 5-54: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BALD EAGLE
BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|-----------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 95 | 120 | 668 | 838 | 72 | 88 | 504 | 615 | 35 | 44 | 248 | 309 |
| 1994 | 58 | 78 | 406 | 547 | 50 | 60 | 347 | 417 | 25 | 31 | 176 | 216 |
| 1995 | 64 | 86 | 448 | 599 | 55 | 66 | 385 | 462 | 27 | 33 | 190 | 231 |
| 1996 | 58 | 84 | 408 | 589 | 44 | 54 | 309 | 377 | 24 | 30 | 167 | 210 |
| 1997 | 45 | 65 | 318 | 458 | 39 | 47 | 271 | 326 | 21 | 26 | 144 | 180 |
| 1998 | 38 | 54 | 265 | 378 | 35 | 42 | 244 | 296 | 17 | 21 | 121 | 148 |
| 1999 | 33 | 49 | 231 | 342 | 28 | 35 | 199 | 246 | 13 | 17 | 92 | 116 |
| 2000 | 31 | 46 | 216 | 321 | 27 | 33 | 188 | 229 | 13 | 15 | 88 | 108 |
| 2001 | 29 | 43 | 201 | 298 | 26 | 32 | 181 | 222 | 12 | 15 | 84 | 105 |
| 2002 | 28 | 42 | 199 | 293 | 28 | 34 | 193 | 236 | 13 | 16 | 90 | 112 |
| 2003 | 25 | 37 | 174 | 259 | 23 | 29 | 162 | 200 | 11 | 14 | 76 | 96 |
| 2004 | 23 | 34 | 163 | 239 | 23 | 28 | 158 | 193 | 11 | 14 | 79 | 99 |
| 2005 | 21 | 31 | 144 | 217 | 17 | 21 | 120 | 147 | 8.5 | 11 | 59 | 74 |
| 2006 | 20 | 30 | 142 | 210 | 19 | 23 | 131 | 161 | 9.0 | 11 | 63 | 77 |
| 2007 | 19 | 28 | 131 | 197 | 16 | 19 | 110 | 133 | 6.9 | 8.5 | 49 | 60 |
| 2008 | 18 | 27 | 124 | 188 | 14 | 17 | 99 | 122 | 6.7 | 8.5 | 47 | 60 |
| 2009 | 18 | 26 | 125 | 183 | 15 | 19 | 107 | 131 | 7.2 | 9.1 | 51 | 64 |
| 2010 | 16 | 23 | 110 | 160 | 14 | 17 | 97 | 120 | 6.0 | 7.5 | 42 | 52 |
| 2011 | 14 | 21 | 96 | 146 | 12 | 15 | 83 | 102 | 5.9 | 7.5 | 41 | 52 |
| 2012 | 14 | 20 | 96 | 141 | 12 | 15 | 84 | 104 | 5.7 | 7.2 | 40 | 50 |
| 2013 | 13 | 19 | 92 | 136 | 13 | 15 | 88 | 108 | 5.5 | 7.0 | 39 | 49 |
| 2014 | 12 | 17 | 83 | 121 | 12 | 15 | 83 | 102 | 5.0 | 6.2 | 35 | 44 |
| 2015 | 11 | 16 | 76 | 112 | 10 | 13 | 73 | 89 | 4.6 | 5.9 | 33 | 41 |
| 2016 | 11 | 16 | 78 | 113 | 12 | 14 | 82 | 100 | 5.0 | 6.2 | 35 | 43 |
| 2017 | 10 | 15 | 67 | 102 | 8.9 | 11 | 62 | 76 | 4.1 | 5.4 | 29 | 38 |
| 2018 | 9.3 | 14 | 65 | 97 | 8.5 | 11 | 60 | 74 | 3.7 | 4.8 | 26 | 33 |

Bold values indicate exceedances

**TABLE 5-55: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS FOR FEMALE BELTED KINGFISHER
BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|------------|-------------|-----------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 269 | 459 | 1799 | 3073 | 134 | 179 | 895 | 1199 | 66 | 116 | 441 | 777 |
| 1994 | 145 | 237 | 973 | 1585 | 111 | 142 | 742 | 950 | 51 | 79 | 343 | 530 |
| 1995 | 134 | 214 | 901 | 1430 | 120 | 161 | 802 | 1078 | 59 | 102 | 394 | 681 |
| 1996 | 135 | 222 | 905 | 1488 | 83 | 107 | 557 | 720 | 40 | 70 | 265 | 466 |
| 1997 | 105 | 188 | 700 | 1258 | 74 | 94 | 494 | 633 | 36 | 57 | 243 | 382 |
| 1998 | 91 | 152 | 608 | 1019 | 66 | 86 | 440 | 579 | 30 | 48 | 199 | 324 |
| 1999 | 79 | 140 | 527 | 938 | 54 | 69 | 359 | 464 | 22 | 34 | 150 | 226 |
| 2000 | 66 | 124 | 442 | 830 | 50 | 65 | 333 | 434 | 20 | 32 | 132 | 213 |
| 2001 | 64 | 116 | 429 | 774 | 52 | 68 | 351 | 453 | 22 | 37 | 149 | 246 |
| 2002 | 60 | 110 | 399 | 737 | 50 | 65 | 337 | 433 | 21 | 33 | 137 | 218 |
| 2003 | 57 | 104 | 381 | 695 | 43 | 56 | 289 | 373 | 18 | 31 | 123 | 205 |
| 2004 | 51 | 96 | 338 | 641 | 41 | 52 | 275 | 351 | 18 | 30 | 120 | 204 |
| 2005 | 49 | 90 | 325 | 601 | 34 | 45 | 231 | 300 | 15 | 24 | 98 | 159 |
| 2006 | 43 | 80 | 287 | 536 | 35 | 46 | 238 | 307 | 14 | 24 | 97 | 163 |
| 2007 | 42 | 79 | 283 | 526 | 32 | 42 | 217 | 282 | 12 | 21 | 82 | 137 |
| 2008 | 39 | 70 | 259 | 471 | 29 | 37 | 193 | 250 | 11 | 18 | 73 | 118 |
| 2009 | 37 | 69 | 248 | 464 | 31 | 40 | 205 | 267 | 13 | 22 | 85 | 149 |
| 2010 | 36 | 67 | 238 | 446 | 30 | 39 | 203 | 261 | 12 | 19 | 77 | 129 |
| 2011 | 33 | 59 | 224 | 393 | 24 | 31 | 159 | 209 | 10 | 16 | 64 | 110 |
| 2012 | 29 | 54 | 197 | 360 | 27 | 36 | 183 | 239 | 11 | 18 | 74 | 122 |
| 2013 | 28 | 51 | 188 | 343 | 26 | 33 | 171 | 218 | 9 | 16 | 63 | 104 |
| 2014 | 27 | 48 | 179 | 323 | 23 | 30 | 154 | 201 | 9 | 15 | 58 | 98 |
| 2015 | 24 | 44 | 163 | 292 | 22 | 29 | 150 | 191 | 8 | 13 | 54 | 85 |
| 2016 | 23 | 41 | 155 | 273 | 22 | 29 | 147 | 197 | 9 | 19 | 62 | 129 |
| 2017 | 22 | 41 | 146 | 277 | 18 | 24 | 120 | 157 | 7 | 13 | 45 | 84 |
| 2018 | 21 | 38 | 137 | 256 | 18 | 23 | 119 | 155 | 6 | 10 | 42 | 68 |

Bold values indicate exceedances

**TABLE 5-56: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS FOR FEMALE BLUE HERON
BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|------------|-------------|-----------|-----------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 291 | 511 | 1949 | 3422 | 142 | 182 | 948 | 1220 | 71 | 95 | 479 | 637 |
| 1994 | 135 | 228 | 903 | 1524 | 114 | 136 | 761 | 912 | 53 | 67 | 357 | 447 |
| 1995 | 124 | 205 | 833 | 1370 | 127 | 164 | 850 | 1098 | 64 | 85 | 430 | 566 |
| 1996 | 130 | 222 | 872 | 1487 | 83 | 99 | 553 | 666 | 41 | 55 | 275 | 369 |
| 1997 | 94 | 183 | 630 | 1227 | 73 | 86 | 486 | 575 | 38 | 48 | 256 | 322 |
| 1998 | 80 | 143 | 534 | 958 | 64 | 79 | 432 | 530 | 31 | 40 | 206 | 265 |
| 1999 | 68 | 131 | 454 | 875 | 51 | 60 | 338 | 401 | 22 | 28 | 150 | 186 |
| 2000 | 54 | 115 | 362 | 771 | 47 | 56 | 312 | 374 | 20 | 25 | 131 | 169 |
| 2001 | 53 | 106 | 357 | 713 | 50 | 60 | 338 | 403 | 23 | 30 | 155 | 201 |
| 2002 | 49 | 101 | 331 | 678 | 49 | 57 | 326 | 385 | 21 | 27 | 142 | 180 |
| 2003 | 48 | 97 | 320 | 647 | 40 | 47 | 269 | 317 | 19 | 24 | 125 | 164 |
| 2004 | 41 | 89 | 277 | 598 | 39 | 45 | 260 | 303 | 19 | 25 | 125 | 164 |
| 2005 | 41 | 84 | 272 | 560 | 31 | 37 | 210 | 247 | 15 | 19 | 100 | 128 |
| 2006 | 34 | 73 | 229 | 486 | 33 | 39 | 224 | 264 | 15 | 20 | 100 | 131 |
| 2007 | 35 | 73 | 233 | 489 | 30 | 36 | 202 | 239 | 12 | 16 | 82 | 108 |
| 2008 | 31 | 64 | 209 | 428 | 26 | 31 | 177 | 207 | 11 | 14 | 72 | 94 |
| 2009 | 30 | 64 | 202 | 430 | 29 | 35 | 195 | 234 | 13 | 18 | 89 | 119 |
| 2010 | 30 | 63 | 200 | 419 | 29 | 35 | 197 | 233 | 12 | 16 | 80 | 105 |
| 2011 | 29 | 55 | 192 | 367 | 22 | 26 | 145 | 173 | 10 | 13 | 64 | 86 |
| 2012 | 25 | 50 | 164 | 333 | 27 | 32 | 179 | 217 | 12 | 15 | 79 | 102 |
| 2013 | 24 | 47 | 159 | 317 | 25 | 29 | 166 | 195 | 10 | 13 | 65 | 85 |
| 2014 | 23 | 45 | 152 | 302 | 22 | 27 | 148 | 178 | 9 | 12 | 61 | 80 |
| 2015 | 20 | 40 | 137 | 268 | 21 | 25 | 144 | 168 | 8 | 11 | 56 | 71 |
| 2016 | 19 | 38 | 130 | 253 | 21 | 27 | 143 | 179 | 10 | 15 | 68 | 98 |
| 2017 | 19 | 39 | 124 | 262 | 17 | 20 | 111 | 133 | 7 | 10 | 47 | 65 |
| 2018 | 17 | 35 | 114 | 236 | 17 | 20 | 111 | 131 | 6 | 8 | 43 | 56 |

Bold values indicate exceedances

**TABLE 5-57: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS FOR FEMALE BALD EAGLES
BASED ON THE SUM OF TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|---------|---------|------------|------------|---------|---------|------------|------------|---------|---------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | NA | NA | 489 | 614 | NA | NA | 369 | 451 | NA | NA | 182 | 226 |
| 1994 | NA | NA | 297 | 400 | NA | NA | 254 | 306 | NA | NA | 129 | 159 |
| 1995 | NA | NA | 328 | 439 | NA | NA | 282 | 339 | NA | NA | 139 | 169 |
| 1996 | NA | NA | 299 | 431 | NA | NA | 226 | 276 | NA | NA | 122 | 154 |
| 1997 | NA | NA | 233 | 336 | NA | NA | 198 | 239 | NA | NA | 105 | 132 |
| 1998 | NA | NA | 194 | 276 | NA | NA | 178 | 217 | NA | NA | 88 | 109 |
| 1999 | NA | NA | 169 | 251 | NA | NA | 146 | 180 | NA | NA | 67 | 85 |
| 2000 | NA | NA | 158 | 235 | NA | NA | 137 | 168 | NA | NA | 64 | 79 |
| 2001 | NA | NA | 148 | 218 | NA | NA | 132 | 162 | NA | NA | 62 | 77 |
| 2002 | NA | NA | 146 | 214 | NA | NA | 141 | 173 | NA | NA | 66 | 82 |
| 2003 | NA | NA | 127 | 189 | NA | NA | 118 | 146 | NA | NA | 56 | 70 |
| 2004 | NA | NA | 119 | 175 | NA | NA | 115 | 141 | NA | NA | 58 | 72 |
| 2005 | NA | NA | 106 | 159 | NA | NA | 88 | 108 | NA | NA | 44 | 54 |
| 2006 | NA | NA | 104 | 154 | NA | NA | 96 | 118 | NA | NA | 46 | 56 |
| 2007 | NA | NA | 96 | 145 | NA | NA | 80 | 98 | NA | NA | 36 | 44 |
| 2008 | NA | NA | 90 | 138 | NA | NA | 72 | 89 | NA | NA | 34 | 44 |
| 2009 | NA | NA | 92 | 134 | NA | NA | 78 | 96 | NA | NA | 37 | 47 |
| 2010 | NA | NA | 80 | 118 | NA | NA | 71 | 88 | NA | NA | 31 | 38 |
| 2011 | NA | NA | 70 | 107 | NA | NA | 61 | 75 | NA | NA | 30 | 38 |
| 2012 | NA | NA | 70 | 103 | NA | NA | 62 | 76 | NA | NA | 29 | 37 |
| 2013 | NA | NA | 67 | 99 | NA | NA | 64 | 79 | NA | NA | 28 | 36 |
| 2014 | NA | NA | 61 | 89 | NA | NA | 61 | 75 | NA | NA | 26 | 32 |
| 2015 | NA | NA | 56 | 82 | NA | NA | 54 | 65 | NA | NA | 24 | 30 |
| 2016 | NA | NA | 57 | 83 | NA | NA | 60 | 74 | NA | NA | 26 | 32 |
| 2017 | NA | NA | 49 | 75 | NA | NA | 46 | 56 | NA | NA | 21 | 28 |
| 2018 | NA | NA | 48 | 71 | NA | NA | 44 | 54 | NA | NA | 19 | 25 |

Bold values indicate exceedances

**TABLE 5-58: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE BELTED KINGFISHER ON TEQ BASIS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|-------------|-------------|-------------|---|-------------|-------------|--------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% |
| | ADD | ADD | ADD | ADD | Conc. | Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 124 | 250 | 1237 | 2498 | 4078 | 8287 | 8157 | 16574 |
| Stillwater (168) | 54 | 103 | 537 | 1027 | 1575 | 3049 | 3149 | 6099 |
| Federal Dam (154) | 12 | 17 | 116 | 174 | 370 | 552 | 741 | 1104 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 11 | 14 | 112 | 138 | 371 | 456 | 742 | 911 |
| 137.2 | 23 | 50 | 226 | 499 | 750 | 1657 | 1500 | 3313 |
| 122.4 | 8.8 | 14 | 88 | 144 | 289 | 475 | 578 | 951 |
| 113.8 | 9.2 | 10 | 92 | 105 | 303 | 340 | 606 | 680 |
| 100 | 4.0 | 9.3 | 40 | 93 | 131 | 248 | 263 | 497 |
| 88.9 | 7.8 | 11 | 78 | 109 | 255 | 351 | 511 | 702 |
| 58.7 | 8.4 | 12 | 84 | 117 | 282 | 372 | 565 | 744 |
| 47.3 | 7.8 | 13 | 78 | 126 | 253 | 379 | 506 | 759 |
| 25.8 | 5.7 | 7.1 | 57 | 71 | 187 | 226 | 374 | 451 |

Bold values indicate exceedances

**TABLE 5-59: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE GREAT BLUE HERON ON TEQ BASIS**

| Location | <<<< ---- Dietary Dose ---- >>>> | | | | <<<< ---- Egg Concentration ---- >>>> | | | |
|----------------------------|----------------------------------|-------------|-------------|-------------|---------------------------------------|-------------|-------------|------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% UCL | vs. Average | vs. 95% |
| | ADD | ADD | ADD | Conc. | Conc. | Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 62 | 125 | 616 | 1245 | 204 | 417 | 340 | 694 |
| Stillwater (168) | 26 | 39 | 256 | 388 | 69 | 98 | 115 | 164 |
| Federal Dam (154) | 5.2 | 7.7 | 52 | 77 | 16 | 23 | 27 | 39 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 5.6 | 6.8 | 56 | 68 | 19 | 23 | 31 | 38 |
| 137.2 | 11 | 25 | 114 | 246 | 38 | 82 | 63 | 137 |
| 122.4 | 4.4 | 7.0 | 44 | 70 | 15 | 23 | 24 | 39 |
| 113.8 | 4.6 | 4.9 | 46 | 49 | 15 | 16 | 25 | 26 |
| 100 | 2.0 | 4.9 | 20 | 49 | 7 | 11 | 11 | 19 |
| 88.9 | 4.0 | 5.6 | 40 | 56 | 13 | 18 | 22 | 30 |
| 58.7 | 4.2 | 5.2 | 42 | 52 | 14 | 16 | 24 | 27 |
| 47.3 | 4.0 | 6.0 | 40 | 60 | 13 | 17 | 21 | 28 |
| 25.8 | 2.9 | 3.6 | 29 | 36 | 10 | 11 | 16 | 19 |

Bold values indicate exceedances

**TABLE 5-60: RATIO OF MODELED DIETARY DOSE AND EGG CONCENTRATIONS TO BENCHMARKS
BASED ON 1993 DATA FOR FEMALE BALD EAGLE ON TEQ BASIS**

| Location | <<<< ----- Dietary Dose ----- >>>> | | | | <<<< ----- Egg Concentration ----- >>>> | | | |
|----------------------------|------------------------------------|------------|-------------|-------------|---|--------------|--------------|--------------|
| | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
| | vs. Average | vs. 95% | vs. Average | vs. 95% | vs. Average | vs. 95% UCL | vs. Average | vs. 95% |
| | ADD | UCL ADD | ADD | UCL ADD | Conc. | Conc | Conc. | UCL Conc. |
| | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard | Hazard |
| Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | Quotient | |
| <i>Upper River</i> | | | | | | | | |
| Thompson Island Pool (189) | 221 | 427 | 2208 | 4272 | 23037 | 44582 | 46075 | 89164 |
| Stillwater (168) | 39 | 51 | 392 | 505 | 4095 | 5274 | 8189 | 10547 |
| Federal Dam (154) | 28 | 51 | 285 | 511 | 2970 | 5337 | 5940 | 10674 |
| <i>Lower River</i> | | | | | | | | |
| 143.5 | 28 | 51 | 285 | 512 | 2970 | 5337 | 5940 | 10674 |
| 137.2 | 107 | 256 | 1074 | 2555 | 11204 | 26663 | 22409 | 53327 |
| 122.4 | 25 | 34 | 250 | 344 | 2611 | 3588 | 5222 | 7177 |
| 113.8 | 23 | 32 | 226 | 317 | 2362 | 3302 | 4723 | 6605 |
| 100 | 26 | 80 | 260 | 803 | 2711 | 8377 | 5423 | 16754 |
| 88.9 | 17 | 32 | 167 | 318 | 1746 | 3322 | 3493 | 6643 |
| 58.7 | 19 | 29 | 194 | 288 | 2029 | 3007 | 4059 | 6014 |
| 47.3 | 22 | 60 | 223 | 599 | 2323 | 6254 | 4645 | 12508 |
| 25.8 | 16 | 31 | 157 | 312 | 1641 | 3260 | 3282 | 6520 |

Bold values indicate exceedances

TABLE 5-61: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BELTED KINGFISHER USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|-----------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 147 | 241 | 1472 | 2412 | 68 | 87 | 682 | 865 | 34 | 60 | 338 | 602 |
| 1994 | 81 | 122 | 809 | 1224 | 56 | 67 | 561 | 669 | 26 | 41 | 260 | 405 |
| 1995 | 75 | 111 | 751 | 1111 | 61 | 78 | 613 | 779 | 30 | 53 | 304 | 529 |
| 1996 | 75 | 116 | 753 | 1158 | 42 | 50 | 421 | 503 | 20 | 36 | 202 | 359 |
| 1997 | 59 | 98 | 587 | 979 | 37 | 44 | 372 | 439 | 19 | 29 | 186 | 294 |
| 1998 | 51 | 79 | 509 | 792 | 33 | 40 | 332 | 402 | 15 | 25 | 151 | 249 |
| 1999 | 44 | 72 | 444 | 724 | 27 | 32 | 268 | 316 | 11 | 17 | 114 | 172 |
| 2000 | 37 | 64 | 373 | 643 | 25 | 29 | 248 | 294 | 10 | 16 | 100 | 163 |
| 2001 | 36 | 60 | 362 | 599 | 26 | 31 | 262 | 310 | 11 | 19 | 113 | 189 |
| 2002 | 34 | 57 | 338 | 569 | 25 | 30 | 253 | 297 | 10 | 17 | 104 | 167 |
| 2003 | 32 | 54 | 322 | 539 | 22 | 25 | 215 | 252 | 9.3 | 16 | 93 | 157 |
| 2004 | 29 | 50 | 285 | 497 | 21 | 24 | 206 | 239 | 9.1 | 16 | 91 | 157 |
| 2005 | 27 | 46 | 273 | 464 | 17 | 20 | 171 | 200 | 7.4 | 12 | 74 | 122 |
| 2006 | 24 | 41 | 243 | 414 | 18 | 21 | 177 | 209 | 7.4 | 13 | 74 | 125 |
| 2007 | 24 | 41 | 239 | 408 | 16 | 19 | 162 | 190 | 6.2 | 11 | 62 | 105 |
| 2008 | 22 | 36 | 218 | 363 | 14 | 17 | 143 | 168 | 5.5 | 9.0 | 55 | 90 |
| 2009 | 21 | 36 | 209 | 360 | 15 | 18 | 153 | 182 | 6.4 | 12 | 64 | 116 |
| 2010 | 20 | 35 | 202 | 346 | 15 | 18 | 153 | 180 | 5.9 | 10 | 59 | 99 |
| 2011 | 19 | 30 | 188 | 304 | 12 | 14 | 118 | 140 | 4.8 | 8.4 | 48 | 84 |
| 2012 | 17 | 28 | 166 | 278 | 14 | 17 | 137 | 165 | 5.6 | 9.4 | 56 | 94 |
| 2013 | 16 | 26 | 159 | 264 | 13 | 15 | 128 | 150 | 4.8 | 8.0 | 48 | 80 |
| 2014 | 15 | 25 | 150 | 249 | 12 | 14 | 115 | 137 | 4.4 | 7.5 | 44 | 75 |
| 2015 | 14 | 22 | 137 | 225 | 11 | 13 | 112 | 130 | 4.1 | 6.5 | 41 | 65 |
| 2016 | 13 | 21 | 131 | 212 | 11 | 14 | 111 | 136 | 4.8 | 10 | 48 | 101 |
| 2017 | 12 | 21 | 123 | 214 | 8.9 | 11 | 89 | 106 | 3.4 | 6.5 | 34 | 65 |
| 2018 | 11 | 19 | 114 | 195 | 8.8 | 10 | 88 | 104 | 3.2 | 5.2 | 32 | 52 |

Bold values indicate exceedances

TABLE 5-62: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE GREAT BLUE HERON USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|-------------|------------|------------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 66 | 112 | 661 | 1124 | 31 | 40 | 313 | 398 | 16 | 21 | 157 | 207 |
| 1994 | 33 | 53 | 330 | 527 | 25 | 30 | 253 | 301 | 12 | 15 | 119 | 146 |
| 1995 | 31 | 48 | 305 | 476 | 28 | 36 | 281 | 359 | 14 | 18 | 141 | 184 |
| 1996 | 31 | 51 | 313 | 508 | 19 | 22 | 186 | 222 | 9.1 | 12 | 91 | 121 |
| 1997 | 23 | 42 | 235 | 423 | 16 | 19 | 164 | 192 | 8.5 | 11 | 85 | 106 |
| 1998 | 20 | 34 | 201 | 335 | 15 | 18 | 146 | 177 | 6.8 | 8.7 | 68 | 87 |
| 1999 | 17 | 31 | 173 | 306 | 12 | 14 | 116 | 136 | 5.0 | 6.2 | 50 | 62 |
| 2000 | 14 | 27 | 141 | 271 | 11 | 13 | 107 | 126 | 4.4 | 5.6 | 44 | 56 |
| 2001 | 14 | 25 | 138 | 251 | 11 | 14 | 115 | 135 | 5.2 | 6.6 | 52 | 66 |
| 2002 | 13 | 24 | 129 | 239 | 11 | 13 | 111 | 129 | 4.7 | 5.9 | 47 | 59 |
| 2003 | 12 | 23 | 124 | 227 | 9.3 | 11 | 93 | 108 | 4.2 | 5.4 | 42 | 54 |
| 2004 | 11 | 21 | 108 | 210 | 8.9 | 10 | 89 | 103 | 4.1 | 5.4 | 41 | 54 |
| 2005 | 10 | 20 | 105 | 196 | 7.3 | 8.4 | 73 | 84 | 3.3 | 4.2 | 33 | 42 |
| 2006 | 9.1 | 17 | 91 | 172 | 7.7 | 9.0 | 77 | 90 | 3.3 | 4.3 | 33 | 43 |
| 2007 | 9.1 | 17 | 91 | 172 | 7.0 | 8.1 | 70 | 81 | 2.8 | 3.6 | 28 | 36 |
| 2008 | 8.2 | 15 | 82 | 151 | 6.1 | 7.1 | 61 | 71 | 2.4 | 3.1 | 24 | 31 |
| 2009 | 7.9 | 15 | 79 | 151 | 6.7 | 7.9 | 67 | 79 | 2.9 | 3.9 | 29 | 39 |
| 2010 | 7.7 | 15 | 77 | 147 | 6.7 | 7.8 | 67 | 78 | 2.7 | 3.4 | 27 | 34 |
| 2011 | 7.3 | 13 | 73 | 129 | 5.0 | 5.9 | 50 | 59 | 2.2 | 2.8 | 22 | 28 |
| 2012 | 6.3 | 12 | 63 | 117 | 6.1 | 7.3 | 61 | 73 | 2.6 | 3.3 | 26 | 33 |
| 2013 | 6.1 | 11 | 61 | 111 | 5.6 | 6.5 | 56 | 65 | 2.2 | 2.8 | 22 | 28 |
| 2014 | 5.8 | 11 | 58 | 105 | 5.0 | 6.0 | 50 | 60 | 2.0 | 2.6 | 20 | 26 |
| 2015 | 5.3 | 9.4 | 53 | 94 | 4.9 | 5.6 | 49 | 56 | 1.9 | 2.3 | 19 | 23 |
| 2016 | 5.0 | 8.9 | 50 | 89 | 4.9 | 6.0 | 49 | 60 | 2.2 | 3.2 | 22 | 32 |
| 2017 | 4.7 | 9.1 | 47 | 91 | 3.8 | 4.5 | 38 | 45 | 1.6 | 2.1 | 16 | 21 |
| 2018 | 4.4 | 8.2 | 44 | 82 | 3.8 | 4.4 | 38 | 44 | 1.4 | 1.8 | 14 | 18 |

Bold values indicate exceedances

TABLE 5-63: RATIO OF MODELED DIETARY DOSE TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BALD EAGLE USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|-----------|------------|------------|-------------|-----------|-----------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 122 | 154 | 1225 | 1536 | 92 | 113 | 924 | 1128 | 45 | 57 | 455 | 567 |
| 1994 | 74 | 100 | 745 | 1002 | 64 | 77 | 635 | 765 | 32 | 40 | 322 | 397 |
| 1995 | 82 | 110 | 821 | 1098 | 71 | 85 | 705 | 848 | 35 | 42 | 348 | 423 |
| 1996 | 75 | 108 | 749 | 1079 | 57 | 69 | 567 | 692 | 31 | 39 | 306 | 386 |
| 1997 | 58 | 84 | 583 | 840 | 50 | 60 | 496 | 598 | 26 | 33 | 263 | 330 |
| 1998 | 49 | 69 | 487 | 692 | 45 | 54 | 447 | 543 | 22 | 27 | 221 | 272 |
| 1999 | 42 | 63 | 423 | 628 | 36 | 45 | 364 | 451 | 17 | 21 | 169 | 213 |
| 2000 | 40 | 59 | 395 | 588 | 34 | 42 | 344 | 421 | 16 | 20 | 161 | 199 |
| 2001 | 37 | 55 | 369 | 546 | 33 | 41 | 331 | 407 | 15 | 19 | 154 | 192 |
| 2002 | 36 | 54 | 364 | 536 | 35 | 43 | 353 | 432 | 16 | 21 | 164 | 206 |
| 2003 | 32 | 47 | 318 | 474 | 30 | 37 | 296 | 366 | 14 | 18 | 140 | 176 |
| 2004 | 30 | 44 | 298 | 438 | 29 | 35 | 289 | 354 | 14 | 18 | 145 | 181 |
| 2005 | 26 | 40 | 265 | 399 | 22 | 27 | 221 | 270 | 11 | 14 | 109 | 136 |
| 2006 | 26 | 39 | 260 | 385 | 24 | 29 | 241 | 295 | 11 | 14 | 115 | 141 |
| 2007 | 24 | 36 | 240 | 362 | 20 | 24 | 201 | 244 | 9 | 11 | 89 | 109 |
| 2008 | 23 | 35 | 227 | 345 | 18 | 22 | 181 | 223 | 9 | 11 | 86 | 109 |
| 2009 | 23 | 34 | 229 | 335 | 20 | 24 | 196 | 240 | 9 | 12 | 93 | 117 |
| 2010 | 20 | 29 | 201 | 294 | 18 | 22 | 179 | 219 | 8 | 10 | 78 | 96 |
| 2011 | 18 | 27 | 176 | 268 | 15 | 19 | 152 | 187 | 8 | 10 | 75 | 96 |
| 2012 | 18 | 26 | 175 | 259 | 15 | 19 | 154 | 190 | 7 | 9 | 73 | 92 |
| 2013 | 17 | 25 | 168 | 249 | 16 | 20 | 160 | 198 | 7 | 9 | 71 | 90 |
| 2014 | 15 | 22 | 153 | 222 | 15 | 19 | 153 | 187 | 6 | 8 | 65 | 80 |
| 2015 | 14 | 21 | 140 | 206 | 13 | 16 | 134 | 163 | 6 | 8 | 60 | 75 |
| 2016 | 14 | 21 | 143 | 207 | 15 | 18 | 150 | 184 | 6 | 8 | 64 | 79 |
| 2017 | 12 | 19 | 123 | 187 | 11 | 14 | 114 | 140 | 5 | 7 | 53 | 69 |
| 2018 | 12 | 18 | 119 | 178 | 11 | 13 | 109 | 135 | 5 | 6 | 48 | 61 |

Bold values indicate exceedances

TABLE 5-64: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BELTED KINGFISHER USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL |
| 1993 | 4201 | 7327 | 8401 | 1465 | 2054 | 2670 | 4108 | 5340 | 1031 | 1920 | 2062 | 3841 |
| 1994 | 2023 | 3383 | 4046 | 677 | 1661 | 2025 | 3322 | 4050 | 778 | 1266 | 1555 | 2531 |
| 1995 | 1868 | 3044 | 3736 | 609 | 1842 | 2402 | 3684 | 4805 | 926 | 1684 | 1852 | 3367 |
| 1996 | 1935 | 3269 | 3870 | 654 | 1217 | 1494 | 2434 | 2987 | 598 | 1128 | 1196 | 2256 |
| 1997 | 1424 | 2714 | 2848 | 543 | 1072 | 1296 | 2145 | 2593 | 555 | 921 | 1110 | 1842 |
| 1998 | 1216 | 2139 | 2431 | 428 | 954 | 1191 | 1908 | 2383 | 448 | 778 | 896 | 1555 |
| 1999 | 1038 | 1958 | 2076 | 392 | 755 | 916 | 1510 | 1832 | 329 | 526 | 658 | 1053 |
| 2000 | 840 | 1727 | 1680 | 345 | 698 | 854 | 1396 | 1709 | 289 | 501 | 578 | 1002 |
| 2001 | 824 | 1601 | 1649 | 320 | 750 | 913 | 1500 | 1826 | 338 | 592 | 677 | 1183 |
| 2002 | 765 | 1523 | 1531 | 305 | 723 | 873 | 1445 | 1746 | 309 | 520 | 618 | 1040 |
| 2003 | 738 | 1449 | 1476 | 290 | 602 | 728 | 1205 | 1455 | 273 | 489 | 547 | 978 |
| 2004 | 643 | 1338 | 1286 | 268 | 580 | 692 | 1160 | 1385 | 271 | 493 | 542 | 986 |
| 2005 | 627 | 1253 | 1254 | 251 | 473 | 573 | 946 | 1146 | 218 | 377 | 437 | 755 |
| 2006 | 536 | 1096 | 1072 | 219 | 499 | 605 | 999 | 1210 | 217 | 391 | 435 | 783 |
| 2007 | 539 | 1094 | 1079 | 219 | 453 | 550 | 906 | 1099 | 180 | 327 | 361 | 654 |
| 2008 | 487 | 964 | 974 | 193 | 397 | 479 | 795 | 959 | 159 | 278 | 319 | 556 |
| 2009 | 470 | 964 | 939 | 193 | 434 | 532 | 869 | 1065 | 192 | 364 | 385 | 727 |
| 2010 | 461 | 936 | 922 | 187 | 436 | 527 | 872 | 1054 | 175 | 312 | 350 | 623 |
| 2011 | 439 | 821 | 879 | 164 | 327 | 400 | 655 | 800 | 141 | 264 | 283 | 527 |
| 2012 | 379 | 747 | 759 | 149 | 396 | 490 | 791 | 980 | 171 | 298 | 342 | 596 |
| 2013 | 367 | 712 | 733 | 142 | 367 | 441 | 735 | 883 | 142 | 251 | 285 | 503 |
| 2014 | 349 | 675 | 699 | 135 | 329 | 403 | 658 | 807 | 132 | 237 | 264 | 473 |
| 2015 | 315 | 603 | 630 | 121 | 319 | 382 | 639 | 764 | 123 | 202 | 246 | 404 |
| 2016 | 300 | 567 | 600 | 113 | 317 | 404 | 633 | 808 | 146 | 323 | 292 | 647 |
| 2017 | 285 | 584 | 569 | 117 | 249 | 306 | 499 | 613 | 102 | 204 | 203 | 409 |
| 2018 | 263 | 529 | 526 | 106 | 248 | 302 | 496 | 603 | 94 | 163 | 188 | 327 |

Bold values indicate exceedances

TABLE 5-65: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS BASED ON FISHRAND FOR FEMALE GREAT BLUE HERON USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|------------|------------|------------|-----------|-----------|-----------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 209 | 367 | 349 | 612 | 102 | 131 | 170 | 218 | 51 | 68 | 86 | 114 |
| 1994 | 97 | 163 | 161 | 272 | 82 | 98 | 136 | 163 | 38 | 48 | 64 | 80 |
| 1995 | 89 | 147 | 149 | 245 | 91 | 118 | 152 | 196 | 46 | 61 | 77 | 101 |
| 1996 | 94 | 160 | 156 | 266 | 59 | 71 | 99 | 119 | 29 | 40 | 49 | 66 |
| 1997 | 68 | 132 | 113 | 219 | 52 | 62 | 87 | 103 | 27 | 35 | 46 | 58 |
| 1998 | 57 | 103 | 96 | 171 | 46 | 57 | 77 | 95 | 22 | 28 | 37 | 47 |
| 1999 | 49 | 94 | 81 | 156 | 36 | 43 | 61 | 72 | 16 | 20 | 27 | 33 |
| 2000 | 39 | 83 | 65 | 138 | 34 | 40 | 56 | 67 | 14 | 18 | 23 | 30 |
| 2001 | 38 | 76 | 64 | 127 | 36 | 43 | 60 | 72 | 17 | 22 | 28 | 36 |
| 2002 | 36 | 73 | 59 | 121 | 35 | 41 | 58 | 69 | 15 | 19 | 25 | 32 |
| 2003 | 34 | 69 | 57 | 116 | 29 | 34 | 48 | 57 | 13 | 18 | 22 | 29 |
| 2004 | 30 | 64 | 50 | 107 | 28 | 32 | 47 | 54 | 13 | 18 | 22 | 29 |
| 2005 | 29 | 60 | 49 | 100 | 23 | 27 | 38 | 44 | 11 | 14 | 18 | 23 |
| 2006 | 25 | 52 | 41 | 87 | 24 | 28 | 40 | 47 | 11 | 14 | 18 | 23 |
| 2007 | 25 | 52 | 42 | 87 | 22 | 26 | 36 | 43 | 9 | 12 | 15 | 19 |
| 2008 | 22 | 46 | 37 | 77 | 19 | 22 | 32 | 37 | 8 | 10 | 13 | 17 |
| 2009 | 22 | 46 | 36 | 77 | 21 | 25 | 35 | 42 | 10 | 13 | 16 | 21 |
| 2010 | 21 | 45 | 36 | 75 | 21 | 25 | 35 | 42 | 9 | 11 | 14 | 19 |
| 2011 | 21 | 39 | 34 | 66 | 16 | 19 | 26 | 31 | 7 | 9 | 12 | 15 |
| 2012 | 18 | 36 | 29 | 60 | 19 | 23 | 32 | 39 | 8 | 11 | 14 | 18 |
| 2013 | 17 | 34 | 29 | 57 | 18 | 21 | 30 | 35 | 7 | 9 | 12 | 15 |
| 2014 | 16 | 32 | 27 | 54 | 16 | 19 | 27 | 32 | 7 | 9 | 11 | 14 |
| 2015 | 15 | 29 | 24 | 48 | 15 | 18 | 26 | 30 | 6 | 8 | 10 | 13 |
| 2016 | 14 | 27 | 23 | 45 | 15 | 19 | 26 | 32 | 7 | 11 | 12 | 18 |
| 2017 | 13 | 28 | 22 | 47 | 12 | 14 | 20 | 24 | 5 | 7 | 8 | 12 |
| 2018 | 12 | 25 | 20 | 42 | 12 | 14 | 20 | 23 | 5 | 6 | 8 | 10 |

Bold values indicate exceedances

TABLE 5-66: RATIO OF MODELED EGG CONCENTRATIONS TO BENCHMARKS BASED ON FISHRAND FOR FEMALE BALD EAGLE USING TEQ FOR THE PERIOD 1993 - 2018

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|--------------|--------------|--------------|--------------|-------------|--------------|--------------|--------------|-------------|-------------|-------------|--------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 12779 | 16033 | 25558 | 32065 | 9646 | 11769 | 19292 | 23538 | 4743 | 5915 | 9486 | 11829 |
| 1994 | 7771 | 10459 | 15543 | 20918 | 6630 | 7983 | 13261 | 15966 | 3362 | 4141 | 6724 | 8282 |
| 1995 | 8564 | 11457 | 17127 | 22914 | 7361 | 8845 | 14721 | 17690 | 3633 | 4417 | 7267 | 8835 |
| 1996 | 7811 | 11260 | 15623 | 22521 | 5916 | 7219 | 11833 | 14438 | 3189 | 4024 | 6377 | 8049 |
| 1997 | 6083 | 8771 | 12166 | 17542 | 5180 | 6241 | 10361 | 12483 | 2748 | 3444 | 5495 | 6888 |
| 1998 | 5078 | 7222 | 10156 | 14444 | 4660 | 5667 | 9319 | 11334 | 2311 | 2837 | 4623 | 5673 |
| 1999 | 4416 | 6550 | 8832 | 13099 | 3803 | 4703 | 7605 | 9406 | 1760 | 2220 | 3520 | 4440 |
| 2000 | 4127 | 6132 | 8254 | 12264 | 3590 | 4388 | 7181 | 8776 | 1678 | 2072 | 3356 | 4145 |
| 2001 | 3854 | 5695 | 7709 | 11389 | 3457 | 4244 | 6915 | 8488 | 1612 | 2003 | 3223 | 4006 |
| 2002 | 3802 | 5597 | 7604 | 11194 | 3684 | 4512 | 7367 | 9025 | 1713 | 2148 | 3427 | 4295 |
| 2003 | 3321 | 4947 | 6641 | 9894 | 3094 | 3821 | 6188 | 7642 | 1461 | 1838 | 2923 | 3675 |
| 2004 | 3109 | 4575 | 6219 | 9151 | 3014 | 3695 | 6029 | 7391 | 1508 | 1891 | 3016 | 3781 |
| 2005 | 2761 | 4160 | 5522 | 8319 | 2303 | 2815 | 4607 | 5629 | 1137 | 1423 | 2274 | 2845 |
| 2006 | 2716 | 4020 | 5432 | 8041 | 2511 | 3075 | 5023 | 6151 | 1199 | 1472 | 2397 | 2944 |
| 2007 | 2503 | 3777 | 5007 | 7554 | 2098 | 2550 | 4195 | 5101 | 929 | 1142 | 1858 | 2283 |
| 2008 | 2364 | 3601 | 4727 | 7203 | 1893 | 2332 | 3786 | 4664 | 900 | 1139 | 1800 | 2279 |
| 2009 | 2390 | 3497 | 4781 | 6994 | 2045 | 2500 | 4090 | 5001 | 969 | 1220 | 1937 | 2441 |
| 2010 | 2101 | 3069 | 4202 | 6139 | 1865 | 2288 | 3729 | 4577 | 809 | 1004 | 1619 | 2008 |
| 2011 | 1840 | 2797 | 3681 | 5595 | 1591 | 1953 | 3181 | 3906 | 786 | 998 | 1571 | 1995 |
| 2012 | 1827 | 2702 | 3655 | 5404 | 1611 | 1983 | 3221 | 3966 | 764 | 964 | 1528 | 1928 |
| 2013 | 1755 | 2599 | 3511 | 5197 | 1675 | 2062 | 3349 | 4124 | 737 | 934 | 1474 | 1869 |
| 2014 | 1597 | 2320 | 3194 | 4639 | 1596 | 1956 | 3191 | 3913 | 674 | 836 | 1348 | 1672 |
| 2015 | 1463 | 2147 | 2926 | 4294 | 1400 | 1706 | 2799 | 3411 | 622 | 784 | 1244 | 1568 |
| 2016 | 1493 | 2161 | 2985 | 4322 | 1570 | 1920 | 3140 | 3841 | 670 | 830 | 1339 | 1659 |
| 2017 | 1285 | 1946 | 2570 | 3893 | 1191 | 1463 | 2381 | 2926 | 550 | 724 | 1099 | 1447 |
| 2018 | 1246 | 1861 | 2492 | 3721 | 1142 | 1408 | 2284 | 2815 | 499 | 641 | 997 | 1281 |

Bold values indicate exceedances

TABLE 5-67: WILDLIFE SURVEY RESULTS - Birds

Hudson River
New York

| Information Source | Date | Contact | Response | Contact Information | Data Available | Information/Findings |
|--|--------------------------|--------------|--------------------------------|---|--|--|
| Birds | | | | | | |
| Hudsonia | 2-Jun-99 | Call/Fax | YES; Spoke with on 6/2/1999 | Eric Kiviat, Executive Director; (914) 758-7273 (7274) OR (914) 758-7053; FAX: (914) 758-7033; EMAIL: kiviat@bard.edu | He has no direct knowledge of the upper Hudson but provided names | WATERFOWL/MALLARD: Steve Brown - Delmar NYSDEC KINGFISHER: Breeding bird atlas - DEC now computerized on web page; Bob Anderle/Janet Carroll - NYSDEC NYSDEC - Natural Resource Damage Assessment |
| NYS Department of Environmental Conservation Endangered Species Unit | 3-Jun-99 | Call | No | Peter Nye (518) 439-7635x9 (Eagle Specialist); www.dec.state.ny.us | Left Message - Will call back | |
| Manomet Center for Conservation Sciences | 2-Jun-99 | Email | No | John M. Hagen, Division Director (Conservation Forestry Staff); jmhagan@ime.net; www.manomet.org; | Left Message - Will call back | |
| Saratoga National Historic Park, Stillwater, NY | 4-Jun-99 | Call | No | Chris (wildlife manager) (518) 664-9821x5; also can contact Richard Beresford | Left Message - Will call back | |
| Federation of New York State Bird Clubs | 3-Jun-99 | Email | No | Valeria Freer, President (vfreer@sullivan.suny.edu); http://www.birds.cornell.edu/fnysbc | | |
| Union College Professor Emeritus | 2-Jun-1999 7-Jun-1999 | Call Call | No Yes | Carl George (518) 388-6330; Bird Expert; (John Waldman - Hudson River Foundation Recommended I call) | He did not have any specific data, but recommended a number of different sources | He recommended that I contact: Bob Daniels (mammals) - NY State Museum; Walter Sabin (Hudson-Mohawk Bird Club, they do an intensive waterbird survey and publish results in the <i>Kingbird Journal</i> (518) 439-7344; Also Union College has survey information for a lake in Scotia near the Hudson for Collins Lake in Scotia (across river from Schenectady) - http://tardis.union.edu/~birds, presents 10 years of bird information - 15 air miles from Hudson; also recommended contacting Robert Yunick for regional baseline information from Audubon Christmas count and the mid-May Big Day |
| Manomet Center for Conservation Sciences | 7-Jun-99 | Email | No | Dr. Treavor Lloyd-Evans (tlloyd-evans@manomet.org) - avian expert | Avian Conservationist | |

TABLE 5-67: WILDLIFE SURVEY RESULTS - Birds

Hudson River
New York

| Information Source | Date | Contact | Response | Contact Information | Data Available | Information/Findings |
|---|----------|-------------|----------|--|--|--|
| American Birding Association - Online | 7-Jun-99 | WWW | No | www.americanbirding.org | Good links - possibility for some bird information on Hudson | |
| Breeding Bird Survey - OnLine | 7-Jun-99 | WWW | No | www.mbr.nbs.gov/bbs/bbs.html | Regional trend analysis by species - region=NY State, some additional details may be available | |
| Hudson-Mohawk Bird Club | 7-Jun-99 | Call | No | Walter Sabin Home: (518) 439-7344 | Intensive waterbird survey every year - publish results in Kingbird Journal | |
| Ornithologist | 7-Jun-99 | Need Number | No | Robert Yunick | regional baseline data from Audubon Christmas count and mid-May Big Day | |
| NYS Department of Environmental Conservation Endangered Species Unit | 8-Jun-99 | WWW | No | www.dec.state.ny.us/website/dfwmr/wildlife/endspec/enspbird.html | Brief summaries, listed by species, for NY State. | <i>Ixobrychus exilis</i> (Least Bittern): Populations along Hudson River Valley, uncommon and rare breeder, declines due to loss of marsh habitat due to drainage, vegetational changes, pollution, insecticides. <i>Rallus elegans</i> (King Rail): Nesting was reported in northern Hudson Valley, however there are no confirmed nests in NY state currently, decline due to degradation of wetlands. <i>Bartramia longicauda</i> (Upland Sand Piper): once common around NY state including Hudson, less than 250 breeding sites to date in NY, decline due to loss of grassland habitat. All considered threatened species. |
| Andrle, R. F. and Carroll, J. R. (ed.) 1988. <u>The Atlas of Breeding Birds in New York State</u> . Cornell University Press, Ithica. | 8-Jun-99 | | | | Regional trend analysis by species - region=NY State, some additional details may be available | <i>Tachycineta bicolor</i> (Tree Swallow): Common breeder throughout entire state. <i>Ceryle alcyon</i> (Belted Kingfisher): Common summer resident throughout entire state. <i>Ardea herodias</i> (Great Blue Heron): Observed in Northern Hudson Valley, possibility of breeding there. <i>Anas platyrhynchos</i> (Mallard): Common breeder in wetlands. In the 1900's, rarely if ever seen as a breeder; creation/improvement of wetlands in mid-1900's and release of captive-bred adults and ducklings in the 1950's caused populations to increase. <u>Birds not found in Northern Hudson Valley</u> : Eagles and Osprey. |

TABLE 5-67: WILDLIFE SURVEY RESULTS - Birds

Hudson River
New York

| Information Source | Date | Contact | Response | Contact Information | Data Available | Information/Findings |
|-----------------------------|---------------------|-----------------------|-----------------|---|---|---|
| NYSDEC | 16-Jun-99 | Call | Yes | Mark Brown (518) 623-3671 | Familiar with the area regarding mammals, birds, and herps. Good source. See General Info page. | This area is rich in birds including water fowl. Bald Eagle is only a winter resident, migrates in the summer. Lots of Canada geese and mallard. Has not seen any Osprey nests. They only feed here and spend most of their time around the near-by lakes. Has also seen tree swallow, kingfisher, and great blue heron. Most of the water fowl and larger birds use the area for feeding but do not breed here. He hasn't seen many nests except those built by species which live in the more wooded areas. Here's a list of the other species he has seen in the area: Common Merganser (Diving Duck), red tailed hawk, sparrow hawk, rough grouse, wild turkey, killdare, wood cock, morning dove, barn owl, bard owl, sawhat owl (occupying nest boxes built for ducks), swallows, ravens, crows, wrens, eastern blue bird, starlings. |
| Ndakinna Wilderness Project | 6/3/1999 6/16/99 | Email Call Call | No No Yes | Jim Brushek (518) 583-9980x3, 23 Middle Grove Road, Greenfield Center, NY 12833; Received address from Saratoga County Information - Annamaria Dalton (annamaria@spa.net) | Professional Tracker | Saw some bald eagles 3 or 4 weeks ago. Hasn't seen any osprey. Great Blue Heron and kingfisher in large numbers. Hasn't seen any tree swallow. Lots of mallards and Canada geese. Could not recall seeing any nests. |

TABLE 5-68: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS FOR FEMALE BATS BASED ON 1993 DATA FOR THE TRI+ CONGENERS

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|---|---|---|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 30 | 52 | 140 | 244 |
| Stillwater (168) | 52 | 244 | 245 | 1146 |
| Federal Dam (154) | 12 | 21 | 58 | 99 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.7 | 4.3 | 8.1 | 20 |
| 137.2 | 3.4 | 14 | 16 | 65 |
| 122.4 | 1.9 | 4.7 | 8.8 | 22 |
| 113.8 | 1.9 | 7.5 | 9.1 | 35 |
| 100 | 0.9 | 6.1 | 4.2 | 29 |
| 88.9 | 0.4 | 0.8 | 2.1 | 3.7 |
| 58.7 | 1.4 | 13 | 6.5 | 59 |
| 47.3 | 1.6 | 11 | 7.3 | 54 |
| 25.8 | 0.5 | 0.8 | 2.2 | 3.7 |

Bold values indicate exceedances

**TABLE 5-69: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE BAT FOR TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 30 | 45 | 140 | 211 | 12 | 13 | 55 | 63 | 5.2 | 6.6 | 24 | 31 |
| 1994 | 33 | 42 | 157 | 197 | 11 | 13 | 52 | 60 | 4.9 | 6.2 | 23 | 29 |
| 1995 | 31 | 39 | 145 | 182 | 10 | 12 | 49 | 56 | 4.5 | 5.7 | 21 | 27 |
| 1996 | 28 | 36 | 133 | 167 | 9.4 | 11 | 44 | 51 | 3.8 | 4.8 | 18 | 23 |
| 1997 | 26 | 32 | 121 | 152 | 8.5 | 10 | 40 | 46 | 3.3 | 4.2 | 15 | 20 |
| 1998 | 23 | 29 | 110 | 138 | 7.7 | 8.8 | 36 | 41 | 2.9 | 3.6 | 13 | 17 |
| 1999 | 21 | 27 | 100 | 126 | 7.0 | 8.1 | 33 | 38 | 2.5 | 3.2 | 12 | 15 |
| 2000 | 20 | 25 | 93 | 117 | 6.6 | 7.6 | 31 | 36 | 2.2 | 2.9 | 11 | 13 |
| 2001 | 19 | 24 | 88 | 110 | 6.5 | 7.4 | 30 | 35 | 2.1 | 2.7 | 10 | 13 |
| 2002 | 18 | 22 | 83 | 104 | 6.2 | 7.1 | 29 | 33 | 2.0 | 2.6 | 9.5 | 12 |
| 2003 | 16 | 20 | 76 | 96 | 5.8 | 6.7 | 27 | 31 | 1.9 | 2.4 | 8.9 | 11 |
| 2004 | 15 | 19 | 70 | 88 | 5.3 | 6.1 | 25 | 29 | 1.7 | 2.2 | 8.0 | 10 |
| 2005 | 14 | 18 | 66 | 83 | 4.9 | 5.7 | 23 | 27 | 1.5 | 2.0 | 7.3 | 9.2 |
| 2006 | 13 | 17 | 63 | 79 | 4.7 | 5.4 | 22 | 25 | 1.4 | 1.8 | 6.7 | 8.6 |
| 2007 | 13 | 16 | 59 | 74 | 4.4 | 5.0 | 20 | 24 | 1.3 | 1.7 | 6.1 | 7.8 |
| 2008 | 12 | 15 | 55 | 69 | 4.1 | 4.7 | 19 | 22 | 1.2 | 1.5 | 5.7 | 7.2 |
| 2009 | 11 | 14 | 52 | 66 | 3.9 | 4.5 | 18 | 21 | 1.2 | 1.5 | 5.5 | 7.0 |
| 2010 | 10 | 13 | 48 | 60 | 3.7 | 4.2 | 17 | 20 | 1.1 | 1.4 | 5.1 | 6.5 |
| 2011 | 9.1 | 11 | 43 | 54 | 3.4 | 3.9 | 16 | 18 | 1.0 | 1.3 | 4.7 | 6.0 |
| 2012 | 8.5 | 11 | 40 | 50 | 3.2 | 3.7 | 15 | 17 | 1.0 | 1.2 | 4.5 | 5.7 |
| 2013 | 8.0 | 10 | 37 | 47 | 3.1 | 3.5 | 14 | 17 | 0.9 | 1.1 | 4.2 | 5.3 |
| 2014 | 7.4 | 9.3 | 35 | 44 | 2.9 | 3.3 | 14 | 16 | 0.8 | 1.0 | 3.8 | 4.9 |
| 2015 | 7.0 | 8.8 | 33 | 41 | 2.8 | 3.2 | 13 | 15 | 0.8 | 1.0 | 3.7 | 4.7 |
| 2016 | 6.6 | 8.2 | 31 | 39 | 2.6 | 3.0 | 12 | 14 | 0.7 | 0.9 | 3.4 | 4.3 |
| 2017 | 6.2 | 7.8 | 29 | 36 | 2.5 | 2.9 | 12 | 13 | 0.6 | 0.8 | 3.0 | 3.9 |
| 2018 | 6.0 | 7.5 | 28 | 35 | 2.4 | 2.8 | 11 | 13 | 0.6 | 0.8 | 2.9 | 3.7 |

Bold values indicate exceedances

**TABLE 5-70: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE BAT BASED ON 1993 DATA ON A TEQ BASIS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|---|---|---|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 133 | 232 | 1328 | 2323 |
| Stillwater (168) | 232 | 1089 | 2324 | 10885 |
| Federal Dam (154) | 55 | 94 | 554 | 943 |
| <i>Lower River</i> | | | | |
| 143.5 | 8 | 20 | 78 | 197 |
| 137.2 | 15 | 62 | 153 | 624 |
| 122.4 | 8.4 | 22 | 84 | 215 |
| 113.8 | 8.7 | 34 | 87 | 339 |
| 100 | 4.0 | 28 | 40 | 276 |
| 88.9 | 2.0 | 3.6 | 20 | 36 |
| 58.7 | 6.2 | 56 | 62 | 562 |
| 47.3 | 7.0 | 51 | 70 | 512 |
| 25.8 | 2.1 | 3.6 | 21 | 36 |

Bold values indicate exceedances

**TABLE 5-71: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE BAT ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|-----------|-----------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 133 | 200 | 1328 | 2001 | 52 | 60 | 522 | 599 | 23 | 29 | 231 | 293 |
| 1994 | 149 | 187 | 1487 | 1866 | 50 | 57 | 498 | 571 | 22 | 28 | 218 | 276 |
| 1995 | 138 | 173 | 1378 | 1730 | 47 | 54 | 467 | 536 | 20 | 25 | 199 | 253 |
| 1996 | 126 | 158 | 1261 | 1582 | 42 | 48 | 420 | 483 | 17 | 22 | 170 | 215 |
| 1997 | 115 | 145 | 1153 | 1446 | 38 | 44 | 380 | 436 | 15 | 19 | 147 | 187 |
| 1998 | 104 | 131 | 1045 | 1311 | 34 | 39 | 344 | 394 | 13 | 16 | 128 | 163 |
| 1999 | 95 | 120 | 953 | 1196 | 31 | 36 | 314 | 360 | 11 | 14 | 112 | 142 |
| 2000 | 88 | 111 | 884 | 1109 | 30 | 34 | 296 | 340 | 10 | 13 | 100 | 128 |
| 2001 | 84 | 105 | 836 | 1049 | 29 | 33 | 288 | 330 | 9 | 12 | 95 | 120 |
| 2002 | 79 | 99 | 786 | 986 | 28 | 32 | 276 | 317 | 9 | 11 | 90 | 115 |
| 2003 | 73 | 91 | 726 | 911 | 26 | 30 | 260 | 298 | 8 | 11 | 84 | 107 |
| 2004 | 67 | 84 | 667 | 837 | 24 | 27 | 238 | 273 | 8 | 10 | 76 | 97 |
| 2005 | 63 | 79 | 627 | 786 | 22 | 25 | 220 | 253 | 6.9 | 8.8 | 69 | 88 |
| 2006 | 60 | 75 | 596 | 749 | 21 | 24 | 209 | 240 | 6.4 | 8.1 | 64 | 81 |
| 2007 | 56 | 70 | 558 | 700 | 19 | 22 | 195 | 224 | 5.8 | 7.4 | 58 | 74 |
| 2008 | 52 | 66 | 525 | 659 | 18 | 21 | 184 | 211 | 5.4 | 6.9 | 54 | 69 |
| 2009 | 50 | 62 | 496 | 623 | 18 | 20 | 176 | 202 | 5.2 | 6.6 | 52 | 66 |
| 2010 | 45 | 57 | 453 | 568 | 16 | 19 | 164 | 188 | 4.9 | 6.2 | 49 | 62 |
| 2011 | 41 | 51 | 407 | 511 | 15 | 17 | 151 | 173 | 4.5 | 5.7 | 45 | 57 |
| 2012 | 38 | 48 | 380 | 476 | 14 | 17 | 144 | 166 | 4.3 | 5.4 | 43 | 54 |
| 2013 | 36 | 45 | 355 | 446 | 14 | 16 | 137 | 157 | 4.0 | 5.1 | 40 | 51 |
| 2014 | 33 | 41 | 330 | 414 | 13 | 15 | 128 | 147 | 3.6 | 4.6 | 36 | 46 |
| 2015 | 31 | 39 | 312 | 392 | 12 | 14 | 124 | 143 | 3.5 | 4.4 | 35 | 44 |
| 2016 | 29 | 37 | 293 | 367 | 12 | 14 | 118 | 135 | 3.2 | 4.1 | 32 | 41 |
| 2017 | 28 | 35 | 275 | 345 | 11 | 13 | 111 | 127 | 2.9 | 3.7 | 29 | 37 |
| 2018 | 27 | 34 | 268 | 336 | 11 | 12 | 108 | 124 | 2.8 | 3.5 | 28 | 35 |

Bold values indicate exceedances

**TABLE 5-72: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE RACCOON BASED ON 1993 DATA FOR THE TRI+ CONGENERS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|--|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 5.8 | 10 | 27 | 47 |
| Stillwater (168) | 9.5 | 42 | 45 | 195 |
| Federal Dam (154) | 2.2 | 3.7 | 10 | 17 |
| <i>Lower River</i> | | | | |
| 143.5 | 0.4 | 0.8 | 1.7 | 3.7 |
| 137.2 | 0.7 | 2.6 | 3.4 | 12 |
| 122.4 | 0.4 | 0.9 | 1.8 | 4.1 |
| 113.8 | 0.4 | 1.3 | 1.9 | 6.2 |
| 100 | 0.2 | 1.3 | 0.8 | 6.1 |
| 88.9 | 0.1 | 0.3 | 0.6 | 1.2 |
| 58.7 | 0.3 | 2.2 | 1.3 | 10 |
| 47.3 | 0.3 | 2.1 | 1.6 | 9.9 |
| 25.8 | 0.1 | 0.2 | 0.6 | 1.0 |

Bold values indicate exceedances

**TABLE 5-73: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE RACCOON FOR TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|-------------|-------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 6.0 | 8.5 | 27.9 | 39.7 | 2.3 | 2.6 | 10.7 | 12.1 | 1.2 | 1.2 | 5.8 | 5.8 |
| 1994 | 6.5 | 7.9 | 30.3 | 37.0 | 2.2 | 2.5 | 10.2 | 11.5 | 1.2 | 1.2 | 5.4 | 5.4 |
| 1995 | 6.0 | 7.4 | 28.2 | 34.5 | 2.1 | 2.3 | 9.7 | 10.9 | 1.1 | 1.1 | 5.0 | 5.0 |
| 1996 | 5.5 | 6.7 | 25.7 | 31.4 | 1.8 | 2.1 | 8.6 | 9.7 | 0.9 | 0.9 | 4.2 | 4.3 |
| 1997 | 5.0 | 6.1 | 23.5 | 28.7 | 1.7 | 1.9 | 7.8 | 8.8 | 0.8 | 0.8 | 3.7 | 3.7 |
| 1998 | 4.5 | 5.5 | 21.2 | 26.0 | 1.5 | 1.7 | 7.0 | 7.9 | 0.7 | 0.7 | 3.2 | 3.2 |
| 1999 | 4.1 | 5.1 | 19.4 | 23.7 | 1.4 | 1.5 | 6.4 | 7.2 | 0.6 | 0.6 | 2.8 | 2.8 |
| 2000 | 3.8 | 4.7 | 17.9 | 21.9 | 1.3 | 1.4 | 6.0 | 6.8 | 0.5 | 0.5 | 2.5 | 2.5 |
| 2001 | 3.6 | 4.4 | 17.0 | 20.8 | 1.3 | 1.4 | 5.9 | 6.6 | 0.5 | 0.5 | 2.4 | 2.4 |
| 2002 | 3.4 | 4.2 | 15.9 | 19.5 | 1.2 | 1.3 | 5.6 | 6.3 | 0.5 | 0.5 | 2.3 | 2.3 |
| 2003 | 3.1 | 3.9 | 14.7 | 18.1 | 1.1 | 1.3 | 5.3 | 6.0 | 0.5 | 0.5 | 2.1 | 2.1 |
| 2004 | 2.9 | 3.5 | 13.5 | 16.6 | 1.0 | 1.2 | 4.9 | 5.5 | 0.4 | 0.4 | 1.9 | 1.9 |
| 2005 | 2.7 | 3.3 | 12.7 | 15.5 | 1.0 | 1.1 | 4.5 | 5.1 | 0.4 | 0.4 | 1.7 | 1.7 |
| 2006 | 2.6 | 3.2 | 12.1 | 14.8 | 0.9 | 1.0 | 4.3 | 4.8 | 0.3 | 0.3 | 1.6 | 1.6 |
| 2007 | 2.4 | 3.0 | 11.3 | 13.9 | 0.8 | 1.0 | 4.0 | 4.5 | 0.3 | 0.3 | 1.5 | 1.5 |
| 2008 | 2.3 | 2.8 | 10.6 | 13.0 | 0.8 | 0.9 | 3.7 | 4.2 | 0.3 | 0.3 | 1.4 | 1.4 |
| 2009 | 2.1 | 2.6 | 10.1 | 12.3 | 0.8 | 0.9 | 3.6 | 4.0 | 0.3 | 0.3 | 1.3 | 1.3 |
| 2010 | 2.0 | 2.4 | 9.2 | 11.3 | 0.7 | 0.8 | 3.4 | 3.8 | 0.3 | 0.3 | 1.2 | 1.2 |
| 2011 | 1.8 | 2.2 | 8.3 | 10.1 | 0.7 | 0.7 | 3.1 | 3.5 | 0.2 | 0.2 | 1.1 | 1.1 |
| 2012 | 1.6 | 2.0 | 7.7 | 9.4 | 0.6 | 0.7 | 3.0 | 3.3 | 0.2 | 0.2 | 1.1 | 1.1 |
| 2013 | 1.5 | 1.9 | 7.2 | 8.8 | 0.6 | 0.7 | 2.8 | 3.2 | 0.2 | 0.2 | 1.0 | 1.0 |
| 2014 | 1.4 | 1.8 | 6.7 | 8.2 | 0.6 | 0.6 | 2.6 | 3.0 | 0.2 | 0.2 | 0.9 | 0.9 |
| 2015 | 1.4 | 1.7 | 6.3 | 7.8 | 0.5 | 0.6 | 2.5 | 2.8 | 0.2 | 0.2 | 0.9 | 0.9 |
| 2016 | 1.3 | 1.6 | 5.9 | 7.3 | 0.5 | 0.6 | 2.4 | 2.7 | 0.2 | 0.2 | 0.8 | 0.8 |
| 2017 | 1.2 | 1.5 | 5.6 | 6.8 | 0.5 | 0.5 | 2.2 | 2.5 | 0.2 | 0.2 | 0.7 | 0.7 |
| 2018 | 1.2 | 1.4 | 5.4 | 6.6 | 0.5 | 0.5 | 2.2 | 2.5 | 0.1 | 0.1 | 0.7 | 0.7 |

Bold values indicate exceedances

**TABLE 5-74: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE RACCOON BASED ON 1993 DATA ON A TEQ BASIS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|--|--|---|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 69 | 107 | 685 | 1067 |
| Stillwater (168) | 150 | 374 | 1504 | 3736 |
| Federal Dam (154) | 19 | 33 | 195 | 328 |
| <i>Lower River</i> | | | | |
| 143.5 | 4.8 | 7.3 | 48 | 73 |
| 137.2 | 8.8 | 23 | 88 | 231 |
| 122.4 | 5.1 | 8.0 | 51 | 80 |
| 113.8 | 5.4 | 12 | 54 | 120 |
| 100 | 2.2 | 36 | 22 | 359 |
| 88.9 | 3.4 | 9.2 | 34 | 92 |
| 58.7 | 2.2 | 20 | 22 | 196 |
| 47.3 | 7.0 | 30 | 70 | 304 |
| 25.8 | 2.6 | 6.5 | 26 | 65 |

Bold values indicate exceedances

**TABLE 5-75: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE RACCOON ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|------------|------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 127 | 144 | 1270 | 1437 | 43 | 44 | 427 | 441 | 19 | 20 | 189 | 200 |
| 1994 | 122 | 133 | 1218 | 1334 | 40 | 42 | 403 | 416 | 17 | 18 | 174 | 184 |
| 1995 | 114 | 125 | 1144 | 1252 | 39 | 40 | 388 | 400 | 17 | 18 | 168 | 177 |
| 1996 | 104 | 114 | 1039 | 1137 | 35 | 36 | 347 | 359 | 14 | 15 | 141 | 150 |
| 1997 | 96 | 105 | 957 | 1047 | 32 | 33 | 316 | 326 | 12 | 13 | 123 | 130 |
| 1998 | 86 | 95 | 864 | 945 | 28 | 29 | 283 | 293 | 11 | 11 | 106 | 113 |
| 1999 | 79 | 86 | 788 | 862 | 26 | 27 | 259 | 267 | 9 | 10 | 94 | 99 |
| 2000 | 72 | 79 | 722 | 790 | 24 | 25 | 238 | 246 | 8.2 | 8.7 | 82 | 87 |
| 2001 | 68 | 75 | 683 | 748 | 23 | 24 | 233 | 240 | 7.6 | 8.1 | 76 | 81 |
| 2002 | 65 | 71 | 645 | 706 | 22 | 23 | 224 | 231 | 7.3 | 7.7 | 73 | 77 |
| 2003 | 60 | 66 | 600 | 656 | 21 | 22 | 213 | 220 | 6.9 | 7.3 | 69 | 73 |
| 2004 | 55 | 60 | 550 | 601 | 20 | 20 | 197 | 204 | 6.3 | 6.7 | 63 | 67 |
| 2005 | 51 | 56 | 509 | 557 | 18 | 18 | 178 | 184 | 5.6 | 5.9 | 56 | 59 |
| 2006 | 49 | 53 | 488 | 535 | 17 | 18 | 171 | 177 | 5.3 | 5.6 | 53 | 56 |
| 2007 | 46 | 50 | 458 | 501 | 16 | 16 | 159 | 164 | 4.8 | 5.1 | 48 | 51 |
| 2008 | 43 | 47 | 428 | 468 | 15 | 15 | 149 | 154 | 4.4 | 4.6 | 44 | 46 |
| 2009 | 41 | 44 | 407 | 445 | 14 | 15 | 142 | 147 | 4.2 | 4.7 | 42 | 47 |
| 2010 | 38 | 41 | 379 | 414 | 14 | 14 | 136 | 140 | 4.1 | 4.3 | 41 | 43 |
| 2011 | 34 | 37 | 336 | 367 | 12 | 13 | 123 | 127 | 3.7 | 3.9 | 37 | 39 |
| 2012 | 31 | 34 | 310 | 339 | 12 | 12 | 116 | 120 | 3.4 | 3.6 | 34 | 36 |
| 2013 | 29 | 32 | 293 | 321 | 11 | 12 | 113 | 116 | 3.3 | 3.5 | 33 | 35 |
| 2014 | 27 | 30 | 270 | 296 | 10 | 11 | 104 | 108 | 3.0 | 3.1 | 30 | 31 |
| 2015 | 25 | 28 | 254 | 279 | 10 | 10 | 100 | 103 | 2.8 | 3.0 | 28 | 30 |
| 2016 | 24 | 27 | 242 | 265 | 10 | 10 | 97 | 100 | 2.7 | 2.9 | 27 | 29 |
| 2017 | 22 | 24 | 223 | 244 | 8.9 | 9.2 | 89 | 92 | 2.3 | 2.5 | 23 | 25 |
| 2018 | 21 | 23 | 211 | 232 | 8.5 | 8.7 | 85 | 87 | 2.2 | 2.3 | 22 | 23 |

Bold values indicate exceedances

TABLE 5-76: RATIO OF OBSERVED MINK AND OTTER PCB CONCENTRATIONS TO BENCHMARKS

| Comparison to Low Range LOAEL | | | | |
|-------------------------------|---------------------|---------------------|---------------|----------------|
| Species and Statistic | North Hudson Valley | South Hudson Valley | Hudson Valley | Other NY State |
| Mink liver - average | 0.5 | 0.6 | | |
| Mink liver - minimum | 0.1 | 0.1 | | |
| Mink liver - maximum | 1.4 | 2.8 | | |
| | | | | |
| Otter liver - average | | | 1.9 | |
| Otter liver - minimum | | | 0.6 | |
| Otter liver - maximum | | | 5.9 | |

| Comparison to Upper Range LOAEL | | | | |
|---------------------------------|---------------------|---------------------|---------------|----------------|
| Species and Statistic | North Hudson Valley | South Hudson Valley | Hudson Valley | Other NY State |
| Mink liver - average | 0.2 | 0.2 | | |
| Mink liver - minimum | 0.0 | 0.0 | | |
| Mink liver - maximum | 0.5 | 1.1 | | |
| | | | | |
| Otter liver - average | | | 0.7 | |
| Otter liver - minimum | | | 0.2 | |
| Otter liver - maximum | | | 2.4 | |

**TABLE 5-77: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE MINK BASED ON 1993 DATA FOR THE TRI+ CONGENERS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|---|---|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 11 | 17 | 359 | 566 |
| Stillwater (168) | 5.8 | 23 | 188 | 760 |
| Federal Dam (154) | 1.8 | 2.8 | 58 | 92 |
| <i>Lower River</i> | | | | |
| 143.5 | 1.0 | 1.3 | 31 | 43 |
| 137.2 | 1.9 | 4.6 | 62 | 150 |
| 122.4 | 0.8 | 1.4 | 25 | 45 |
| 113.8 | 0.8 | 1.3 | 27 | 43 |
| 100 | 0.4 | 1.0 | 12 | 34 |
| 88.9 | 0.6 | 0.8 | 19 | 26 |
| 58.7 | 0.7 | 1.8 | 24 | 58 |
| 47.3 | 0.7 | 1.7 | 22 | 56 |
| 25.8 | 0.5 | 0.6 | 15 | 18 |

Bold values indicate exceedances

TABLE 5-78: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS FOR FEMALE OTTER BASED ON 1993 DATA FOR THE TRI+ CONGENERs

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|--|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 89 | 173 | 2906 | 5623 |
| Stillwater (168) | 16 | 21 | 520 | 671 |
| Federal Dam (154) | 12 | 21 | 375 | 673 |
| <i>Lower River</i> | | | | |
| 143.5 | 12 | 21 | 375 | 673 |
| 137.2 | 43 | 103 | 1413 | 3362 |
| 122.4 | 10 | 14 | 329 | 453 |
| 113.8 | 9.2 | 13 | 298 | 417 |
| 100 | 11 | 33 | 342 | 1057 |
| 88.9 | 6.8 | 13 | 220 | 419 |
| 58.7 | 7.9 | 12 | 256 | 379 |
| 47.3 | 9.0 | 24 | 293 | 789 |
| 25.8 | 6.4 | 13 | 207 | 411 |

Bold values indicate exceedances

**TABLE 5-79: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE MINK FOR TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|------------|-----------|-----------|---------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 5.6 | 7.1 | 181 | 231 | 2.4 | 2.6 | 79 | 86 | 0.9 | 1.0 | 29 | 34 |
| 1994 | 5.7 | 6.6 | 185 | 216 | 2.3 | 2.5 | 75 | 82 | 0.9 | 1.0 | 28 | 32 |
| 1995 | 5.6 | 6.5 | 182 | 212 | 2.2 | 2.4 | 72 | 79 | 0.8 | 1.0 | 27 | 32 |
| 1996 | 4.8 | 5.6 | 156 | 182 | 1.8 | 1.9 | 58 | 63 | 0.7 | 0.8 | 21 | 25 |
| 1997 | 4.3 | 5.0 | 139 | 163 | 1.8 | 1.9 | 57 | 62 | 0.6 | 0.7 | 19 | 22 |
| 1998 | 3.8 | 4.5 | 123 | 145 | 1.4 | 1.5 | 44 | 49 | 0.5 | 0.6 | 16 | 19 |
| 1999 | 3.5 | 4.1 | 113 | 133 | 1.3 | 1.4 | 42 | 46 | 0.4 | 0.5 | 15 | 17 |
| 2000 | 3.2 | 3.8 | 104 | 123 | 1.2 | 1.3 | 38 | 42 | 0.4 | 0.5 | 14 | 16 |
| 2001 | 3.1 | 3.6 | 100 | 117 | 1.2 | 1.3 | 40 | 43 | 0.4 | 0.4 | 12 | 15 |
| 2002 | 2.8 | 3.4 | 93 | 109 | 1.1 | 1.2 | 36 | 40 | 0.4 | 0.4 | 12 | 14 |
| 2003 | 2.6 | 3.1 | 86 | 101 | 1.1 | 1.2 | 34 | 38 | 0.4 | 0.4 | 11 | 13 |
| 2004 | 2.4 | 2.9 | 79 | 93 | 1.0 | 1.1 | 32 | 35 | 0.3 | 0.4 | 10 | 12 |
| 2005 | 2.3 | 2.7 | 74 | 87 | 0.9 | 1.0 | 29 | 32 | 0.3 | 0.3 | 9.3 | 11 |
| 2006 | 2.2 | 2.6 | 71 | 83 | 0.9 | 1.0 | 28 | 31 | 0.3 | 0.3 | 8.6 | 10 |
| 2007 | 2.0 | 2.4 | 66 | 78 | 0.8 | 0.9 | 25 | 28 | 0.3 | 0.3 | 8.2 | 9.5 |
| 2008 | 1.9 | 2.2 | 62 | 73 | 0.7 | 0.8 | 23 | 26 | 0.2 | 0.3 | 7.2 | 8.3 |
| 2009 | 1.8 | 2.1 | 59 | 69 | 0.7 | 0.8 | 24 | 26 | 0.2 | 0.3 | 7.5 | 8.6 |
| 2010 | 1.7 | 1.9 | 54 | 63 | 0.7 | 0.7 | 22 | 24 | 0.2 | 0.3 | 7.2 | 8.2 |
| 2011 | 1.5 | 1.7 | 48 | 57 | 0.6 | 0.7 | 19 | 21 | 0.2 | 0.2 | 6.5 | 7.4 |
| 2012 | 1.4 | 1.6 | 45 | 53 | 0.6 | 0.7 | 20 | 22 | 0.2 | 0.2 | 6.8 | 7.7 |
| 2013 | 1.3 | 1.5 | 42 | 50 | 0.6 | 0.6 | 18 | 20 | 0.2 | 0.2 | 5.9 | 6.8 |
| 2014 | 1.2 | 1.4 | 39 | 46 | 0.5 | 0.6 | 18 | 19 | 0.2 | 0.2 | 5.5 | 6.3 |
| 2015 | 1.1 | 1.3 | 37 | 44 | 0.5 | 0.6 | 17 | 18 | 0.2 | 0.2 | 5.4 | 6.1 |
| 2016 | 1.1 | 1.3 | 35 | 41 | 0.5 | 0.6 | 16 | 18 | 0.2 | 0.2 | 5.3 | 6.0 |
| 2017 | 1.0 | 1.2 | 33 | 38 | 0.4 | 0.5 | 14 | 16 | 0.1 | 0.2 | 4.3 | 4.9 |
| 2018 | 1.0 | 1.2 | 32 | 37 | 0.4 | 0.5 | 14 | 16 | 0.1 | 0.2 | 4.4 | 5.0 |

Bold values indicate exceedances

**TABLE 5-80: RATIO OF MODELED DIETARY DOSE TO TOXICITY BENCHMARKS
FOR FEMALE OTTER FOR TRI+ CONGENERS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|------------|------------|-------------|-------------|------------|------------|------------|------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 50 | 62 | 1615 | 2025 | 37 | 46 | 1217 | 1485 | 18 | 23 | 598 | 746 |
| 1994 | 30 | 41 | 983 | 1322 | 26 | 31 | 837 | 1008 | 13 | 16 | 424 | 523 |
| 1995 | 33 | 45 | 1083 | 1448 | 29 | 34 | 929 | 1116 | 14 | 17 | 459 | 557 |
| 1996 | 30 | 44 | 988 | 1423 | 23 | 28 | 747 | 911 | 12 | 16 | 402 | 508 |
| 1997 | 24 | 34 | 769 | 1108 | 20 | 24 | 654 | 788 | 11 | 13 | 347 | 434 |
| 1998 | 20 | 28 | 643 | 913 | 18 | 22 | 588 | 715 | 9.0 | 11 | 292 | 358 |
| 1999 | 17 | 25 | 559 | 828 | 15 | 18 | 480 | 594 | 6.8 | 8.6 | 222 | 280 |
| 2000 | 16 | 24 | 522 | 775 | 14 | 17 | 453 | 554 | 6.5 | 8.0 | 212 | 261 |
| 2001 | 15 | 22 | 488 | 720 | 13 | 16 | 436 | 536 | 6.3 | 7.8 | 203 | 253 |
| 2002 | 15 | 22 | 481 | 707 | 14 | 18 | 465 | 569 | 6.7 | 8.3 | 216 | 271 |
| 2003 | 13 | 19 | 420 | 625 | 12 | 15 | 391 | 482 | 5.7 | 7.1 | 184 | 232 |
| 2004 | 12 | 18 | 393 | 578 | 12 | 14 | 381 | 466 | 5.9 | 7.3 | 190 | 239 |
| 2005 | 11 | 16 | 349 | 526 | 8.9 | 11 | 291 | 355 | 4.4 | 5.5 | 143 | 180 |
| 2006 | 11 | 16 | 344 | 508 | 10 | 12 | 317 | 388 | 4.7 | 5.7 | 151 | 186 |
| 2007 | 10 | 15 | 317 | 477 | 8.1 | 10 | 265 | 322 | 3.6 | 4.4 | 117 | 144 |
| 2008 | 9.2 | 14 | 299 | 455 | 7.4 | 9.1 | 239 | 294 | 3.5 | 4.4 | 114 | 144 |
| 2009 | 9.3 | 14 | 302 | 442 | 7.9 | 10 | 258 | 316 | 3.8 | 4.7 | 122 | 154 |
| 2010 | 8.2 | 12 | 266 | 388 | 7.2 | 8.9 | 235 | 289 | 3.1 | 3.9 | 102 | 127 |
| 2011 | 7.2 | 11 | 233 | 354 | 6.2 | 7.6 | 201 | 247 | 3.1 | 3.9 | 99 | 126 |
| 2012 | 7.1 | 11 | 231 | 341 | 6.3 | 7.7 | 203 | 250 | 3.0 | 3.7 | 96 | 122 |
| 2013 | 6.8 | 10 | 222 | 328 | 6.5 | 8.0 | 211 | 260 | 2.9 | 3.6 | 93 | 118 |
| 2014 | 6.2 | 9.0 | 202 | 293 | 6.2 | 7.6 | 201 | 247 | 2.6 | 3.2 | 85 | 105 |
| 2015 | 5.7 | 8.4 | 185 | 271 | 5.4 | 6.6 | 177 | 215 | 2.4 | 3.0 | 78 | 99 |
| 2016 | 5.8 | 8.4 | 189 | 273 | 6.1 | 7.5 | 198 | 242 | 2.6 | 3.2 | 84 | 105 |
| 2017 | 5.0 | 7.6 | 163 | 246 | 4.6 | 5.7 | 150 | 185 | 2.1 | 2.8 | 69 | 91 |
| 2018 | 4.8 | 7.2 | 158 | 235 | 4.4 | 5.5 | 144 | 178 | 1.9 | 2.5 | 63 | 81 |

Bold values indicate exceedances

**TABLE 5-81: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE MINK BASED ON 1993 DATA ON A TEQ BASIS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|--|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 28 | 44 | 792 | 1233 |
| Stillwater (168) | 18 | 55 | 510 | 1536 |
| Federal Dam (154) | 4.3 | 6.8 | 120 | 191 |
| <i>Lower River</i> | | | | |
| 143.5 | 2.5 | 3.4 | 69 | 96 |
| 137.2 | 4.9 | 12 | 137 | 322 |
| 122.4 | 2.0 | 3.5 | 57 | 97 |
| 113.8 | 2.2 | 3.3 | 61 | 93 |
| 100 | 0.9 | 4.3 | 26 | 121 |
| 88.9 | 1.6 | 2.5 | 46 | 71 |
| 58.7 | 1.8 | 4.3 | 50 | 120 |
| 47.3 | 2.0 | 5.0 | 55 | 139 |
| 25.8 | 1.2 | 1.7 | 34 | 49 |

Bold values indicate exceedances

**TABLE 5-82: RATIO OF MODELED DIETARY DOSES TO BENCHMARKS
FOR FEMALE OTTER BASED ON 1993 DATA ON A TEQ BASIS**

| Location | LOAEL vs. Average ADD Hazard Quotient | LOAEL vs. 95% UCL ADD Hazard Quotient | NOAEL vs. Average ADD Hazard Quotient | NOAEL vs. 95% UCL ADD Hazard Quotient |
|----------------------------|---|---|--|--|
| <i>Upper River</i> | | | | |
| Thompson Island Pool (189) | 225 | 434 | 6286 | 12140 |
| Stillwater (168) | 45 | 60 | 1254 | 1683 |
| Federal Dam (154) | 29 | 52 | 817 | 1467 |
| <i>Lower River</i> | | | | |
| 143.5 | 29 | 52 | 808 | 1453 |
| 137.2 | 108 | 258 | 3038 | 7230 |
| 122.4 | 25 | 35 | 711 | 978 |
| 113.8 | 23 | 32 | 644 | 904 |
| 100 | 26 | 82 | 735 | 2309 |
| 88.9 | 17 | 32 | 476 | 910 |
| 58.7 | 20 | 30 | 550 | 827 |
| 47.3 | 23 | 61 | 635 | 1720 |
| 25.8 | 16 | 32 | 447 | 890 |

Bold values indicate exceedances

**TABLE 5-83: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE MINK ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|-----------|-----------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL |
| 1993 | 19 | 22 | 522 | 612 | 7.5 | 7.9 | 209 | 221 | 2.9 | 3.1 | 80 | 88 |
| 1994 | 18 | 20 | 510 | 569 | 7.0 | 7.5 | 197 | 209 | 2.7 | 3.0 | 76 | 83 |
| 1995 | 18 | 20 | 498 | 556 | 6.9 | 7.3 | 192 | 203 | 2.7 | 2.9 | 75 | 82 |
| 1996 | 15 | 17 | 431 | 482 | 5.6 | 5.9 | 156 | 166 | 2.1 | 2.3 | 60 | 66 |
| 1997 | 14 | 16 | 388 | 435 | 5.4 | 5.7 | 152 | 161 | 1.9 | 2.1 | 53 | 58 |
| 1998 | 12 | 14 | 345 | 388 | 4.4 | 4.7 | 122 | 130 | 1.6 | 1.8 | 45 | 50 |
| 1999 | 11 | 13 | 317 | 356 | 4.1 | 4.4 | 115 | 122 | 1.4 | 1.6 | 40 | 44 |
| 2000 | 10 | 12 | 290 | 327 | 3.7 | 3.9 | 103 | 109 | 1.3 | 1.4 | 37 | 40 |
| 2001 | 10 | 11 | 277 | 311 | 3.8 | 4.0 | 106 | 113 | 1.2 | 1.3 | 34 | 37 |
| 2002 | 9.2 | 10 | 259 | 291 | 3.5 | 3.7 | 99 | 105 | 1.1 | 1.3 | 32 | 35 |
| 2003 | 8.6 | 10 | 240 | 270 | 3.3 | 3.5 | 93 | 99 | 1.1 | 1.2 | 31 | 34 |
| 2004 | 7.9 | 8.8 | 220 | 248 | 3.1 | 3.3 | 87 | 92 | 1.0 | 1.1 | 28 | 30 |
| 2005 | 7.4 | 8.3 | 206 | 232 | 2.8 | 3.0 | 78 | 83 | 0.9 | 1.0 | 25 | 28 |
| 2006 | 7.0 | 7.9 | 197 | 221 | 2.7 | 2.9 | 76 | 81 | 0.8 | 0.9 | 24 | 26 |
| 2007 | 6.6 | 7.4 | 184 | 207 | 2.5 | 2.6 | 69 | 74 | 0.8 | 0.9 | 22 | 24 |
| 2008 | 6.2 | 6.9 | 173 | 194 | 2.3 | 2.4 | 64 | 68 | 0.7 | 0.8 | 19 | 21 |
| 2009 | 5.9 | 6.6 | 165 | 185 | 2.3 | 2.4 | 64 | 68 | 0.7 | 0.8 | 20 | 22 |
| 2010 | 5.4 | 6.1 | 151 | 170 | 2.1 | 2.3 | 60 | 63 | 0.7 | 0.7 | 19 | 21 |
| 2011 | 4.8 | 5.4 | 135 | 151 | 1.9 | 2.0 | 52 | 56 | 0.6 | 0.7 | 17 | 19 |
| 2012 | 4.5 | 5.0 | 126 | 141 | 1.9 | 2.1 | 54 | 57 | 0.6 | 0.7 | 18 | 19 |
| 2013 | 4.2 | 4.7 | 118 | 133 | 1.8 | 1.9 | 50 | 53 | 0.6 | 0.6 | 16 | 17 |
| 2014 | 3.9 | 4.4 | 110 | 123 | 1.7 | 1.8 | 48 | 50 | 0.5 | 0.6 | 15 | 16 |
| 2015 | 3.7 | 4.1 | 103 | 116 | 1.6 | 1.7 | 45 | 48 | 0.5 | 0.5 | 14 | 15 |
| 2016 | 3.5 | 3.9 | 98 | 110 | 1.6 | 1.7 | 44 | 47 | 0.5 | 0.5 | 14 | 15 |
| 2017 | 3.2 | 3.6 | 91 | 102 | 1.4 | 1.5 | 39 | 41 | 0.4 | 0.4 | 11 | 12 |
| 2018 | 3.1 | 3.5 | 87 | 98 | 1.4 | 1.5 | 38 | 41 | 0.4 | 0.4 | 11 | 12 |

Bold values indicate exceedances

**TABLE 5-84: RATIO OF MODELED DIETARY DOSES TO TOXICITY BENCHMARKS
FOR FEMALE OTTER ON A TEQ BASIS FOR THE PERIOD 1993 - 2018**

| Year | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL | LOAEL | LOAEL | NOAEL | NOAEL |
|------|------------|------------|-------------|-------------|-----------|------------|-------------|-------------|------------|------------|-------------|-------------|
| | 189 | 189 | 189 | 189 | 168 | 168 | 168 | 168 | 154 | 154 | 154 | 154 |
| | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL | Average | 95% UCL |
| 1993 | 128 | 160 | 3591 | 4479 | 95 | 115 | 2653 | 3227 | 47 | 58 | 1302 | 1619 |
| 1994 | 80 | 106 | 2227 | 2961 | 66 | 79 | 1834 | 2200 | 33 | 41 | 927 | 1138 |
| 1995 | 87 | 115 | 2434 | 3223 | 73 | 87 | 2031 | 2432 | 36 | 43 | 1000 | 1212 |
| 1996 | 79 | 113 | 2220 | 3158 | 58 | 71 | 1636 | 1988 | 31 | 39 | 877 | 1103 |
| 1997 | 62 | 88 | 1744 | 2476 | 51 | 61 | 1434 | 1721 | 27 | 34 | 756 | 944 |
| 1998 | 52 | 73 | 1463 | 2047 | 46 | 56 | 1290 | 1562 | 23 | 28 | 636 | 778 |
| 1999 | 46 | 66 | 1276 | 1857 | 38 | 46 | 1055 | 1299 | 17 | 22 | 486 | 610 |
| 2000 | 43 | 62 | 1190 | 1737 | 36 | 43 | 996 | 1211 | 17 | 20 | 462 | 569 |
| 2001 | 40 | 58 | 1113 | 1614 | 34 | 42 | 959 | 1172 | 16 | 20 | 444 | 550 |
| 2002 | 39 | 57 | 1095 | 1584 | 36 | 44 | 1019 | 1244 | 17 | 21 | 471 | 588 |
| 2003 | 34 | 50 | 960 | 1403 | 31 | 38 | 859 | 1055 | 14 | 18 | 402 | 504 |
| 2004 | 32 | 46 | 898 | 1297 | 30 | 36 | 836 | 1020 | 15 | 19 | 414 | 518 |
| 2005 | 29 | 42 | 799 | 1180 | 23 | 28 | 641 | 780 | 11 | 14 | 313 | 391 |
| 2006 | 28 | 41 | 785 | 1140 | 25 | 30 | 697 | 849 | 12 | 14 | 330 | 404 |
| 2007 | 26 | 38 | 724 | 1071 | 21 | 25 | 584 | 706 | 9.1 | 11 | 256 | 314 |
| 2008 | 24 | 36 | 683 | 1020 | 19 | 23 | 527 | 646 | 8.9 | 11 | 248 | 313 |
| 2009 | 25 | 35 | 688 | 990 | 20 | 25 | 568 | 691 | 10 | 12 | 266 | 335 |
| 2010 | 22 | 31 | 607 | 871 | 19 | 23 | 518 | 633 | 8.0 | 10 | 223 | 276 |
| 2011 | 19 | 28 | 532 | 793 | 16 | 19 | 443 | 541 | 7.7 | 10 | 216 | 274 |
| 2012 | 19 | 27 | 526 | 764 | 16 | 20 | 448 | 548 | 7.5 | 9.4 | 210 | 264 |
| 2013 | 18 | 26 | 505 | 735 | 17 | 20 | 465 | 569 | 7.2 | 9.1 | 203 | 256 |
| 2014 | 16 | 23 | 460 | 657 | 16 | 19 | 442 | 540 | 6.6 | 8.2 | 185 | 229 |
| 2015 | 15 | 22 | 422 | 608 | 14 | 17 | 389 | 472 | 6.1 | 7.7 | 171 | 215 |
| 2016 | 15 | 22 | 429 | 611 | 16 | 19 | 435 | 529 | 6.6 | 8.1 | 184 | 227 |
| 2017 | 13 | 20 | 370 | 551 | 12 | 14 | 331 | 405 | 5.4 | 7.1 | 151 | 198 |
| 2018 | 13 | 19 | 358 | 526 | 11 | 14 | 318 | 389 | 4.9 | 6.3 | 137 | 175 |

Bold values indicate exceedances

TABLE 5-85: WILDLIFE SURVEY RESULTS Mammals

Hudson River
New York

| Information Source | Date | Contact | Response | Contact Information | Data Available | Information/Findings |
|--|---------------------|-----------------------|----------------------------|---|--|--|
| Mammals | | | | | | |
| Hudsonia | 2-Jun-99 | Call/Fax | YES; spoke with on 6/2/999 | Eric Kiviat, Executive Director; (914) 758-7273 (7274) OR (914) 758-7053; FAX: (914) 758-7033; EMAIL: kiviat@bard.edu; inside.bard.edu/specialprog/arch/hudsonia.html | He has no direct knowledge of the upper Hudson but provided names | RIVER OTTER: very rare; he has only seen one on the Hudson in 30 years RACCOON: Fur bearer unit - NYSDEC; trapper prices currently very low so may not have information LITTLE BROWN BAT: Endangered species Unit - Allen Hicks (Delmar NYSDEC Endangered Species) NYSDEC - Natural Resource Damage Assessment |
| NYS Department of Environmental Conservation - Endangered Species Unit | 3-Jun-99 | Call | Yes | Al Hicks (Mammal Biologist) (518) 478-3056; www.dec.state.ny.us | Left Message - Will call back | |
| The New York River Otter Project | 2-Jun-99 | Email | No | Dennis Money, Dennis_Money@rge.com; www.nyotter.org | Left Message - Will call back | |
| Professional Trapper | 4-Jun-99 | Call | No | Jim Comstock | Left Message - Will call back | |
| New York State Trappers Association | 4-Jun-99 | Email | Yes | Jerry Leggieir (montcalm@earthlink.net) | Asked me to give him a call at night; also suggested that I call Everett Nack (518) 851-2901 - a commercial fisherman on the river | |
| Professional Fisherman on the Hudson | | | | Everett Nack (518) 851-2901 | Recommended by Jerry Leggieir | |
| NYSDEC | 16-Jun-99 | Call | Yes | Mark Brown (518) 623-3671 | Familiar with the area regarding mammals, birds, and herps. Good source. See General Info page. | Otter, Mink, Musk Rat present. PCB contamination reduced their numbers severly but in the past 10 years, they have rebounded after clean-up work. Has also seen raccoon, short and long tail weasels, big and little brown bat, skunk, opossum. The red fox, grey fox, and coyote especially common in the northern Hudson, and plenty of white tail deer suggesting no bears. |
| Ndakinna Wilderness Project | 6/3/1999 6/16/99 | Email Call Call | No No Yes | Jim Brushek (518) 583-9980x3, 23 Middle Grove Road, Greenfield Center, NY 12833; Received address from Saratoga County Information - Annamaria Dalton (annamaria@spa.net) | Professional Tracker | Quite a few otter. Mink numbers are large and increasing. Tons of raccoons ("road-kill count is staggering"). Some musk rat. Lots of beavers. Very recent reports of moose in the center of Saratoga, about 5 miles from Hudson. He expects moose to inhabit the Hudson very soon but he thinks they are already there. Sees fisher cats cruising the water occassionally. Frequently sees red fox, grey fox, and deer visiting the water. The coyote population is very large. Coyotes and foxes will feed on the smaller aquatic mammals. Sees the occassional black bear. |

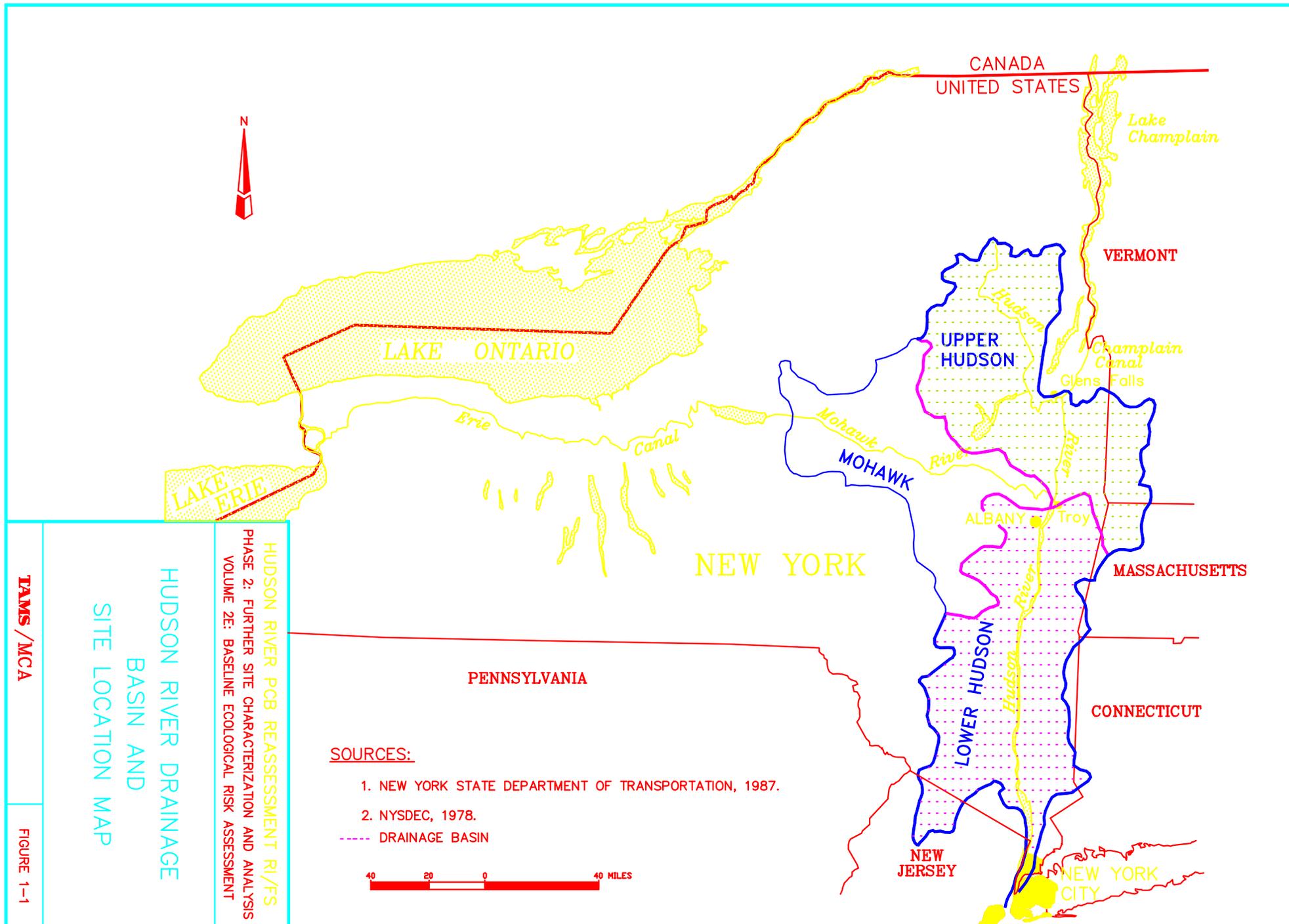
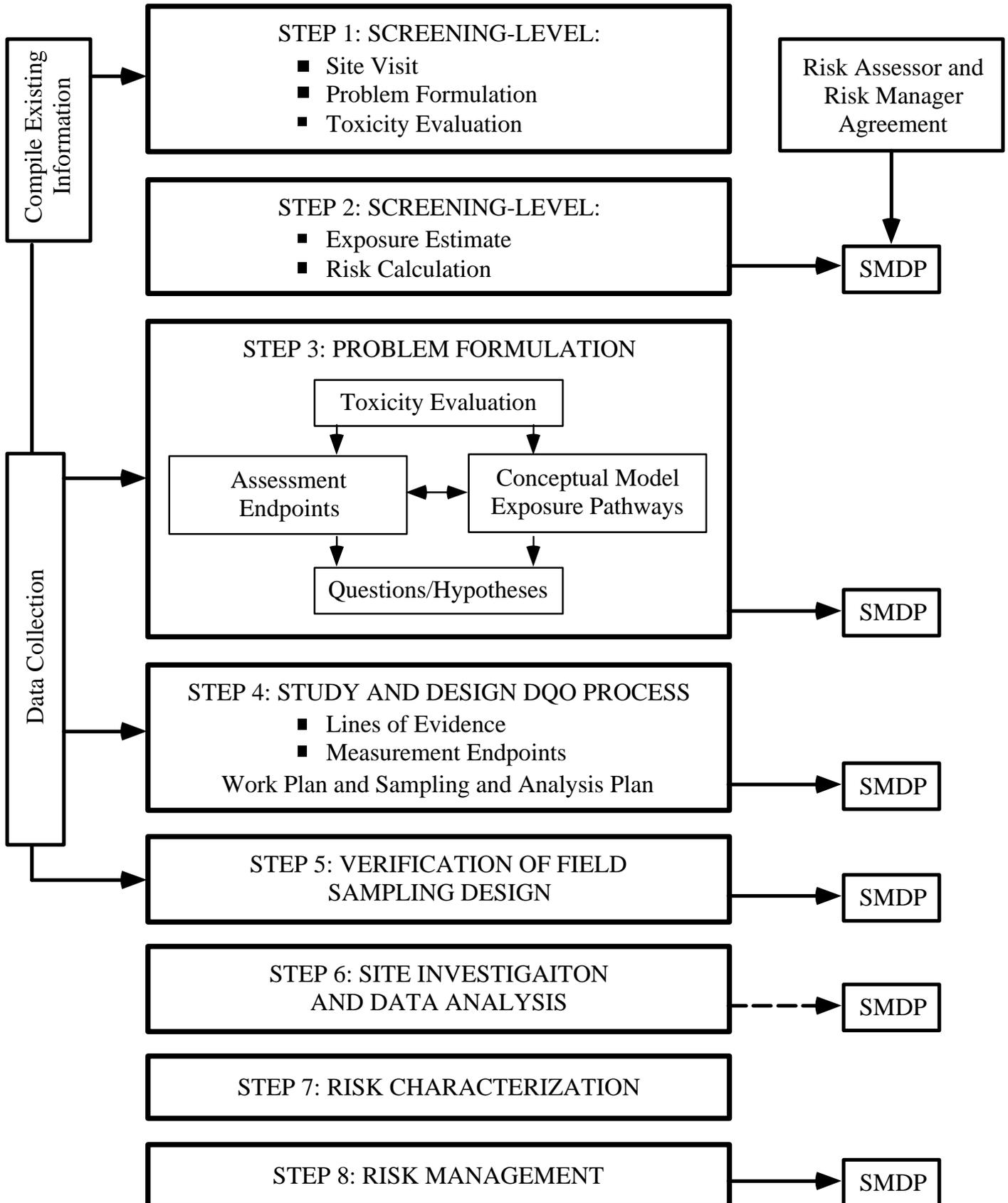
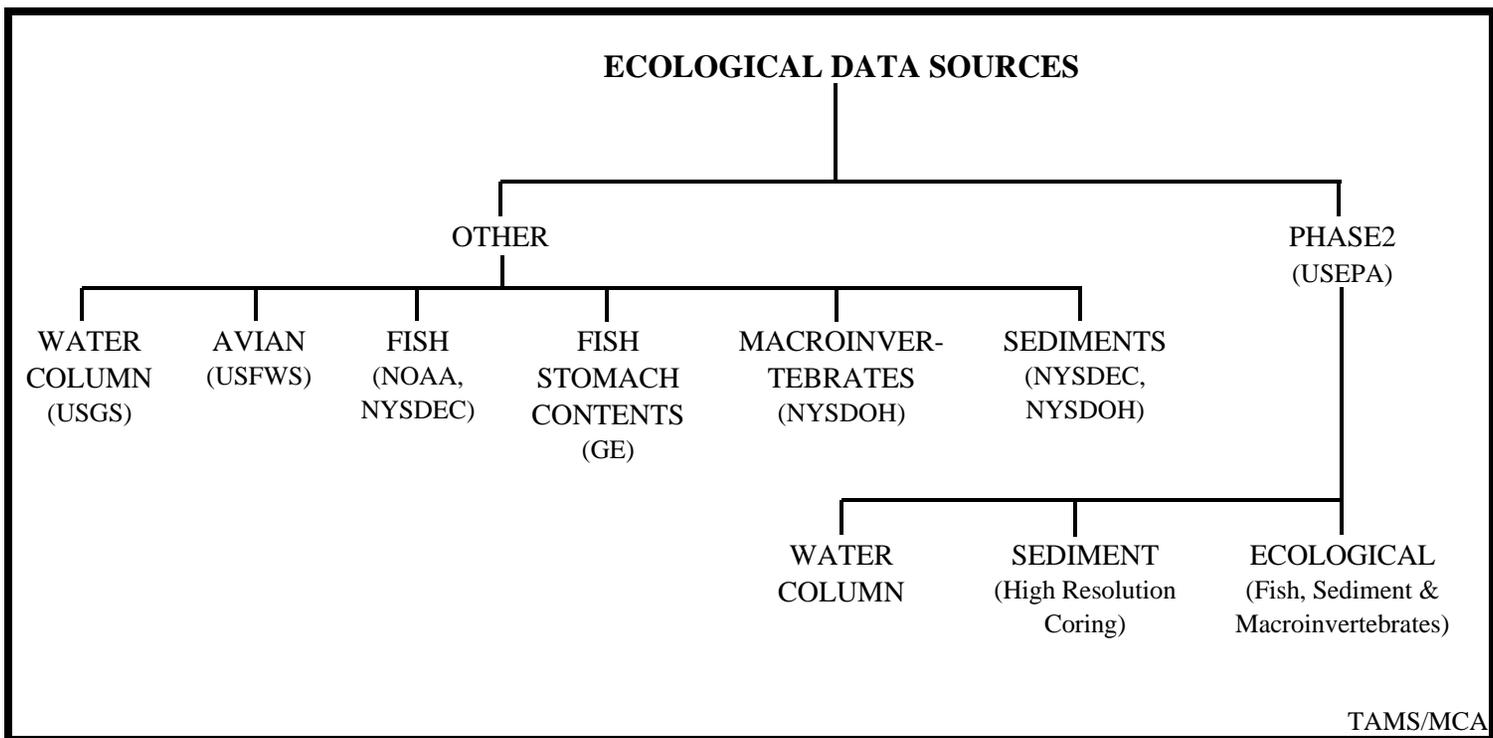


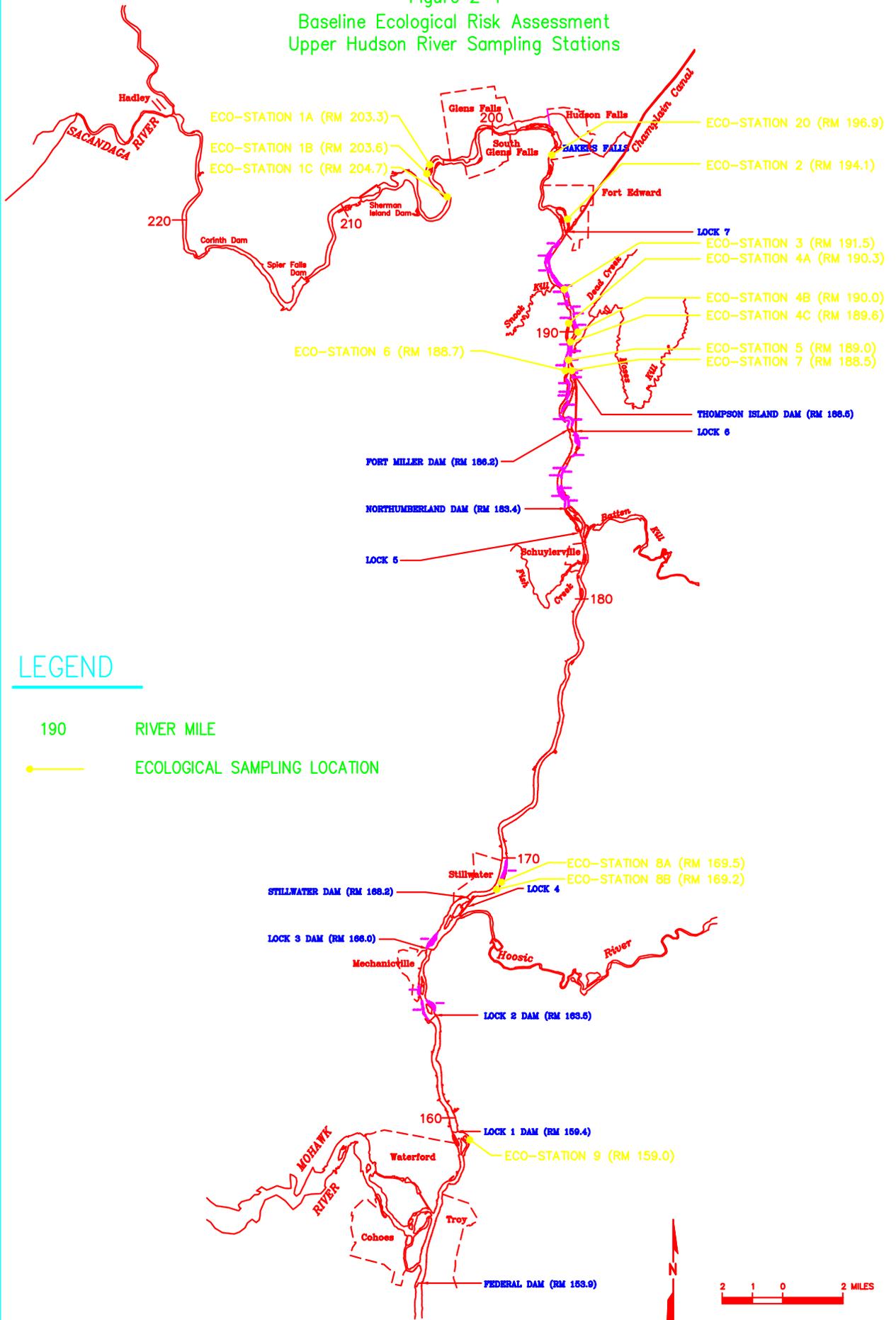
Figure 1-2
Eight-Step Ecological Risk Assessment Process for Superfund
Hudson River PCB Reassessment
Ecological Risk Assessment





**Figure 1-3
Hudson River ERA Data Sources**

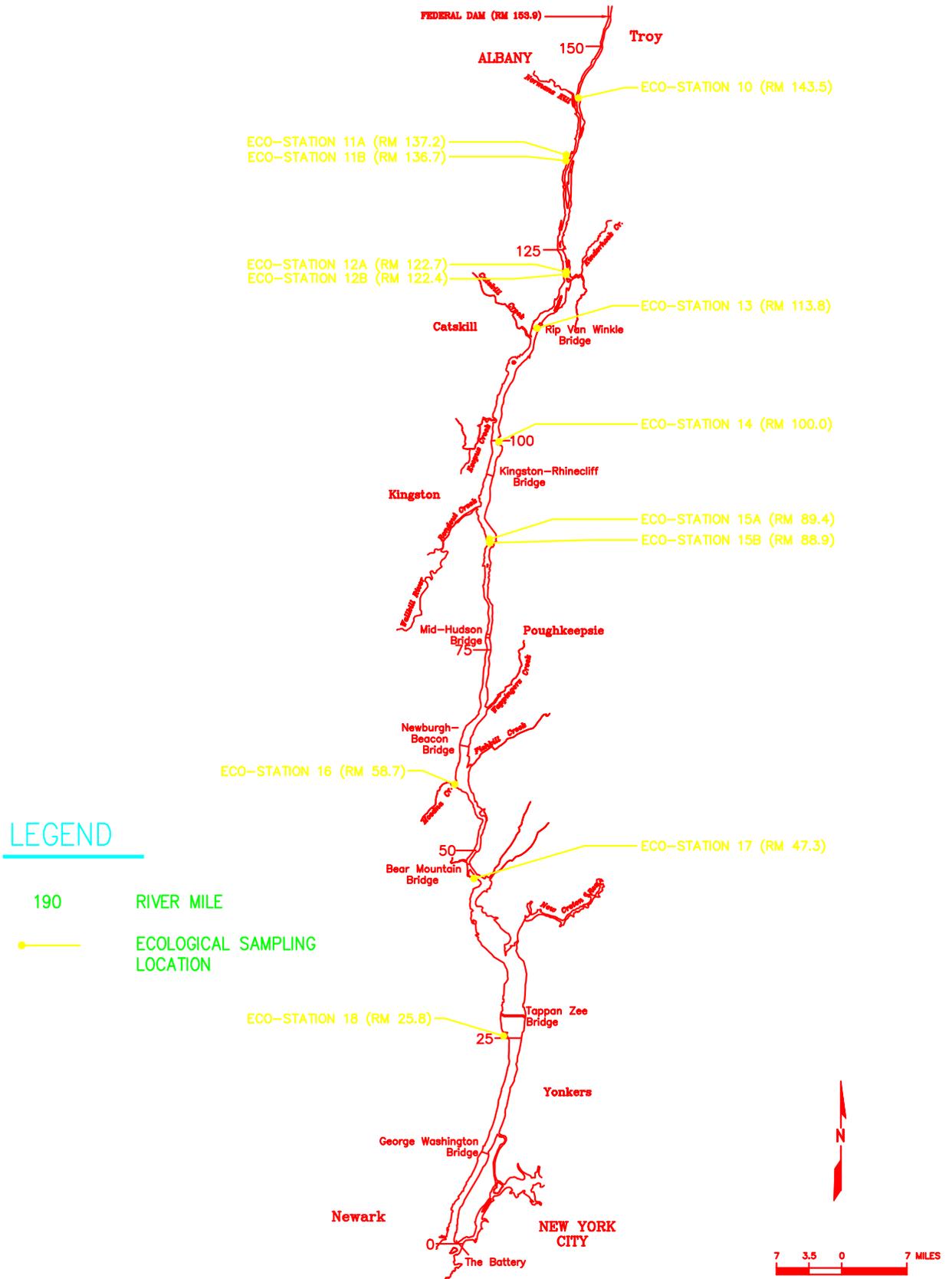
Figure 2-1
Baseline Ecological Risk Assessment
Upper Hudson River Sampling Stations

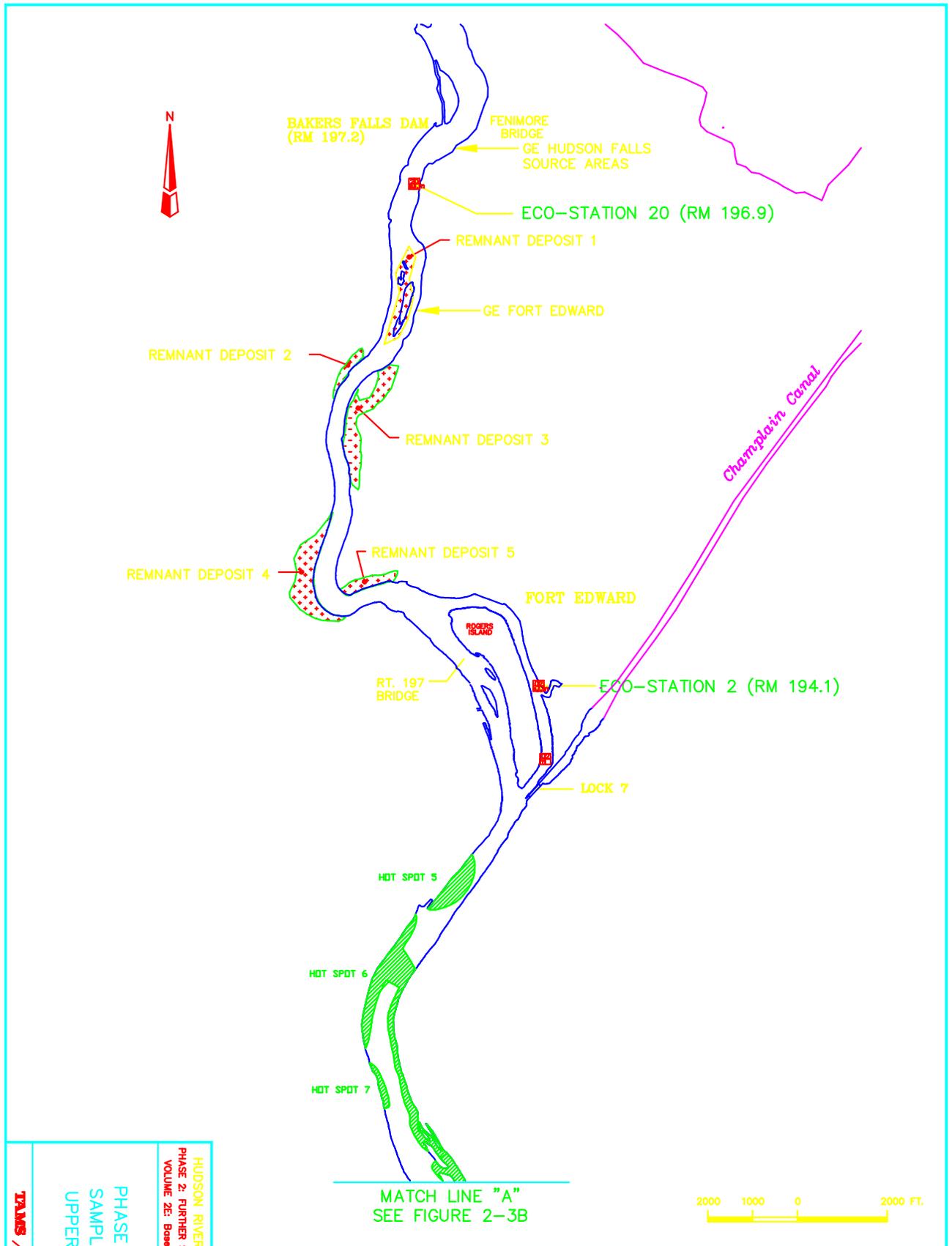


LEGEND

- 190 RIVER MILE
- ECOLOGICAL SAMPLING LOCATION

Figure 2-2
Baseline Ecological Risk Assessment
Lower Hudson River Sampling Stations

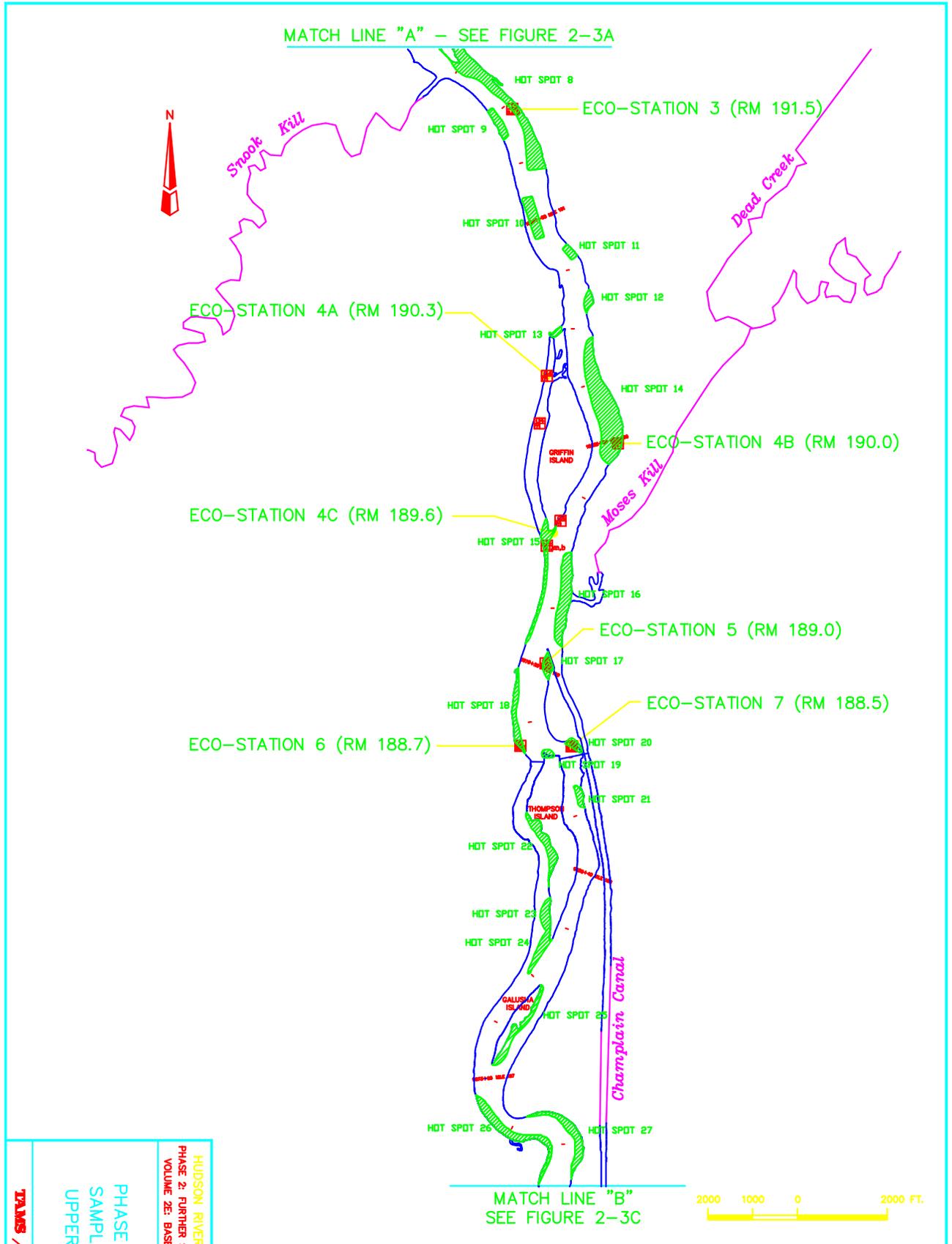




TAMS / MCA
 PHASE 2 ECOLOGICAL SAMPLING LOCATIONS—UPPER HUDSON RIVER
 HUDSON RIVER PCB REASSESSMENT RI/FS
 PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
 VOLUME 2E: Baseline Ecological Risk Assessment
 FIGURE 2-3A

LEGEND:
 ● ECOLOGICAL SAMPLING STATION

SOURCES:
 SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.
 1. REMNANT DEPOSIT 1 AREA OBTAINED FROM NYSDEC (1993) AND AREAS 2, 3, 4, 5 OBTAINED FROM GENERAL ELECTRIC (CANONIE ENVIRONMENTAL, 1991).
 2. HOT SPOTS 5 THROUGH 8 WERE DIGITIZED FROM NYSDEC'S "PCB RECLAMATION PROJECT" DRAWINGS (DECEMBER 1985) AT A SCALE OF 1" = 200'.
 3. HOT SPOTS 1 THROUGH 4 (AROUND ROGERS ISLAND) ARE NOT SHOWN SINCE THEIR CONTINUED EXISTENCE IS UNCERTAIN DUE TO CHANNEL MAINTENANCE DREDGING SUBSEQUENT TO NYSDEC'S 1977/78 SAMPLING.

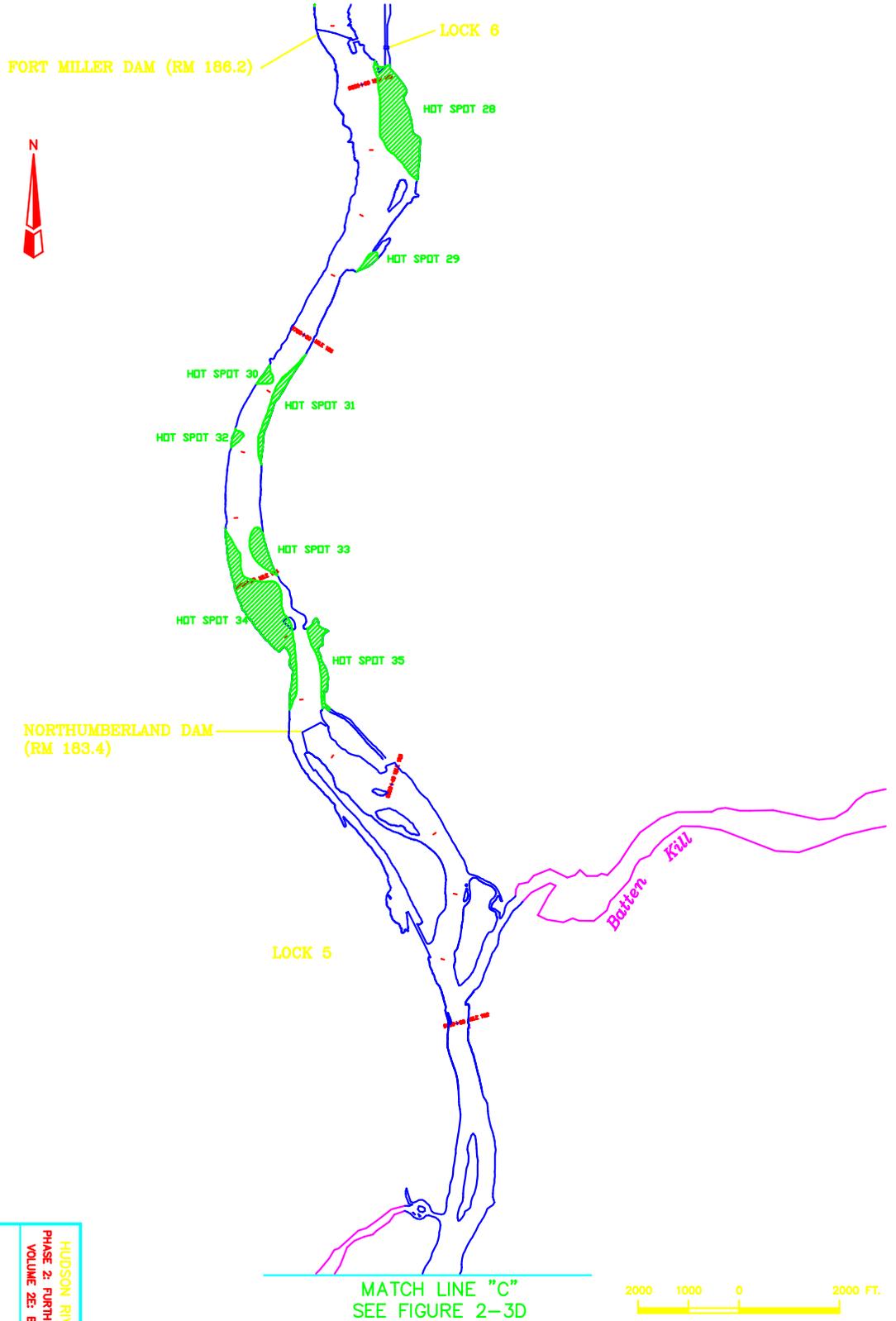


TAMS / MCA
PHASE 2 ECOLOGICAL SAMPLING LOCATIONS—UPPER HUDSON RIVER
HUDSON RIVER PCB REASSESSMENT RI/FS
PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT
FIGURE 2-38

LEGEND:
 ● ECOLOGICAL SAMPLING STATION

SOURCES:
 SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.
 1. HOT SPOTS 8 THROUGH 20 WERE DIGITIZED FROM NYSDEC'S "PCB RECLAMATION PROJECT" DRAWINGS (DECEMBER 1985) AT A SCALE OF 1" = 200'.
 2. HOT SPOTS 21 THROUGH 27 WERE DIGITIZED FROM NUS CORPORATION'S "UPPER HUDSON RIVER AREA" DRAWINGS (APRIL 1984) AT A SCALE OF 1-1/4" = 1 MILE.

MATCH LINE "B" - SEE FIGURE 2-3B



MATCH LINE "C" SEE FIGURE 2-3D



LEGEND:
 ● ECOLOGICAL SAMPLING STATION

SOURCES:
 SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.
 1. HOT SPOTS 28 THROUGH 35 WERE DIGITIZED FROM NUS CORPORATION'S "UPPER HUDSON RIVER AREA" DRAWINGS (APRIL 1984) AT A SCALE OF 1-1/4" = 1 MILE.

HUDSON RIVER PCB REASSESSMENT RI/FS
 PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
 VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

PHASE 2 ECOLOGICAL
 SAMPLING LOCATIONS-
 UPPER HUDSON RIVER

TPAWS / MCA

FIGURE 2-3C

MATCH LINE "C" - SEE FIGURE 2-3C



Fish Creek

180

MATCH LINE "D"
SEE FIGURE 1-3E



TAMS / MCA

PHASE 2 ECOLOGICAL
SAMPLING LOCATIONS--
UPPER HUDSON RIVER

FIGURE 2-3D

HUDSON RIVER POB REASSESSMENT RI/FS
PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

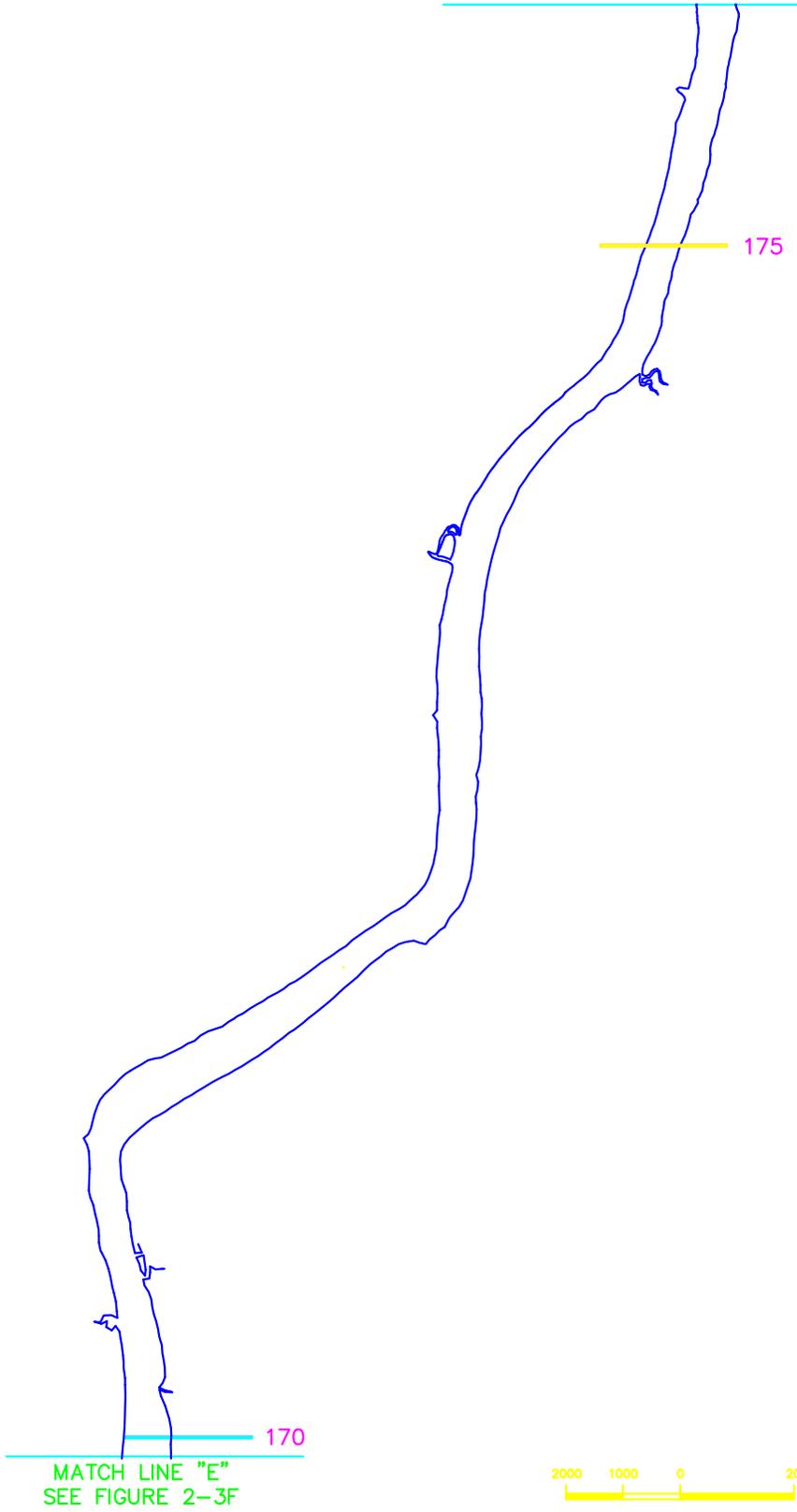
LEGEND:

— 180 RIVER MILE (RM) UPSTREAM OF THE BATTERY

SOURCES:

SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.

MATCH LINE "D" - SEE FIGURE 2-3D



MATCH LINE "E" SEE FIGURE 2-3F

LEGEND:

— 175 RIVER MILE (RM) UPSTREAM OF THE BATTERY

SOURCES:

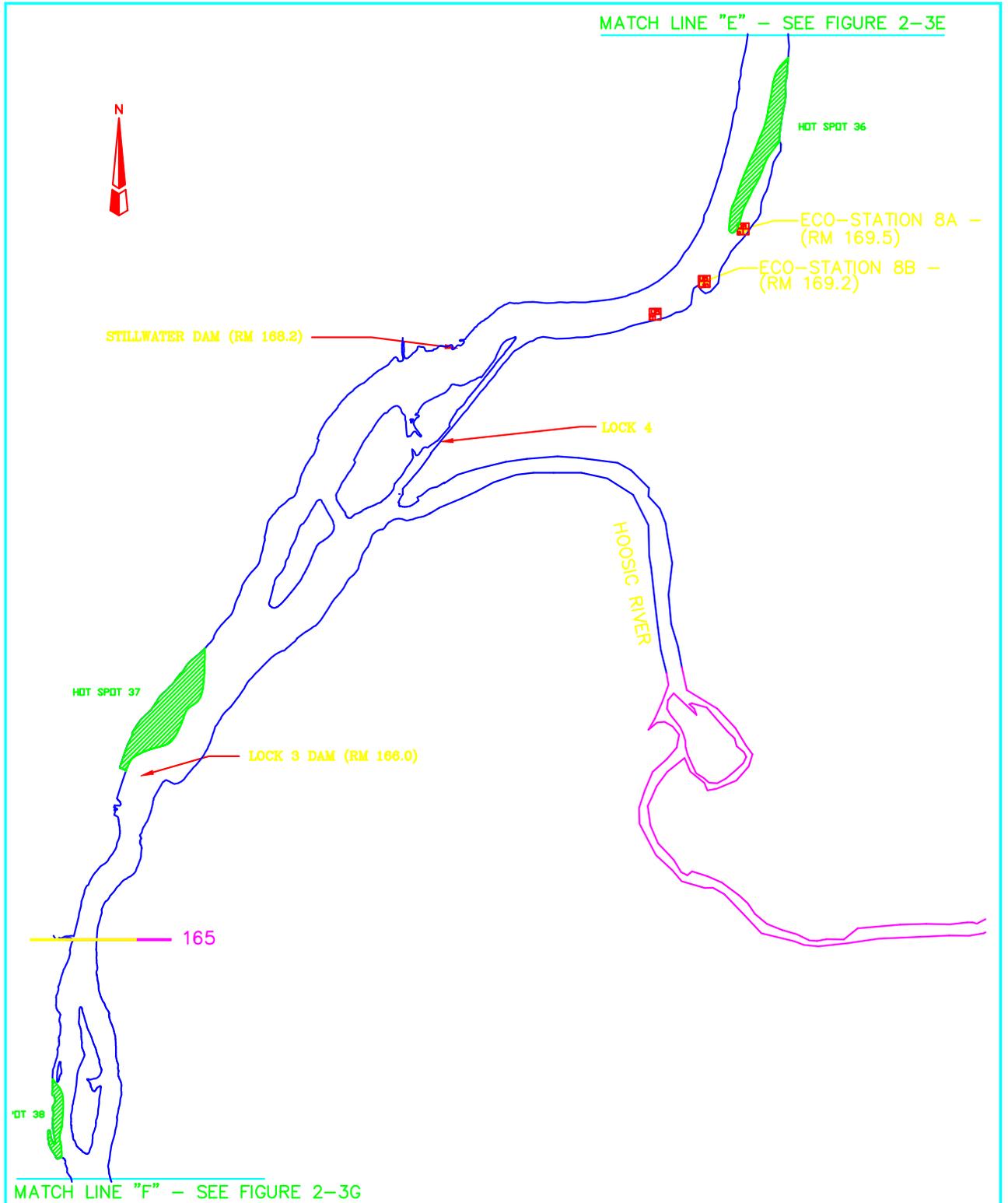
SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.

HUDSON RIVER PCB REASSESSMENT RI/FIS
 PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
 VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

PHASE 2 ECOLOGICAL
 SAMPLING LOCATIONS—
 UPPER HUDSON RIVER

TJAMS / MCA

FIGURE 2-3E



LEGEND:

- 165 RIVER MILE (RM) UPSTREAM OF THE BATTERY
- ECOLOGICAL SAMPLING STATION

SOURCES:

- SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.
1. HOT SPOTS 36 THROUGH 38 WERE DIGITIZED FROM NUS CORPORATION'S "UPPER HUDSON RIVER AREA" DRAWINGS (APRIL 1984) AT A SCALE OF 1-1/4" = 1 MILE.

TRAMS / MCA

FIGURE 2-3F

PHASE 2 ECOLOGICAL SAMPLING LOCATIONS- UPPER HUDSON RIVER

HUDSON RIVER PCB REASSESSMENT RI/FS
 PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
 VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

MATCH LINE "F" - SEE FIGURE 2-3F



HOT SPOT 39

HOT SPOT 40

QUACK ISLAND

LOCK 2 DAM (RM 163.5)

160

LOCK 1 DAM (RM 159.4)

ECO-STATION 9 (RM 159.0)

MATCH LINE "G" SEE FIGURE 2 - 3H



LEGEND:

- 160 RIVER MILE (RM) UPSTREAM OF THE BATTERY
- ECOLOGICAL SAMPLING STATION

SOURCES:

- SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.
- HOT SPOTS 39 AND 40 WERE DIGITIZED FROM NUS CORPORATION'S "UPPER HUDSON RIVER AREA" DRAWINGS (APRIL 1984) AT A SCALE OF 1-1/4" = 1 MILE.

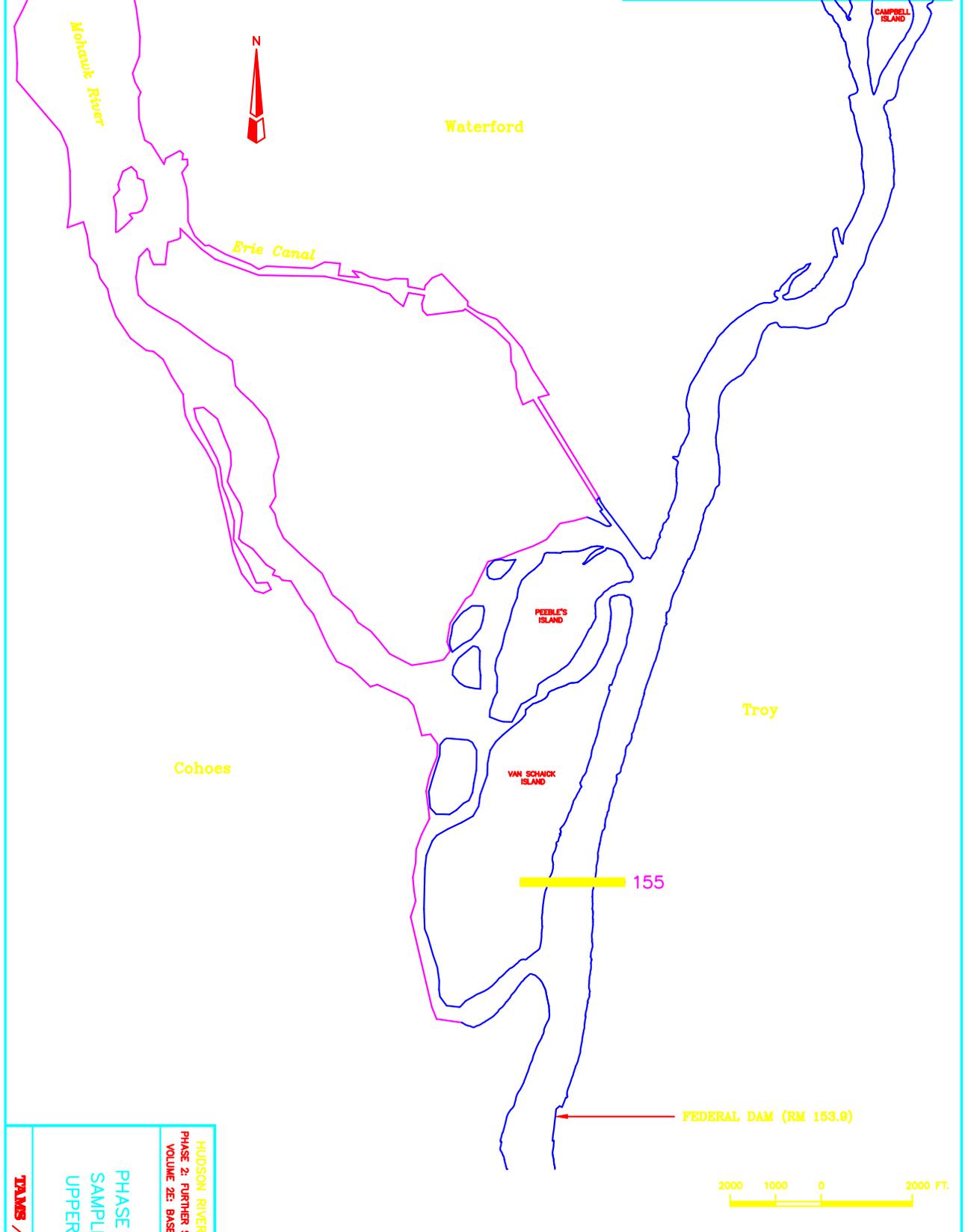
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FIGURE 2-3G

PHASE 2 ECOLOGICAL AND SAMPLING LOCATIONS-UPPER HUDSON RIVER

HUDSON RIVER PCB REASSESSMENT RI/FS
PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

MATCH LINE "G" - SEE FIGURE 2-3G



HUDSON RIVER PCB REASSESSMENT RI/FIS
 PHASE 2: FURTHER SITE CHARACTERIZATION AND ANALYSIS
 VOLUME 2E: BASELINE ECOLOGICAL RISK ASSESSMENT

PHASE 2 ECOLOGICAL
 SAMPLING LOCATIONS-
 UPPER HUDSON RIVER

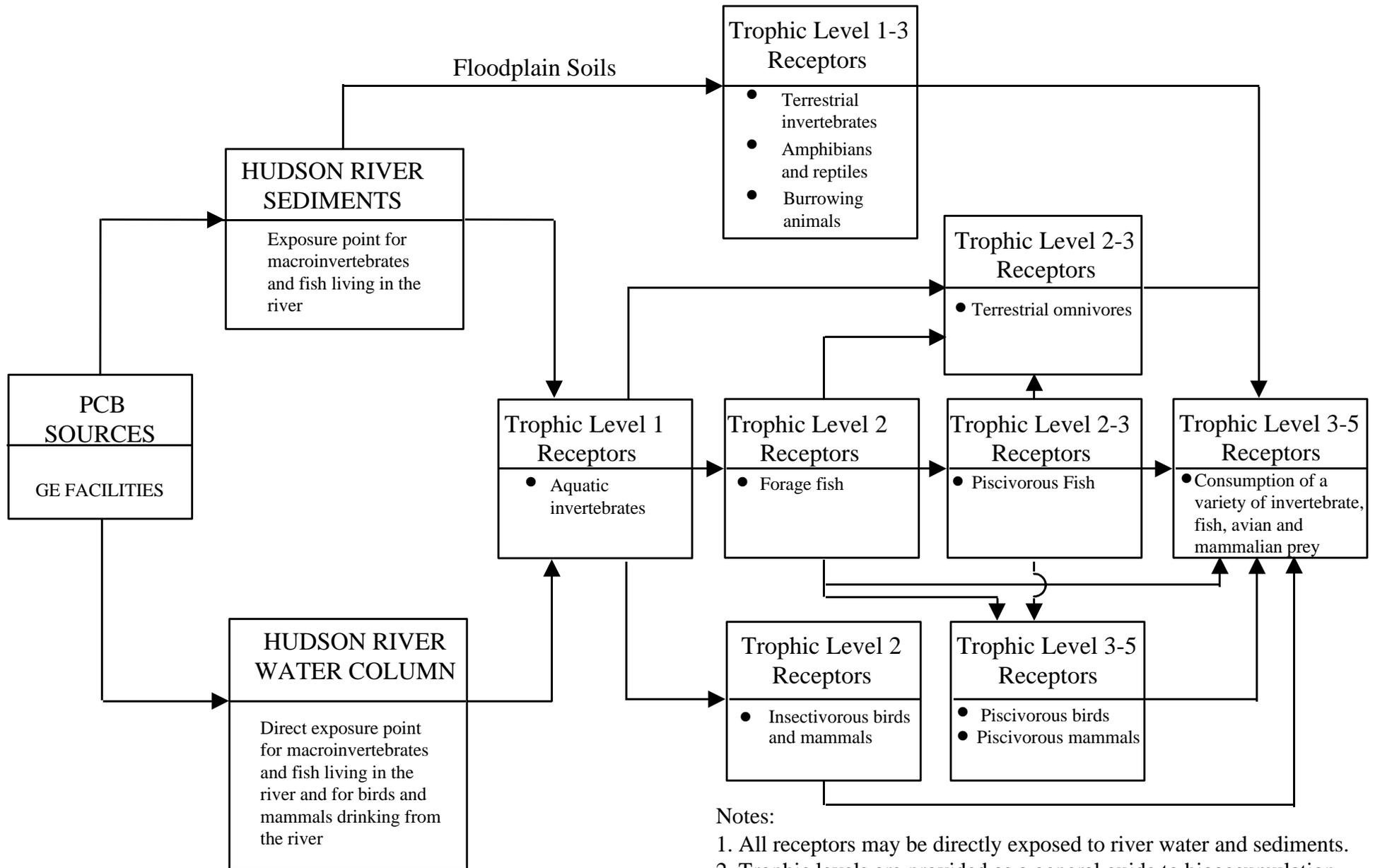
TJAMS / MCA

FIGURE 2 - 3H

LEGEND:
 — 155 RIVER MILE (RM) UPSTREAM OF THE BATTERY

SOURCES:
 SHORELINES AND RM DESIGNATIONS ARE APPROXIMATE.

Figure 2-4
Hudson River PCB Reassessment
Conceptual Model Diagram Including Floodplain Soils



Notes:

1. All receptors may be directly exposed to river water and sediments.
2. Trophic levels are provided as a general guide to bioaccumulation potential but vary according to species and food availability

FIGURE 3-1: AVERAGE WET WEIGHT PCB CONCENTRATIONS IN SELECTED FISH SPECIES BASED ON NYSDEC DATA

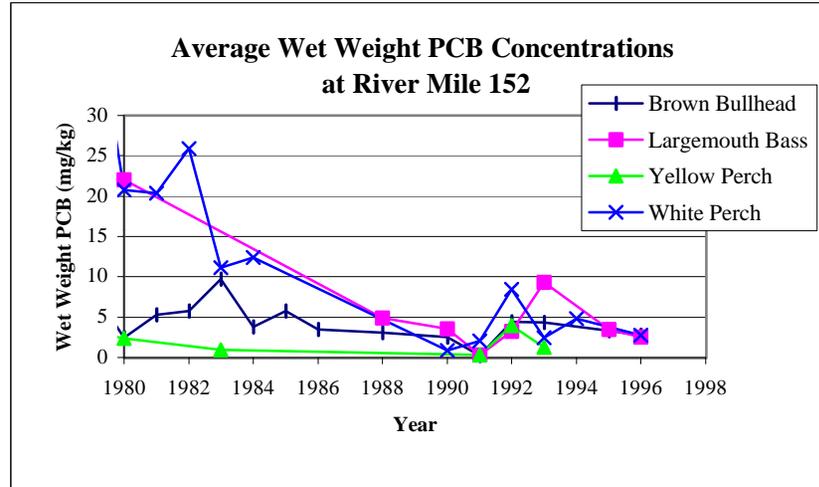
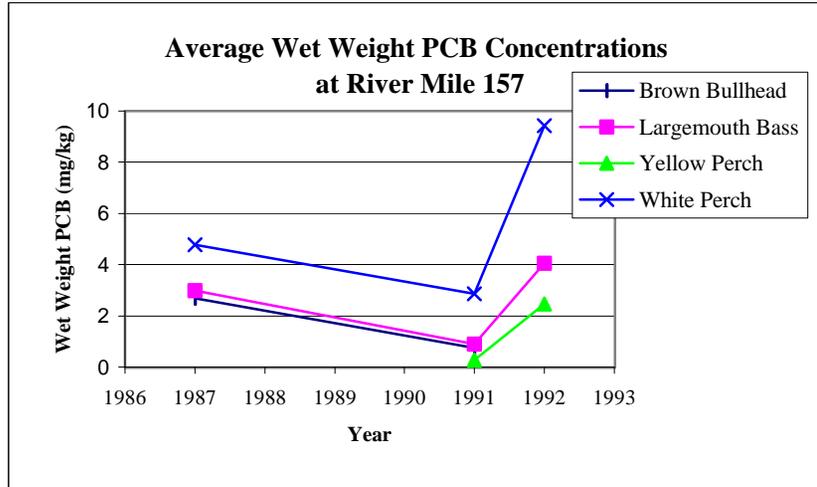
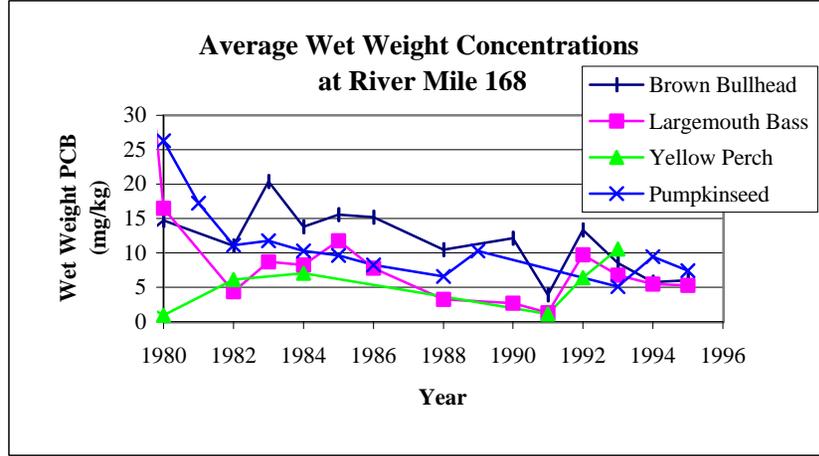
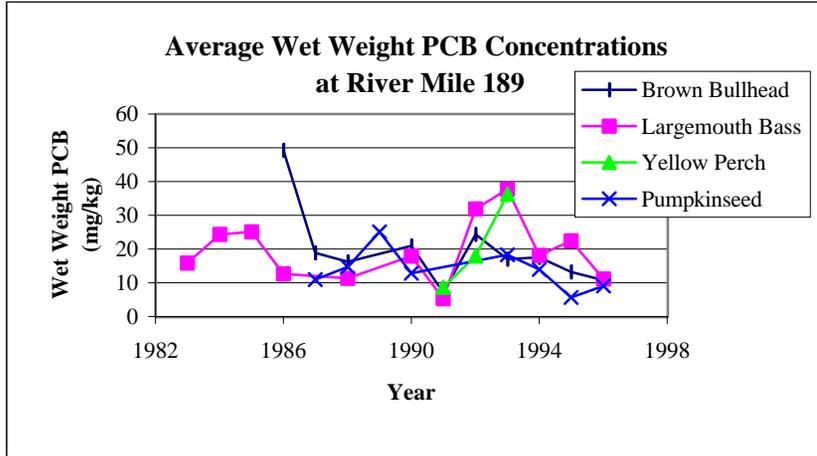
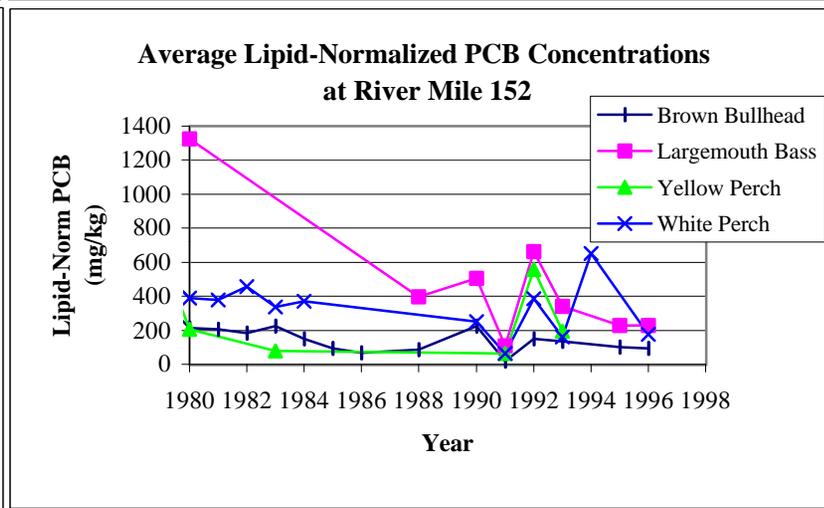
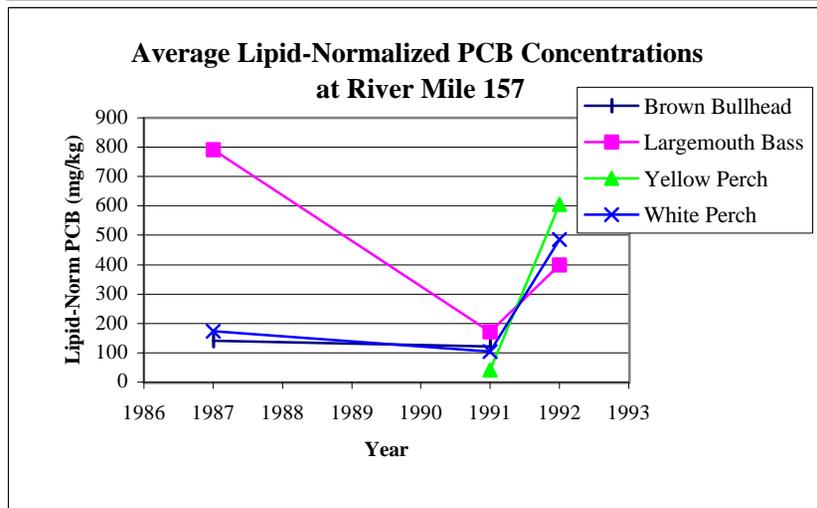
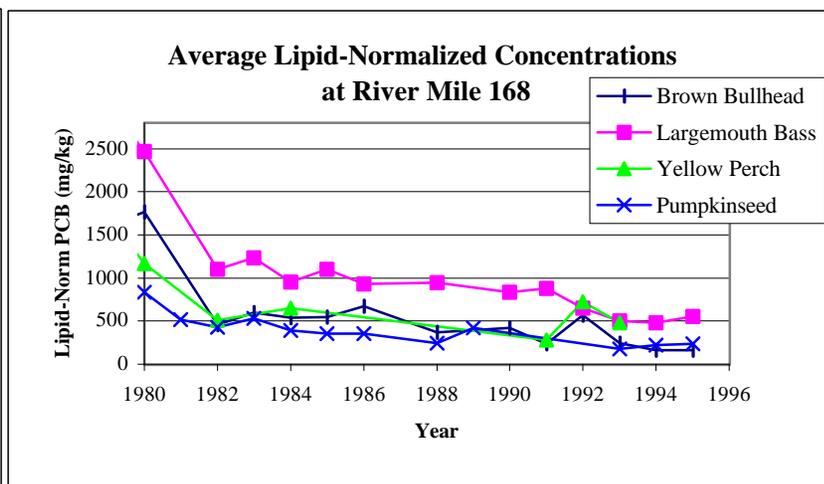
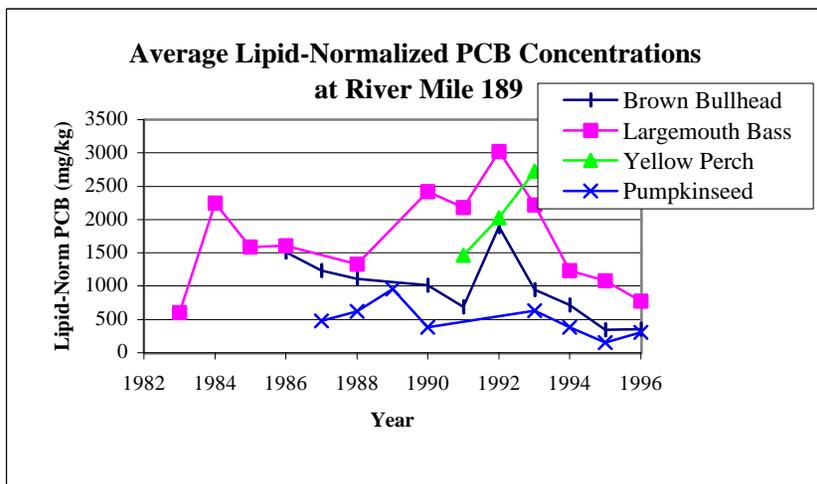


FIGURE 3-2: AVERAGE LIPID-NORMALIZED PCB CONCENTRATIONS IN SELECTED FISH SPECIES BASED ON NYSDEC DATA



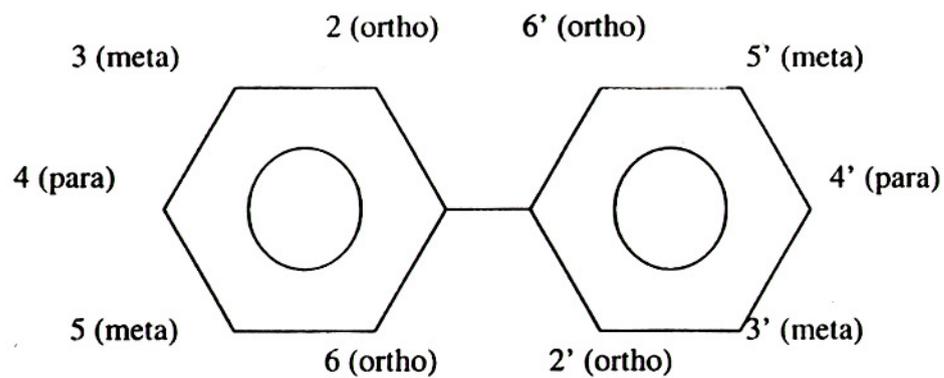


Figure 4-1: Shape of Biphenyl and Substitution Sites

Figure 4-2
Selected Fish Aroclor and Total PCB Toxicity Endpoints

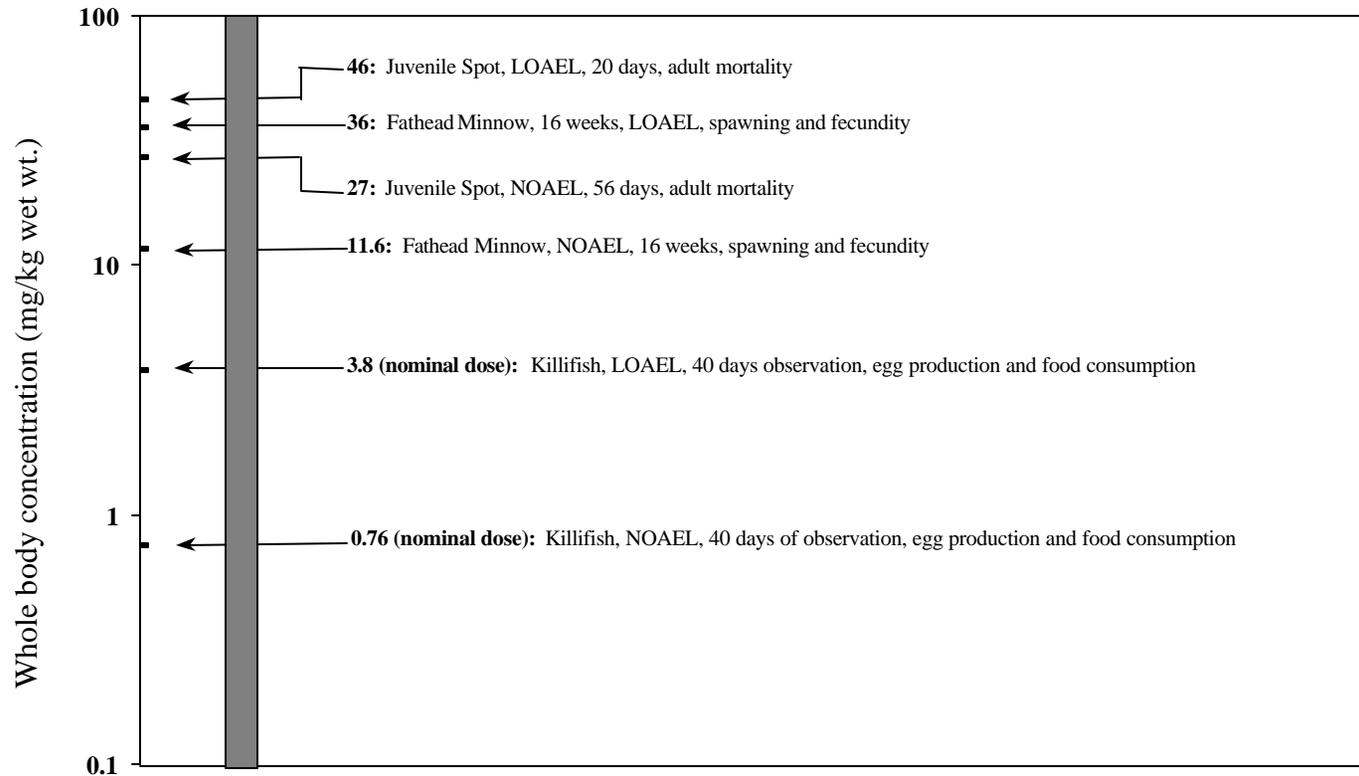


Figure 4-3
Selected Fish Egg Dioxin Equivalent Toxicity Endpoints

Endpoint: Early Life Stage Mortality

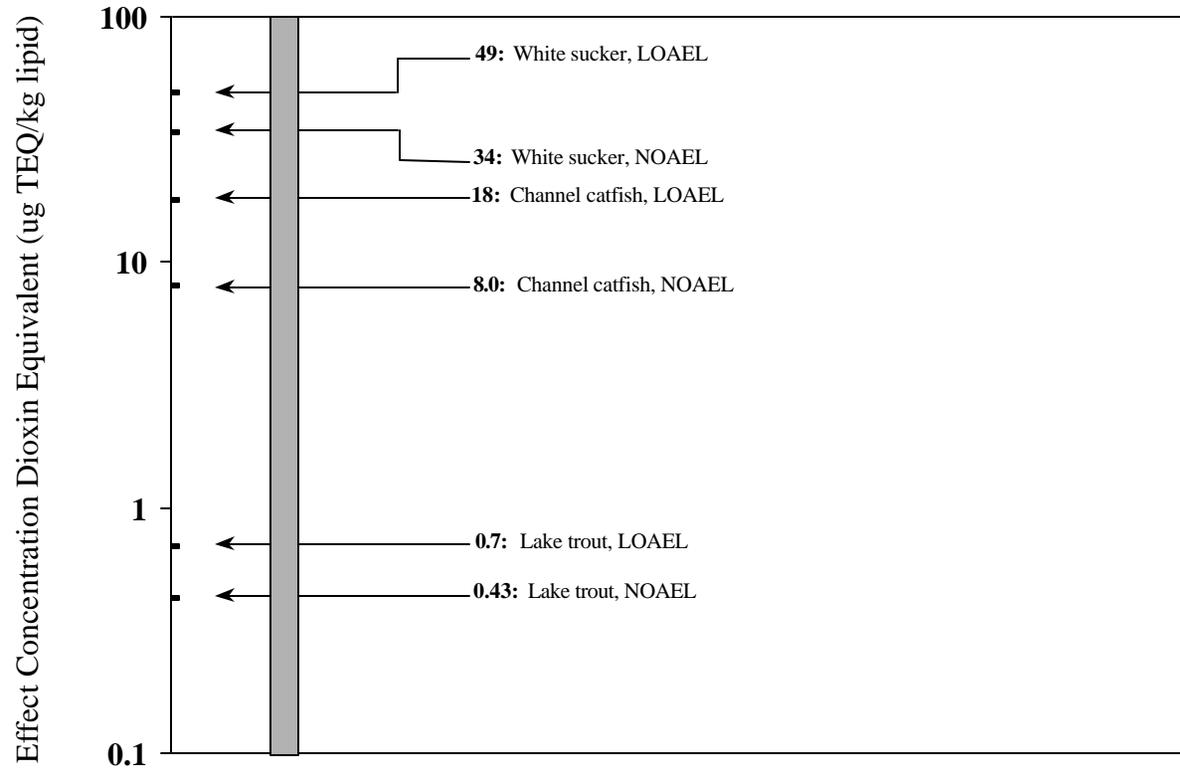


Figure 4-4
 Selected Bird Diet Aroclor and Total PCB Toxicity Endpoints

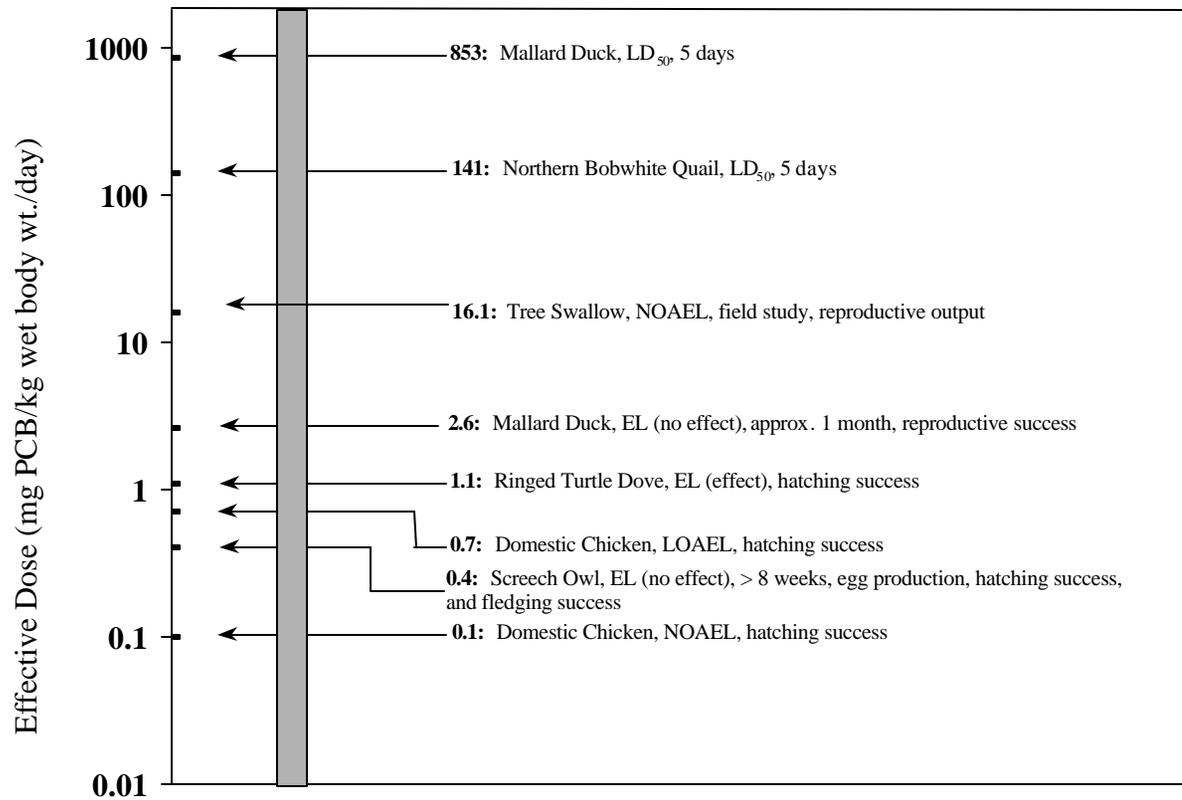


Figure 4-5
Selected Bird Diet Dioxin Equivalent Toxicity Endpoints

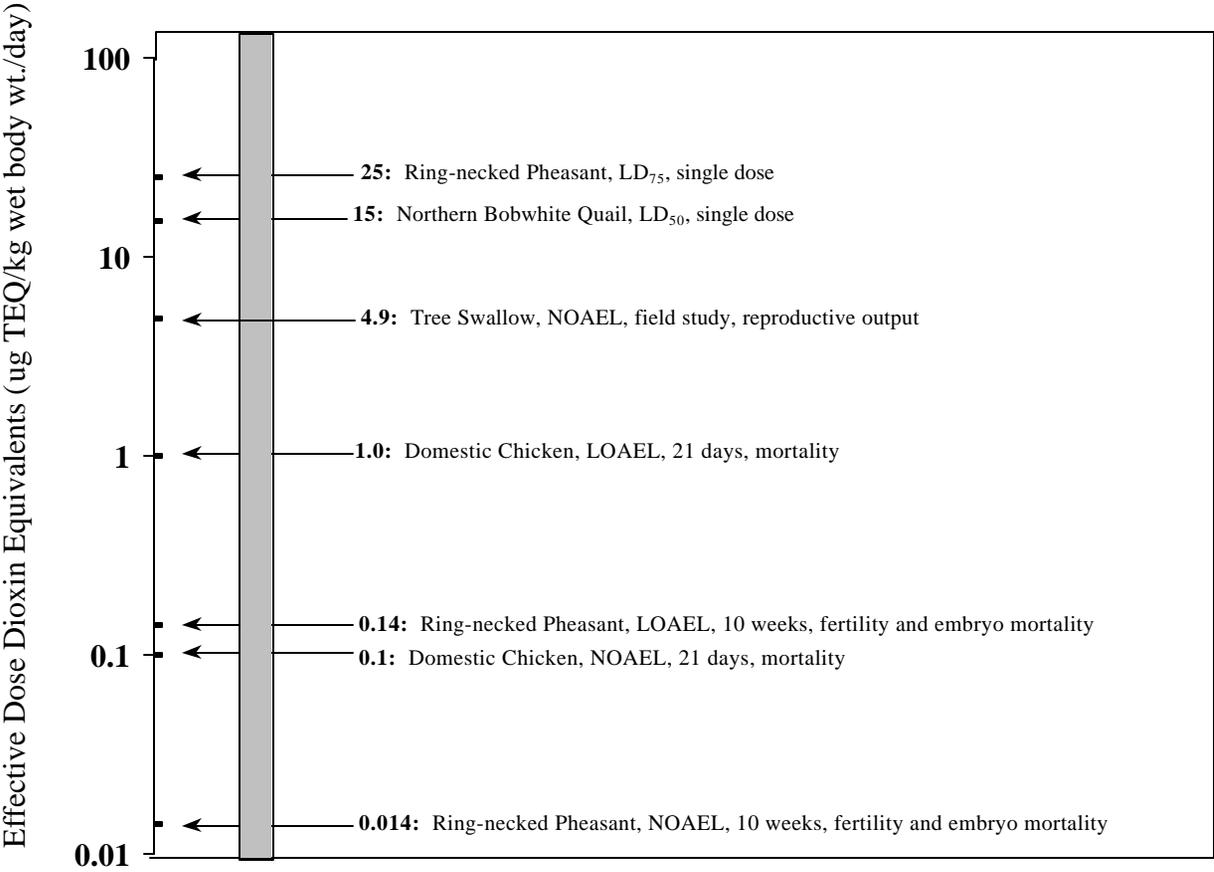


Figure 4-6
Selected Bird Egg Aroclor and Total PCB Toxicity Endpoints

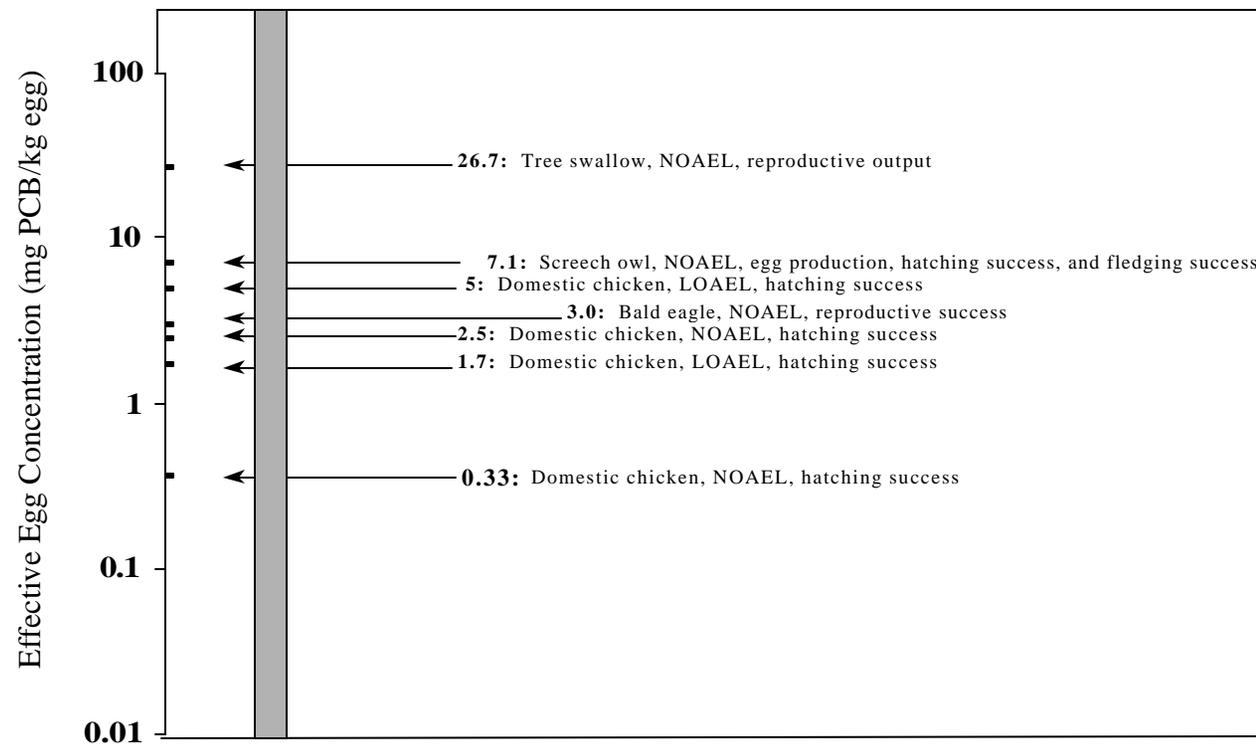


Figure 4-7
 Selected Bird Egg Dioxin Equivalent Toxicity Endpoints

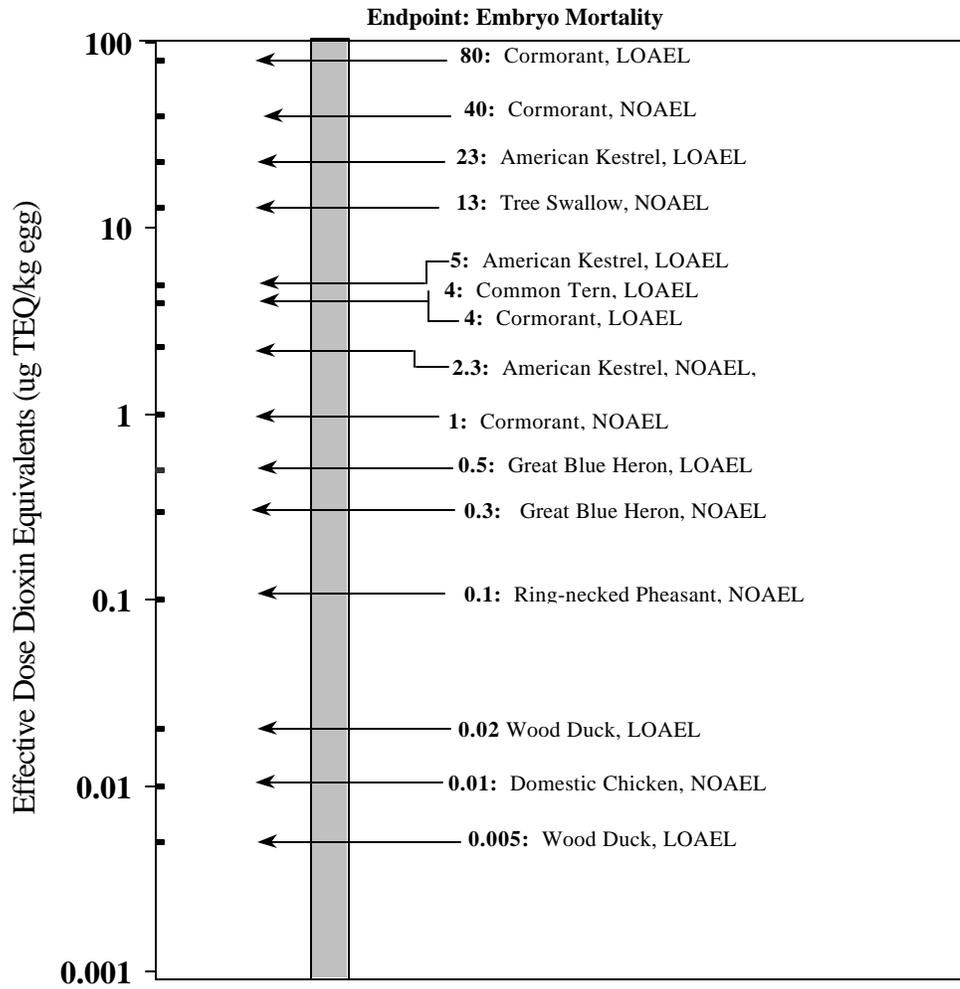


Figure 4-8
Selected Mink Aroclor and Total PCB Toxicity Endpoints

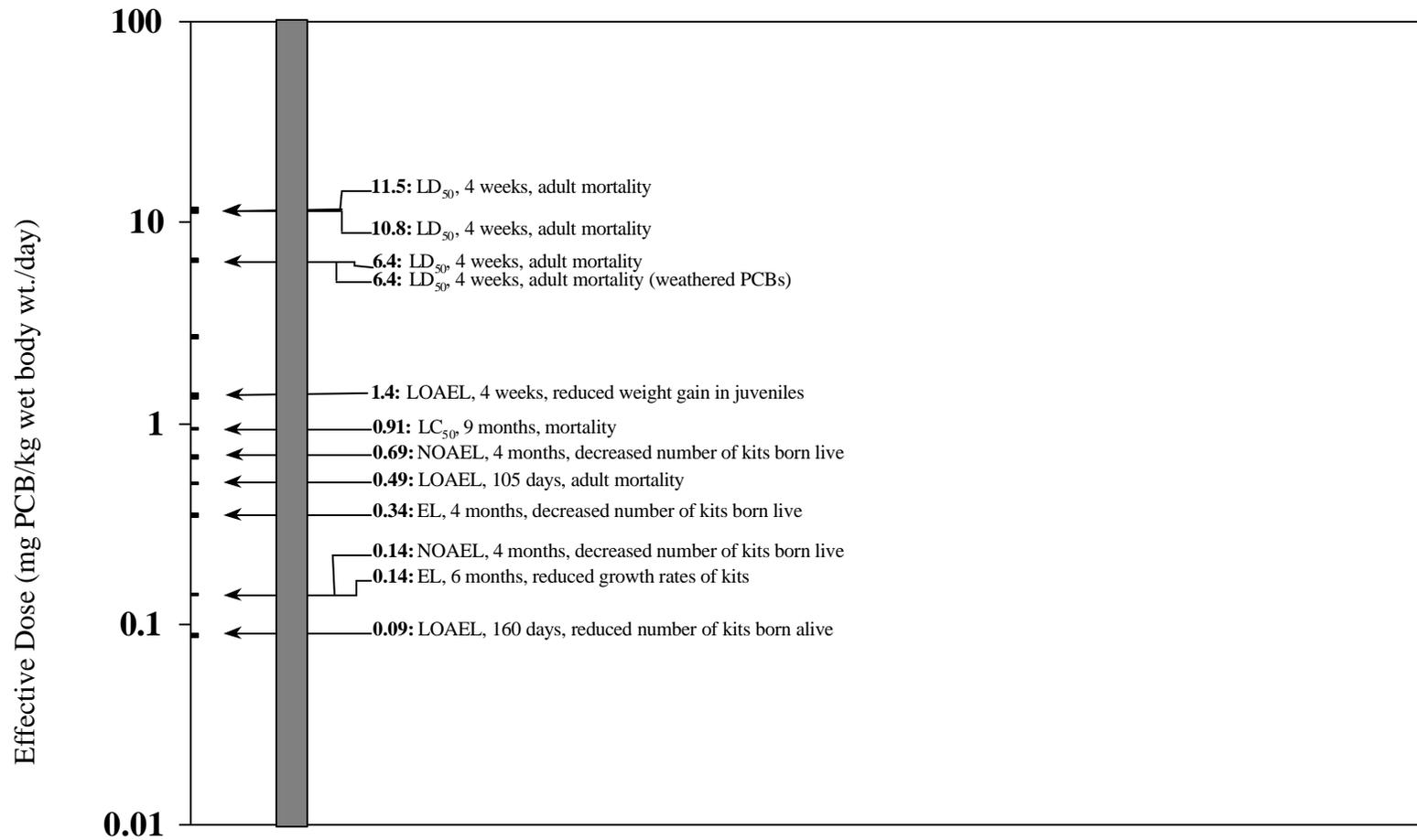


Figure 4-9
Selected Mammal Aroclor and Total PCB Toxicity Endpoints

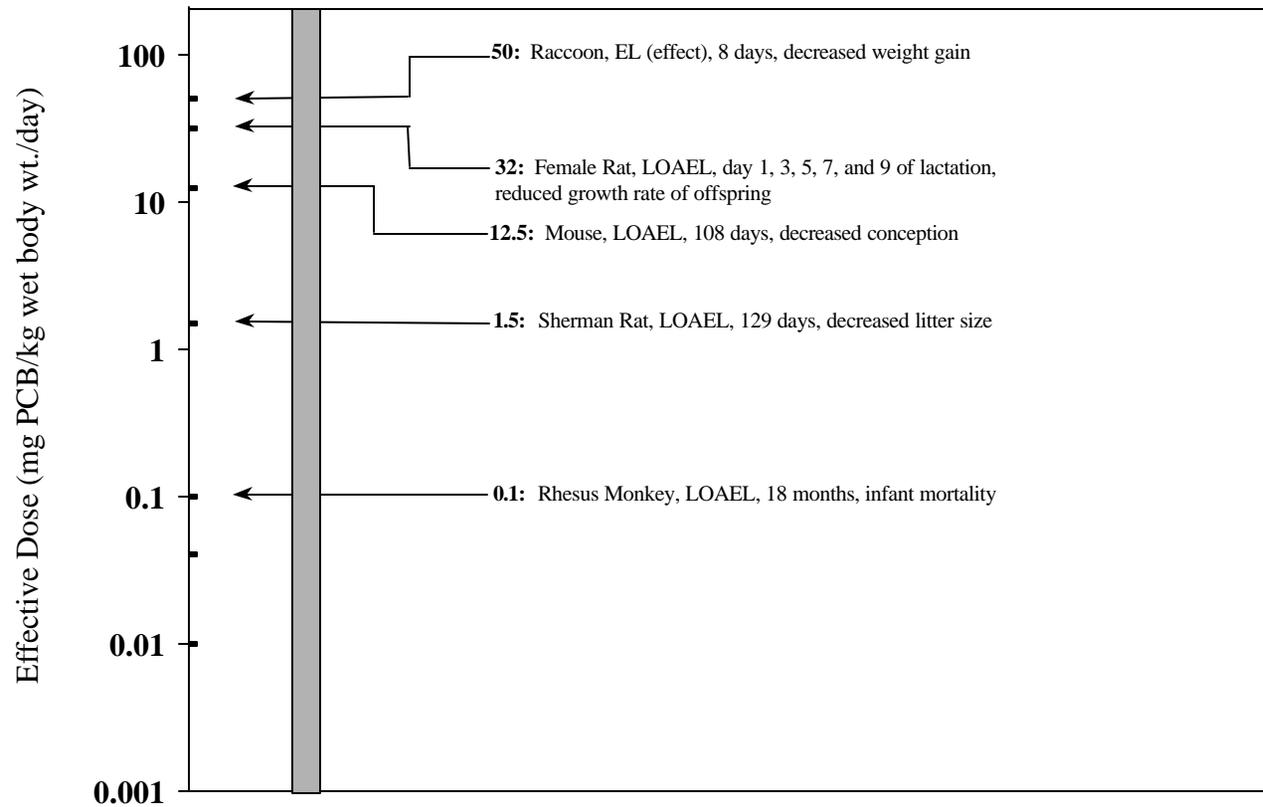
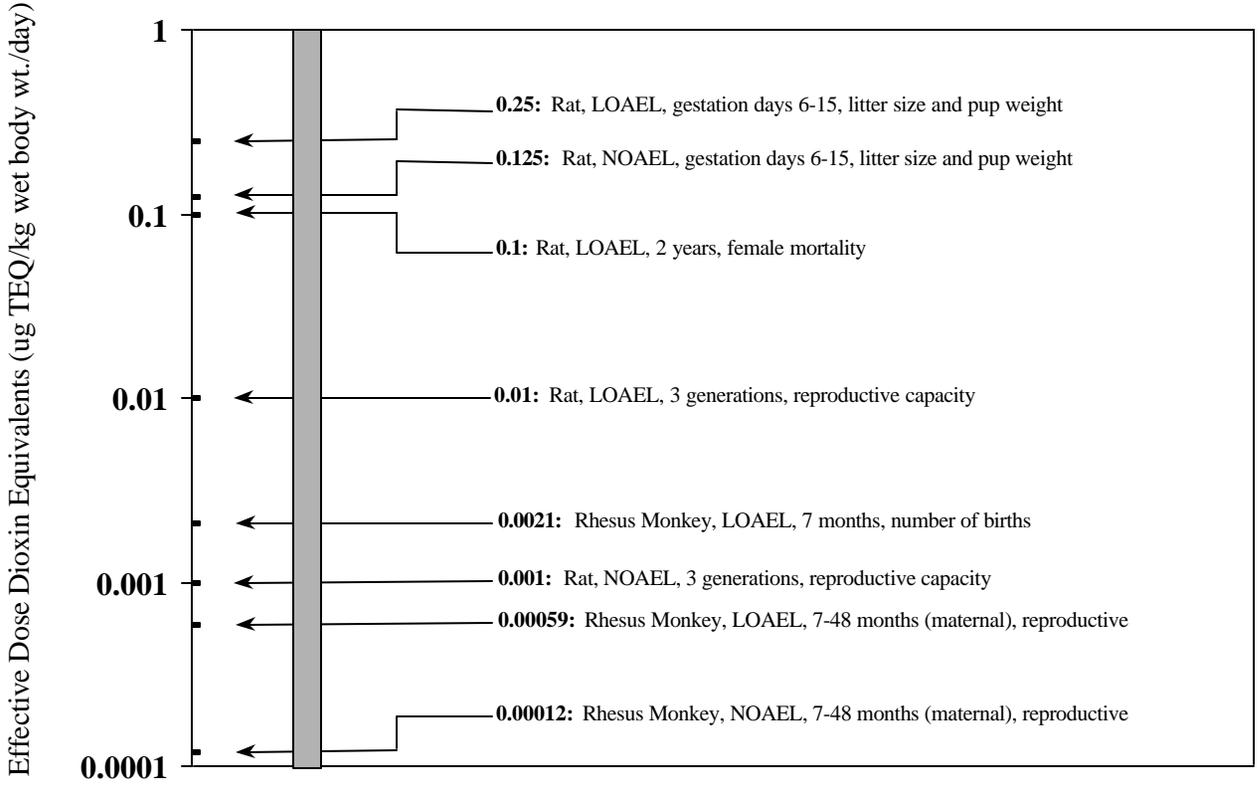
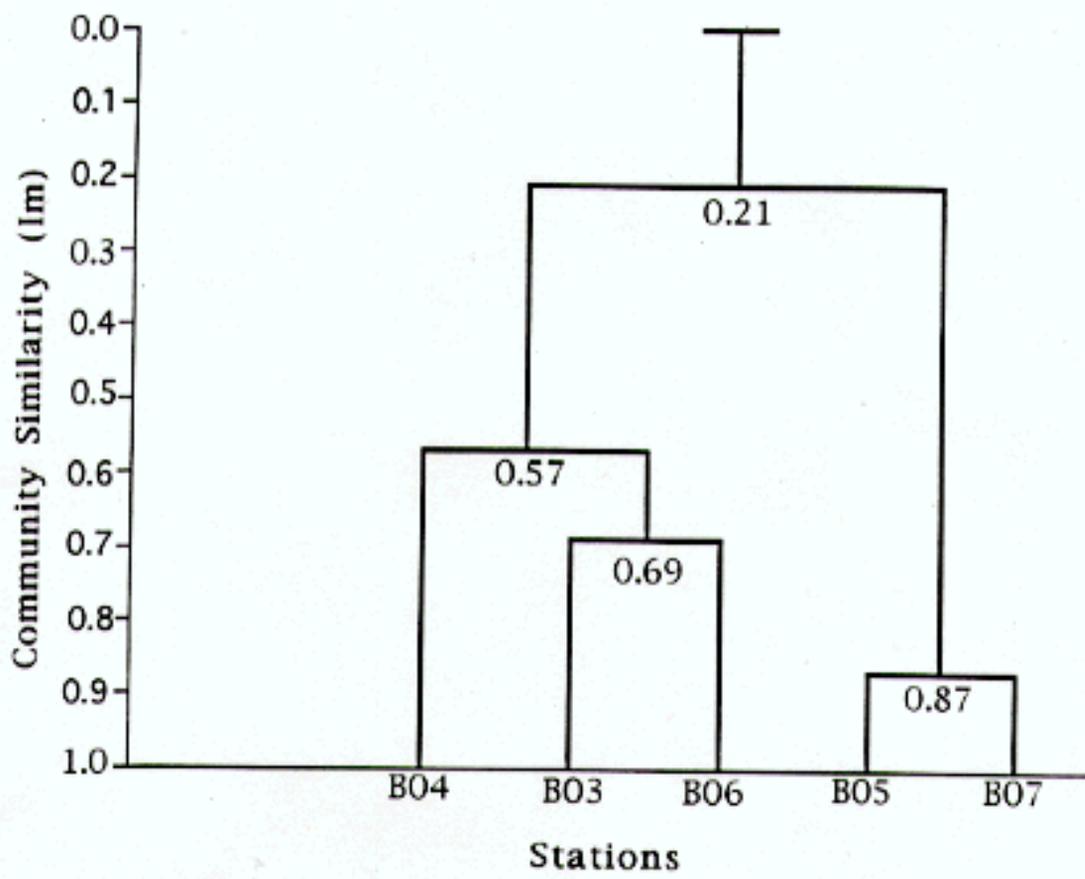


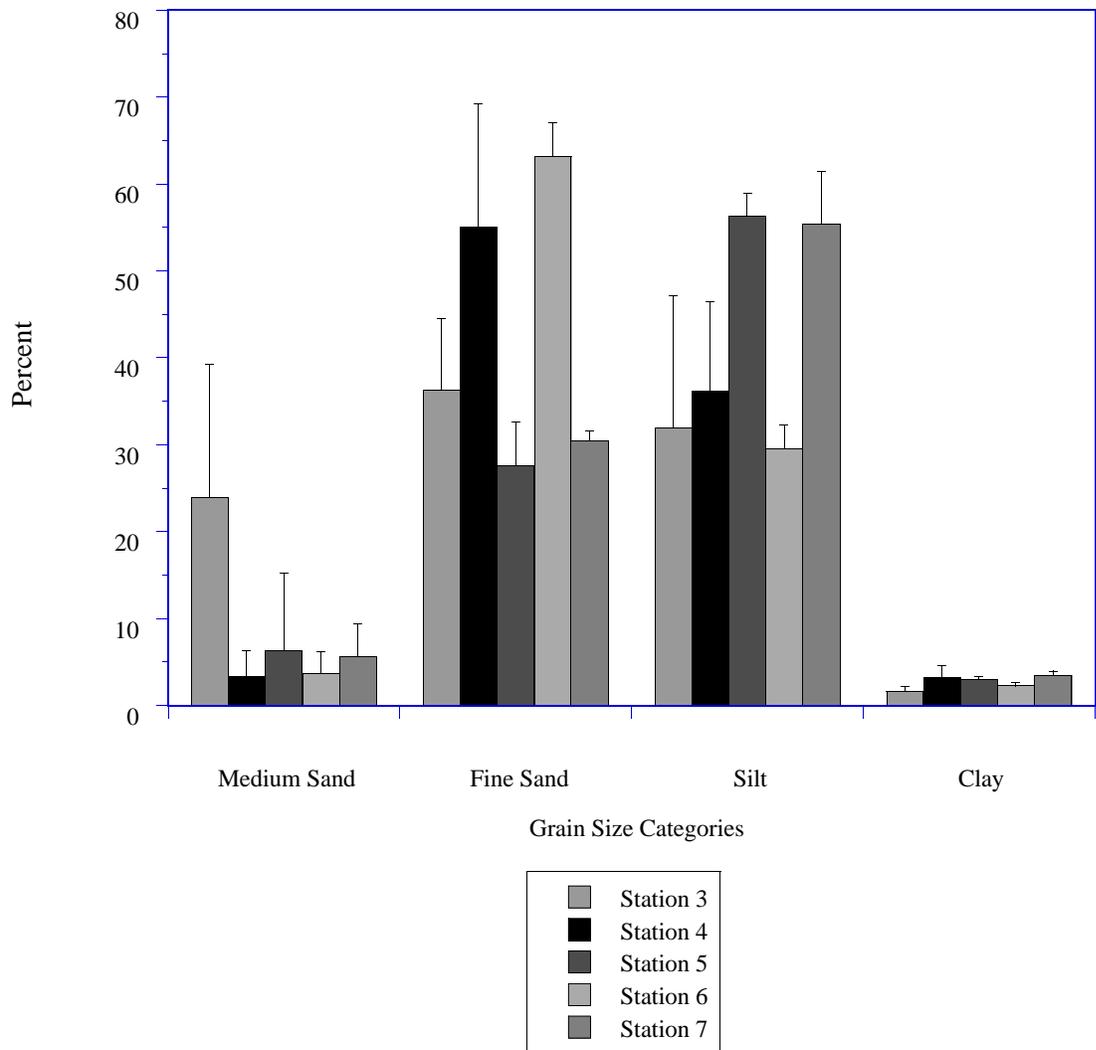
Figure 4-10
 Selected Mammal Dioxin Equivalent Toxicity Endpoints





Note: The dendrogram is based on Morista's Index (I_m) of community similarity and the computed fusion value of each junction is given.

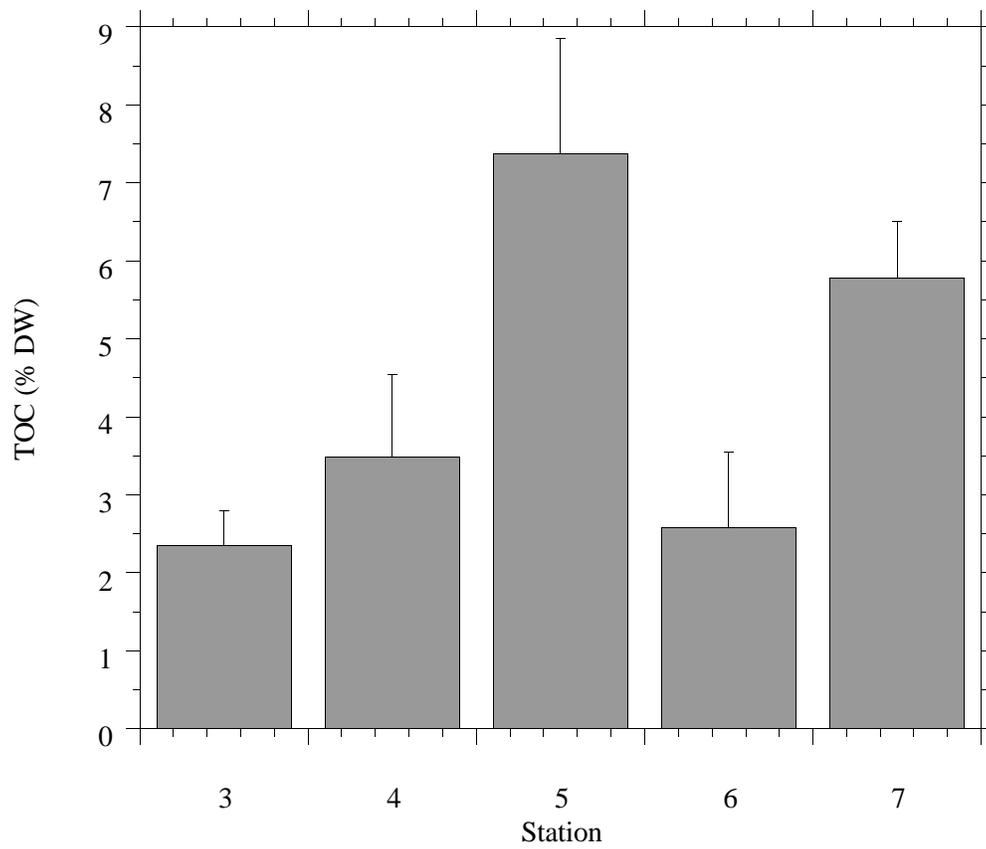
Figure 5-1
Complete Linkage Clustering - TI Pool



Note: Error bars represent one standard deviation.

TAMS/MCA

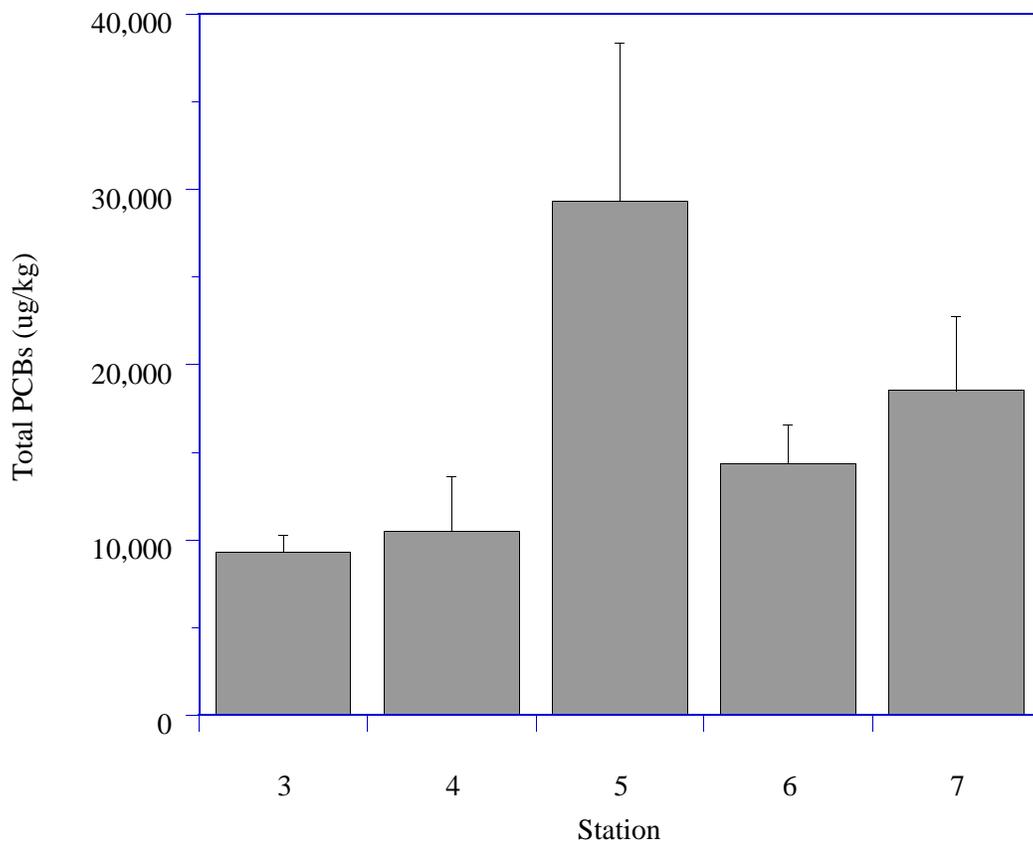
Figure 5-2
Relative Percent Grain Size Classes - TI Pool



Note: Error bars represent one standard deviation.

TAMS/MCA

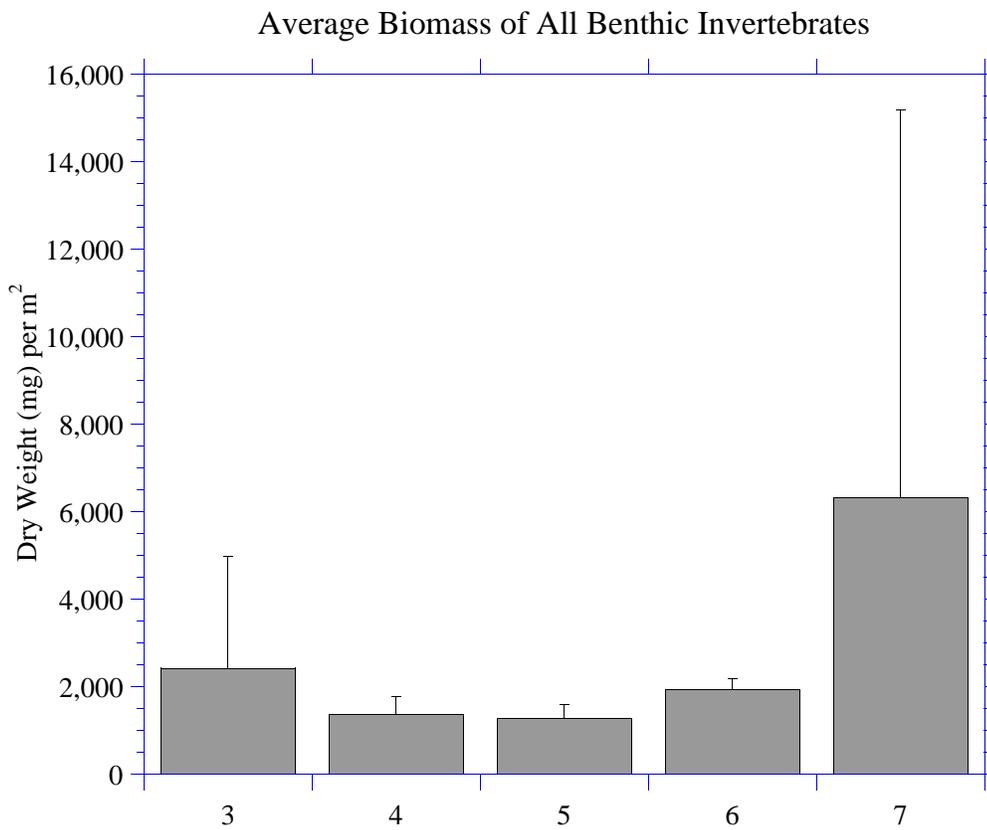
Figure 5-3
Mean Sediment TOC - TI Pool



Note: Error bars represent one standard deviation.

TAMS/MCA

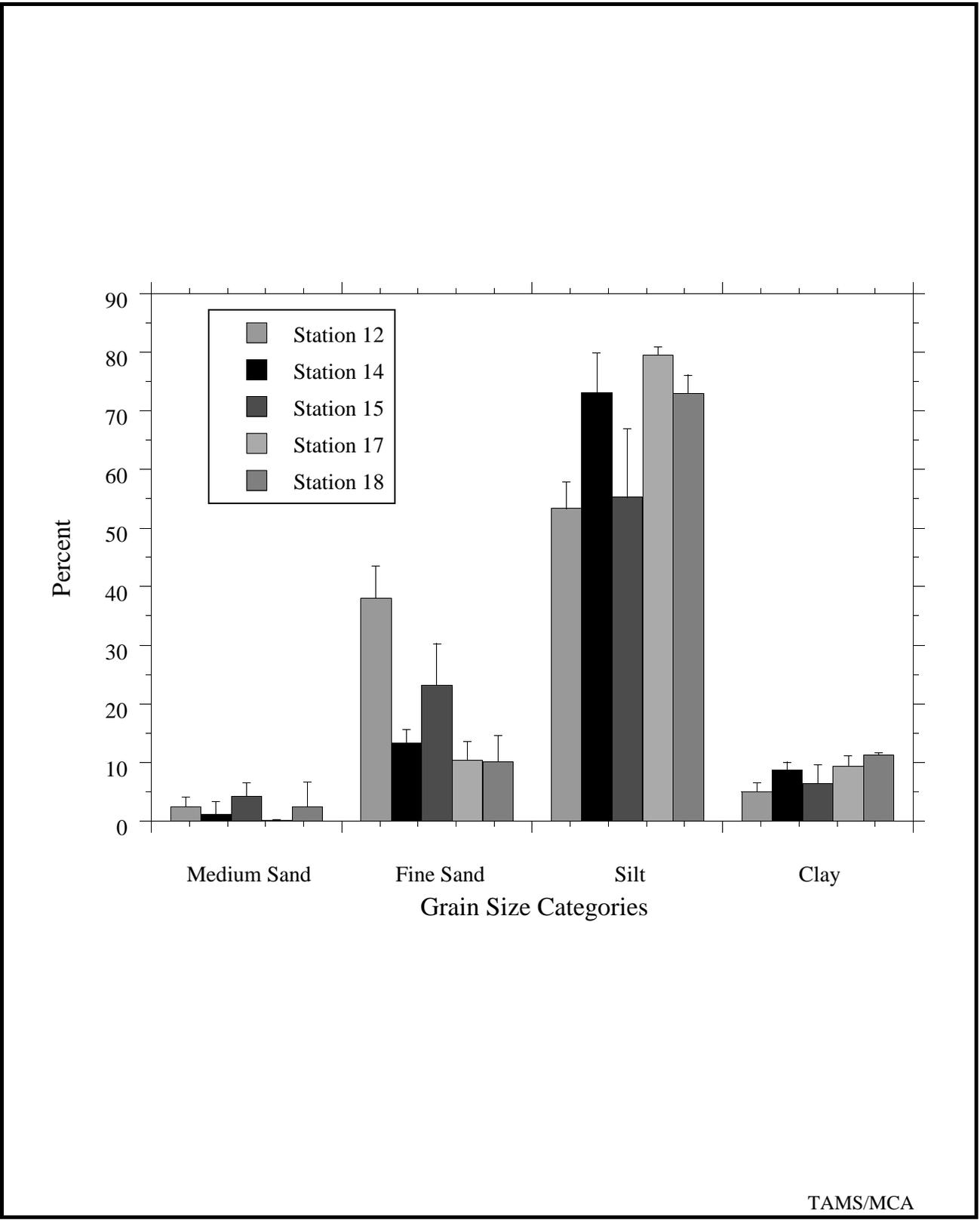
Figure 5-4
Mean Total PCB Concentration in Sediment - TI Pool



Note: Error bars represent one standard deviation.

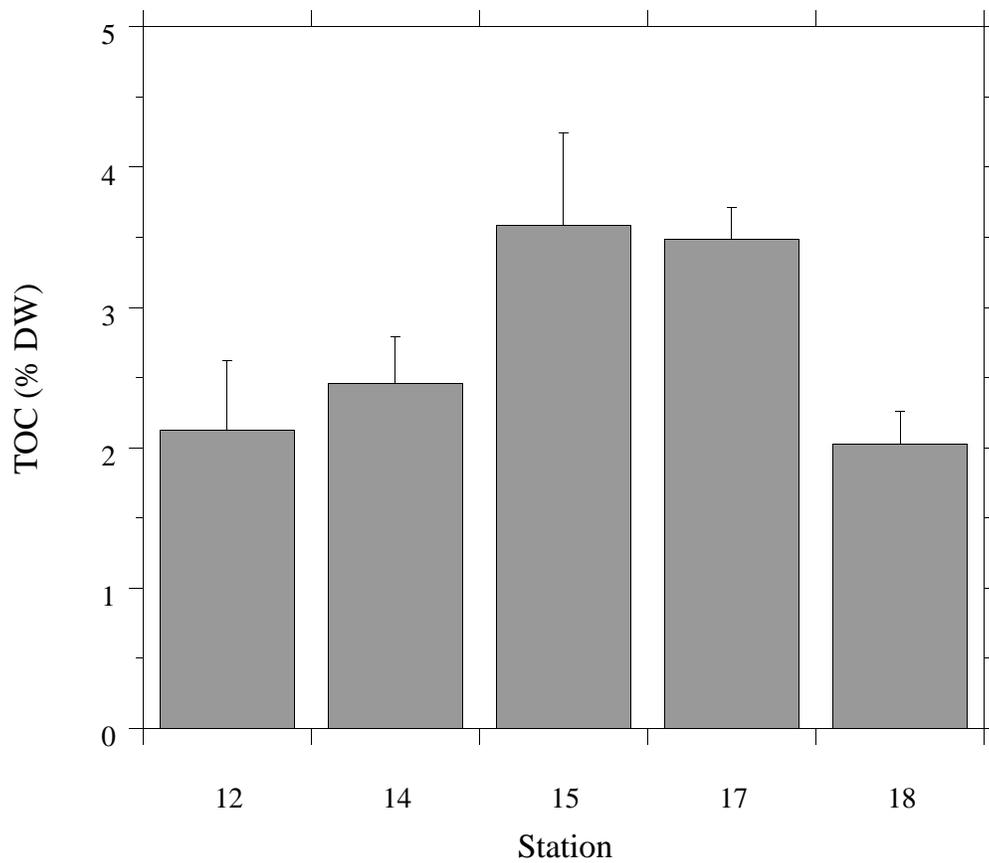
TAMS/MCA

Figure 5-5
Biomass of Benthic Invertebrates - TI Pool



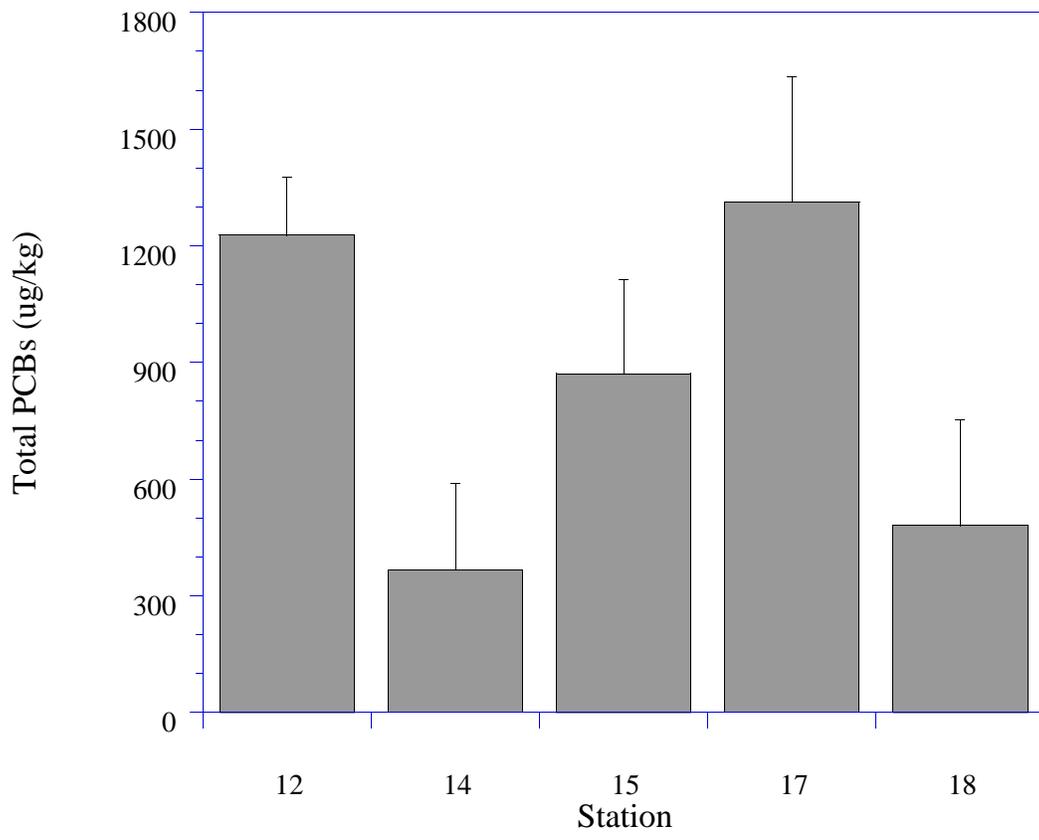
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Figure 5-6
Relative Percent Grain Size Classes - Lower Hudson River



TAMS/MCA

Figure 5-7
Mean Sediment TOC - Lower Hudson River



Note: Error bars represent one standard deviation.

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Figure 5-8
Mean Total PCB Concentration in Sediment - Lower Hudson River